

University of Southern Queensland
Faculty of Engineering and Surveying

Application of an Urban Runoff Model Using Surface Specific Parameters

A dissertation submitted by
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ABSTRACT

Catchment urbanisation has led to a dramatic decline waterway health. The high percentage of impervious surfaces prevalent in urban areas can affect the ecological, economic and social function of our creeks, rivers and bays. Current engineering practice aims to limit this decline using the principles of water sensitive urban design (WSUD). An essential component of this design process is the estimation of stormwater pollutant loads based on catchment characteristics. Current practice involves the estimation of pollutant loads based on land use type. As WSUD seeks to emphasise ‘at source’ controls there is a need to estimate for finer scale catchments. It is anticipated that the use of surface type rather than land use type to determine catchment pollutant loads will prove more accurate at a finer scale. A mass balance model developed by Brodie (2006) utilises surface type to determine pollutant loads.

This dissertation applies Brodie’s mass balance model to a small catchment in Brisbane, QLD to determine the ability of the mass balance model to predict suspended solid loads. This dissertation compares the measured and predicted event mean concentrations of suspended solids over a number of storm events. The result was a poor correlation between predicted and measured data sets due to large inaccuracies in the measured rainfall and use of previously calibrated particle accumulation and washoff parameters not specific to the catchment. A major outcome involves providing comments on the further application of the mass balance model and conducting an error analysis to determine highly sensitive model parameters. The outcome is applicable to stormwater engineers and future users of the mass balance model.

Certification

I certify that the ideas, designs and experimental work, results, analyses and conclusions set out in this dissertation are entirely my own effort, except where otherwise indicated and acknowledged

I further certify that the work is original and has not been previously submitted for assessment in any other course or institution, except where specifically stated

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Signature

Date

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ABBREVIATIONS

ABBREVIATION	MEANING
WSUD	Water Sensitive Urban Design
SQID	Stormwater Quality Improvement Devices
NCP	Non-Coarse Particles
TSS	Total Suspended Solids
BCC	Brisbane City Council
Brisbane CBD	Brisbane Central Business District
BOM	Bureau of Meteorology
ENSO	EL Nino-Southern Oscillation
SOI	Southern Oscillation Index
HSA	Hoyland Street A inlet
HSB	Hoyland Street B inlet
HSC	Hoyland Street C inlet
HSL	Hoyland Street Level Sensor
HSO	Hoyland Street Outlet
HSOa	Hoyland Street Outlet a
HSOb	Hoyland Street Outlet b
VSS	Volatile Suspended Solids
TP	Total Phosphorous
TN	Total Nitrogen
EMC	Event Mean Concentration

DCIA	Directly Connected Impervious Area
ICIA	Indirectly Connected Impervious Area
US	United States
USGS	United States Geological Survey
SSC	Suspended Sediment Concentration
RDI	Rainfall Detachment Index
ADP	Antecedent Dry Period
SRM	Statistical Rational Method
AMC	Antecedent Moisture Condition
USQ	University of Southern Queensland
RCP	Reinforced Concrete Pipe
ALS	Airbourne Laser Survey

CHAPTER 1

INTRODUCTION

1.0 INTRODUCTION

“We’re all downstream”

Margaret and Jim Drescher, Nova Scotia

Extensive catchment urbanisation has led to a dramatic change in the natural hydrology and ecology of many urban catchments. The high percentage of impervious area prevalent in major cities and towns has produced an increase in runoff volumes and a decrease in stormwater quality. The end result is a modification of stream flow processes causing erosion and habitat degradation. It is now widely accepted that a notable relationship exists between stream degradation and impervious area.

The effects of catchment urbanisation on stormwater discharge has gained considerable awareness within the community over the past 20 years, resulting in a shift in traditional engineering approach and application to stormwater. Local governments have been placing increasing emphasis on the treatment of urban stormwater through stormwater quality improvement devices (SQIDs). More recently, the focus has moved from the broad scale application of end-of-pipe solutions to source controls such as stormwater retention and infiltration (O’Loughlin, 2007). This reflects the emergence of water sensitive urban design (WSUD), which seeks to combine all elements of the water cycle.

An essential aspect of stormwater management practice involves estimating stormwater pollutant loads based on catchment characteristics. Current engineering practice is to estimate stormwater pollutant loads from urban catchments based on land use parameters. These land use parameters are often divided based on the zoning of the land use type (eg residential, industrial). This method proves practical and

reasonably accurate for larger catchments, however for smaller catchments it has been considered that stormwater loads determined from a specific land surface (eg. roofs, gassed areas) are more effective. As current practice seeks to emphasise “at source” controls, the need for accurate information on smaller catchments is increasing. This necessitates a need for accurate modelling of these smaller catchments to assist in design and planning.

A recent University of Southern Queensland research project (Brodie, 2006) has developed a simple empirically based model to estimate suspended solid loads in stormwater runoff from various urban surfaces. Non-coarse Particles (NCP) as opposed to Total Suspended Solids (TSS) were used to measure the suspended solids in the analysis. Data for the model was obtained utilising runoff samples from 40 storm events from December 2004 to January 2006 within the inner city of Toowoomba, Australia. Runoff samples were obtained using an innovative flow splitting device for 5 surface types; galvanised roof, grass, bare soil, commercial carpark and road. According to Brodie, the model was able to provide reasonable estimates ($R^2 = 0.74$ to 0.97) when compared to the measured loads. The model also performed well in comparison to three models used in current engineering practice; the Arithmetic Mean, Logarithmic Mean and Stochastic EMC methods.

1.1 Project Outline and Objectives

The purpose of this dissertation is to assess the effectiveness of Brodie’s mass balance model in predicting the stormwater loads of total suspended solids from a small urban catchment. This catchment is located at Hoyland Street in Northern Brisbane, Australia and drains into a bioretention basin before discharging into Bald Hills Creek. The inflows to this bioretention system have been monitored as

part of the WSUD monitoring program instituted by Brisbane City Council (BCC).

The catchment is significantly different in terms of climatic conditions and surface characteristics to the catchments sampled by Brodie. It is expected that significant variation in predicted and measured suspended solids will be evident over a number of storm events. The effectiveness of the mass balance model will be determined by direct comparison with the measured and modelled data, following the application of modelled runoff and measured rainfall data. The comparison will also focus on determining the extent of any relationship between the accuracy of the model and any rainfall event or catchment characteristics. This will aid in determining any major contributing factor to either the accuracy or inaccuracy of the modelled data.

Due to the nature and design of the bioretention system and monitoring arrangement, large inconsistencies in the recorded inflow data are evident. During rainfall events ponding within the system can cause backflow effects. This effectively slows the flow rate and thus causes a delay in the measured runoff hydrograph. In larger events negative inflows have been recorded due to the water level within the system being at a higher elevation than the monitoring equipment. This phenomenon has been investigated by Deletic and Fletcher (2006) and it was confirmed that backwater effects do occur at two of the three monitored inlet pipes. Despite the inconsistencies in discharge data, event specific suspended solid concentrations are available for the bioretention system and are consistent with expected results.

It was initially considered that the results of the mass balance model could be directly compared to measured suspended solids loads in order to provide a direct comparison. Due to the data inconsistencies discussed previously this is not

possible and it has been necessary to model the sub-catchment event discharge in order to apply it to both the measured data and the mass balance model to determine event suspended solids loads. The sub-catchment discharges were modelled using DRAINS hydrologic and hydraulic modelling software. The total event discharge was then compared to the measured discharge for storms where no backwater effects occurred in order to calibrate the model and aid in selecting appropriate input variables.

The outline of this study is to achieve the following broad aims:

1. Conduct a literature review on the prediction of stormwater pollutant loads using urban surface characteristics, restricted to suspended solids.
2. Determine the physical characteristics of the catchment.
3. Gather rainfall, runoff and suspended solids data for the Hoyland Street catchment and select appropriate rainfall events.
4. Develop and calibrate DRAINS model.
5. Test mass balance model.
6. Compare and verify model results with measured suspended solid loads.
7. Write and submit a dissertation on the application and effectiveness of the model in predicting suspended solid loads in urban stormwater runoff.

CHAPTER 2

BACKGROUND

2.0 BACKGROUND

2.1 Climate and Geography of Bald Hills

The city of Brisbane is located in the south-east corner of Queensland, on the Brisbane River adjacent to a large shallow bay known as Moreton Bay. Bald Hills is the northernmost suburb in Brisbane (Figure 2.1) and shares its northern and western borders with the Moreton Bay Regional Council. Largely a residential suburb, Bald Hills has undergone significant development within the last 30 years with many rural allotments being subdivided and gradually taken over as a low-density residential area.

Bald Hills is situated at the confluence of the South and North Pine Rivers which subsequently flow into Moreton Bay. As a result the suburb drains into the South Pine River to its west and Bald Hills Creek to its east. Bald Hills Creek discharges into the environmentally sensitive Tinchi Tamba Wetlands Reserve before entering the Pine River.

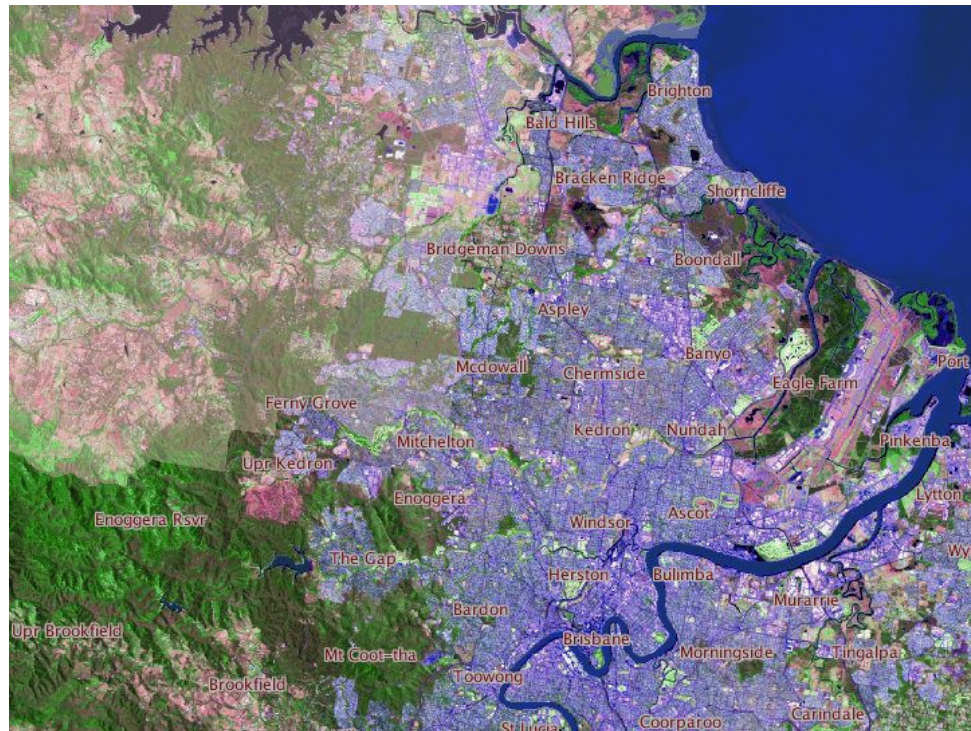


Figure 2.1 Location of Bald Hills Relative to the Brisbane CBD (Source: Brisbane City Council)

Brisbane experiences a subtropical climate with four distinct yet temperate seasons. Average temperatures range from 9.5°C to 20.9°C in the winter, and 20.6°C to 29.1°C in the summer (BOM, 2000), with maximum and minimum recorded temperatures of 39.6°C and 0.6°C respectively. Cool, clear nights are prevalent in the winter months while cloud cover is nearly twice as common in summer. Rainfall is also much greater in the summer months as is the prevalence of high intensity storms with a short duration.

Bald Hills has a typical Brisbane Climate. Figure 2.2 below displays the rainfall and temperature patterns at Sandgate Post Office, approximately 5.5 km east of Bald Hills. It can be seen that rainfall within the catchment is subject to a high degree of variation within the year, with average rainfall in the summer months over twice that of winter.

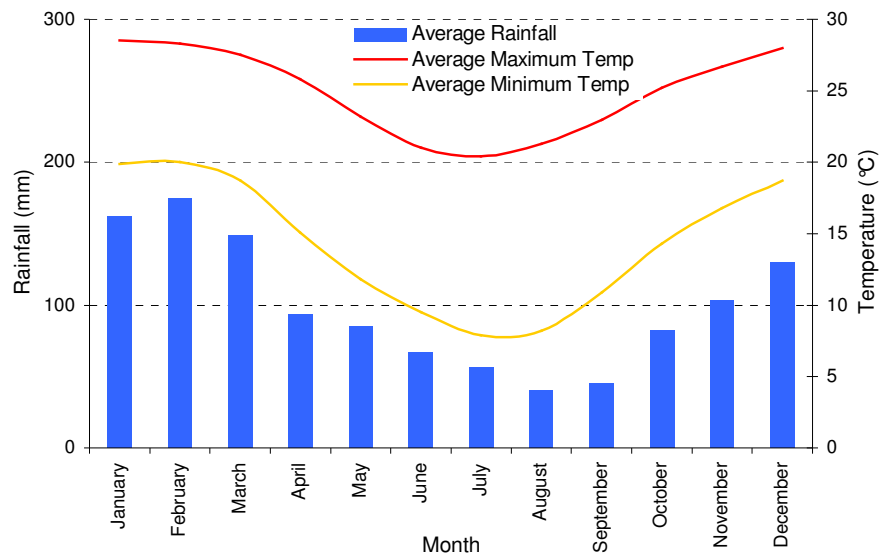


Figure 2.2 Monthly Rainfall and Temperature Averages at Sandgate Post Office (Bureau of Meteorology)

As well as the increased summer rainfall, evaporation is dramatically increased in the summer months, particularly from October to March. As seen in Figure 2.3, this is largely dependant on an associated increase in solar energy associated with seasonal variation leading to increased summer temperatures. Increased wind speeds in the warmer months create an additional transport mechanism which would also result in an increase for evaporation.

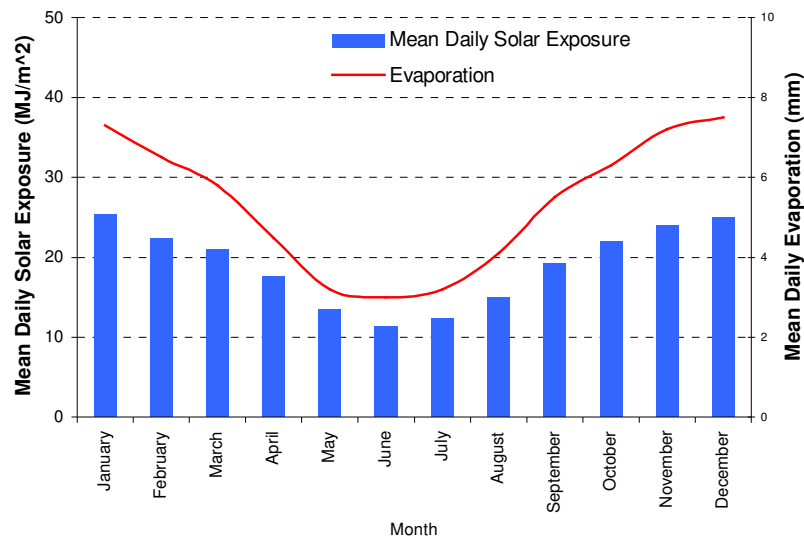


Figure 2.3 Monthly Evaporation and Solar Exposure Averages at Brisbane Airport (Bureau of Meteorology)

Distinct inter-annual rainfall variation occurs along the East Coast of Australia as a result of a phenomenon known as El Nino-Southern Oscillation (ENSO). This phenomenon is influenced by ocean temperature, ocean currents as well as the atmospheric pressure over Darwin and Tahiti. The result of an ENSO event is a large decrease in annual rainfall, which can follow a two to seven year cycle, however is extremely variable.

For Brisbane, including Bald Hills, this event results in dry winters, occasionally with no rain at all for the entire winter months. A prolonged El Nino event has recently occurred for the past five to six years resulting in a prolonged drought. Trends in the Southern Oscillation Index (SOI) indicate that this event ended in late 2007, however accurate predictions of this poorly understood phenomenon are not possible.

2.2 Hoyland Street Catchment

As previously explained, a bioretention system is located at Hoyland Street, Bald Hills and has been monitored as part of a WSUD monitoring program by the Brisbane City Council. The catchment area draining into this system is approximately 1.8ha and is located 20km North of the Brisbane CBD within the wider Bald Hills Creek catchment (BCC, 2006). Three sub-catchments drain into the system through separate inlets at the upstream end of the bioretention system.

2.2.1 Bioretention System

The Hoyland Street bioretention system is located at the intersection of Hoyland and Parer Streets, an Arterial Route connecting the Gympie Arterial Road with the Gateway Motorway (Figure 2.4). The system was constructed in late 2001 as a demonstration project for Water Sensitive Urban Design (WSUD) during a major road upgrade with a view to monitor the site for pollutant removal efficiency.



Figure 2.4 Location of Hoyland Street Bioretention Basin

A bioretention system is a vegetated area where runoff is filtered through a filter media layer as it percolates through the ground (2006, Healthy Waterways). Perforated under-drains collect the filtered stormwater before discharging into the receiving environment. Essentially runoff is treated by mechanical filtration as it passes through the filter media and dense vegetation, and also by adsorption onto the filter media as well as biological uptake by the vegetation. Bioretention basins also limit peak discharge downstream by providing significant flow attenuation.

The three sub-catchments previously described, drain into separate pipes, which in turn flow into the bioretention system. Figure 2.5 illustrates this, and clearly shows the three inlet pipes, known as HSA, HSB and HSC respectively. A row of six 150 mm slotted PVC agricultural pipes at three metre centres convey the filtered runoff into an outlet chamber at the downstream end of the system. A

maintenance pipe is also located within the basin to enable cleaning of the slotted PVC pipes.

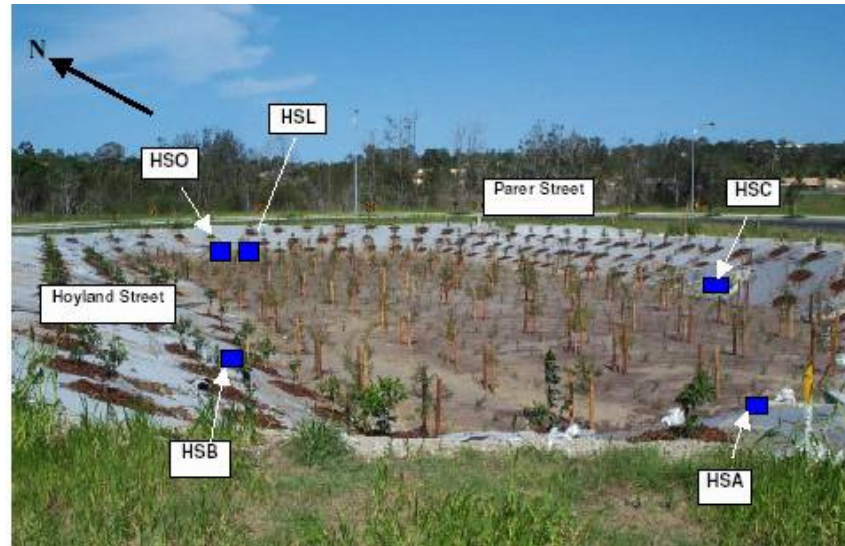


Figure 2.5 Hoyland Street following construction in January 2002 (BCC, 2006)

In addition to the perforated under-drains, a high flow bypass pit with two 675 mm outlet pipes known as HSO1 and HSO2 drain any surface water that has not infiltrated through the soil into nearby Bald Hills Creek. The pit is set 0.5 m higher than the surrounding basin allowing considerable extended detention. This configuration allows the basin to act as a fully on-line system, conveying runoff from major rainfall events directly through the basin before exiting the system through the high-flow bypass.

Planted vegetation within the system was confined to a single tree species and two ground cover species. *Melaleuca quinquenervia* (Broad-leaved paperbark) and *Lomandra longifolia* (Spiny-headed mat-rush) were planted within the basin itself, while *Pittosporum revolutum* (Sweet Pittosporum) was planted in higher

densities on the batters. Figure 2.6 shows the extent of vegetative cover just two years after installation.



Figure 2.6 Hoyland Street in June 2004 (BCC, 2006)

2.2.2 Catchment Characteristics

The total catchment boundary of the bioretention system is detailed in Figure 2.7. It can be seen that the vast majority of the catchment is formed by Hoyland Street itself. Fractions of eight allotments between 500 and 800 m² also drain into the bioretention system as does a portion of steep grassed embankment in the upper part of the catchment and a significant portion of grassed area. Some of this grassed area drains directly into the bioretention system by overland sheet flow; however the rest of the catchment is directly connected to the bioretention by stormwater infrastructure. The bioretention basin itself comprises around 4% of the total catchment area.

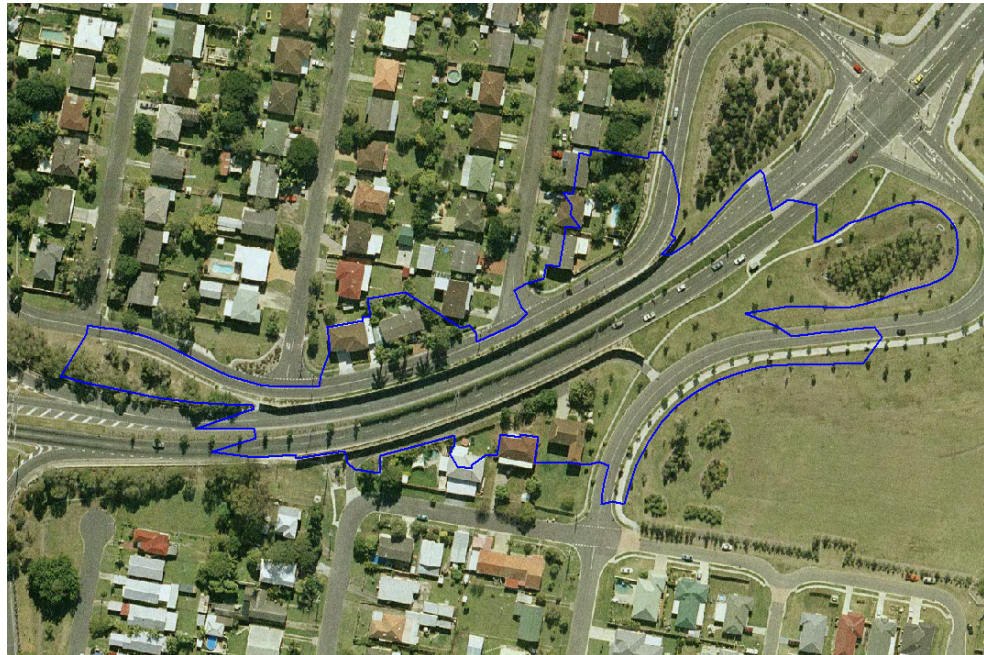


Figure 2.7 Aerial photograph depicting the Hoyland Street catchment

The upper road catchment connecting to the Gympie Road overpass is a moderately steep ramp of around 6%. This grade gradually lessens approaching the bioretention system to just over 1%. The short distance to gully pits, moderate ramp grade and distinct lack of any overland flow path indicate a relatively short time of concentration for the catchment. This coupled with the high percentage of directly connected impervious area result in inflows to the bioretention system during even small rainfall events.

2.2.3 Monitoring Data

Water quality data has been monitored at the Hoyland Street bioretention system since 2002 as part of Brisbane City Council's WSUD monitoring program in order to test the pollutant removal efficiency of the system. Recent analysis has shown that the system has a mean infiltration rate of 101 mm/h

and reduces the mean TSS concentrations by 71% ranging from 32% to 88% (BCC, 2007). According to City Design (2006) samples were tested for a range of pollutants including:

- Total suspended solids (TSS),
- Volatile suspended solids (VSS),
- Total phosphorous (TP),
- Total nitrogen (TN),
- Ammonia,
- NO_x,
- Ortho-P, and
- Heavy metals (zinc, lead, nickel, cadmium, chromium, copper).

Analysis of nutrient species (Ortho-P, NO_x and Ammonia) was only conducted for five events between 2003 and 2006, however all other parameters were consistently tested. Testing of nutrient species required immediate chilling prior to sample collection, which could not be established in all cases.

Flow volumes were also measured at each of the three inlet pipes and outlet pipe using flow meters. In addition, a water level sensor (HSL) was located within the system close to the outlet to measure the ponded level during events. This provided an indication as to when the infiltration capacity of the soil was exceeded and subsequently when the maximum ponded level was exceeded and runoff was exiting the system via the high flow bypass pit.

As previously discussed, several deficiencies exist with the monitoring data at Hoyland Street. More specifically, errors associated with the measurement of inflow and outflow discharge as well as level and velocity. This phenomenon was described by Deletic and Fletcher (2006) and it results in an inability to

establish pollutant loads for the majority of runoff events. These problems include:

- Outlet flow volume occasionally greater than inlet flow volume,
- The backing up of water in the inlet pipes causing inaccurate and sometimes negative flows to be recorded,
- The assumption that equal flow occurs in both outlet pipes was found to be false,
- High tailwater levels from Bald Hills Creek causing inaccuracies in the outflow measurements.

Deletic and Fletcher (2006) assessed the existing monitoring configuration as well as available data and recommended the calculation of pollutant loads using a catchment weighting. A major conclusion of this assessment was that the inflow and outflow data could not be reliably used to estimate pollutant loads entering or being discharged from the system. As a result, the calculation of loads using the catchment area as a weighted average was recommended.

2.2.4 Rainfall Data

Information on rainfall events within the BCC local government area is available for a large number of locations due to a widespread network of pluviometers and stream height gauges. With respect to the bioretention system, rainfall data is available from 6 pluviometers, all within 7km of the system. These pluviometers are tipping buckets with a 1mm resolution. This means that once 1mm of rain has been recorded the bucket tips and the interval between tips is recorded. Unfortunately, the closest rainfall gauge (BDR712) has not collected data for the entire monitoring period.

In addition to direct rainfall data, Rainfall contour maps are available for individual events of a specified duration. These contour maps provide an indication as to the spatial variation in rainfall during events by interpolating the rainfall at points between the rainfall gauges. Figure 2.8 below displays an example of this for a rainfall event on the 03/01/2008. These maps can prove an indispensable tool in selecting rainfall events that do not exhibit a large spatial variation between the rainfall gauge and the monitoring site.

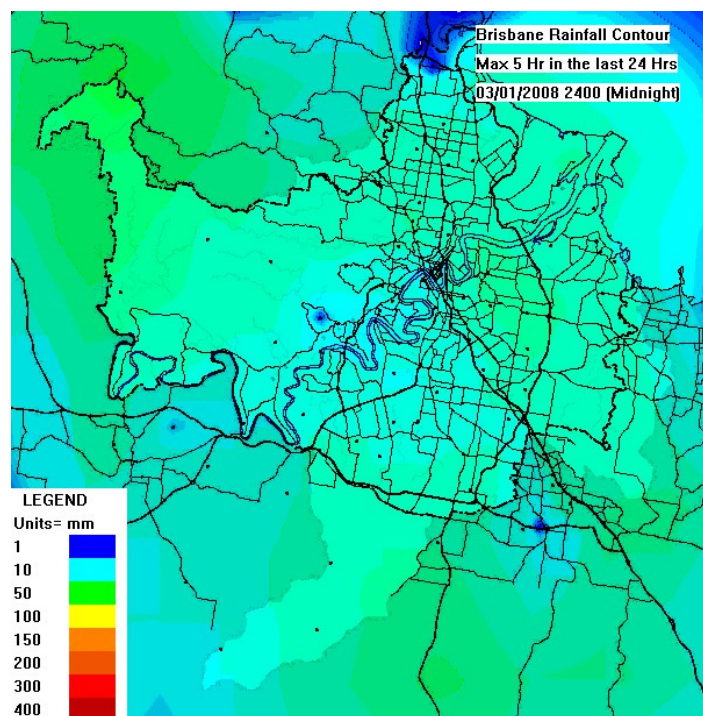


Figure 2.8 Brisbane Rainfall Contours for 03/01/2008

CHAPTER 3

LITERATURE REVIEW

3.0 LITERATURE REVIEW

3.1 Impervious and Pervious Areas

Urban catchments comprise of two distinct, hydrologically important areas that contribute to stormwater runoff through different mechanisms; impervious and pervious areas. Pervious areas allow infiltration and usually provide a greater degree of initial storage of runoff while impervious areas do not allow infiltration and usually have a small initial storage. As a result, impervious areas such as roads and roofs contribute runoff quicker and at a higher rate, whereas pervious areas such as parks and lawns require higher intensity or more prolonged rainfall to contribute similar amounts of runoff. (Boyd et al., 1993).

The connected nature of impervious area is of vital importance when describing rainfall, runoff and pollutant relationships. Directly connected impervious area (DCIA) refers to impervious areas with a dedicated connection to the drainage system such as roof runoff when conveyed directly to the street or underground drainage systems. Conversely, indirectly connected impervious area (ICIA) refers to impervious area not directly connected to the drainage system and either stored (such as with a rainwater tank) or conveyed to an additional pervious area.

Impervious and pervious areas also differ in their contribution of pollutants to receiving environments following rainfall events. The processes of accumulation and wash-off occur more frequently and easily from impervious areas. Automobiles also regularly frequent impervious areas such as roads, and drainage systems typically provide an excellent transport mechanism for

pollutants. By contrast, pervious areas generally infiltrate a large portion of the total rainfall, removing pollutants in the process. This phenomenon is particularly noticeable in small rainfall events, however it has been confirmed to still occur in events up to 75 mm when comparing forested and developed catchments (Christopher et al., 1997).

In addition to runoff quality, numerous studies have established a direct relationship between the amount of impervious surface within a watershed and pollution of its surface waters (Civco and Sleavinl, 2000). The degradation is largely due to a combination of increased runoff volumes and decreased water quality (Arnold and Gibbons, 1996). However, this change in catchment hydrology can be combined with the clearing of riparian vegetation and the straightening and formalising of drainage channels to result in severely degraded receiving environments.

3.2 Urban Drainage Systems

Drainage systems in urban areas are primarily designed for the effective conveyance of runoff into a receiving environment. Receiving environments vary from dry gullies, overland flow paths, creeks, rivers, groundwater storages, lakes and oceans. In addition to runoff conveyance, urban drainage systems can also be designed to provide retention, treatment and/or reuse of stormwater and also provide surcharge from sewerage systems. Components of an urban drainage system typically comprise of:

- Property drainage,
- Street drainage,
- Trunk drainage,
- Detention systems,

- WSUD elements.

Individual allotments comprise of portions of pervious and impervious area both of which are connected to the drainage system either directly or indirectly. Pervious area generally consists of surfaces such as lawns, gardens, grassed road reserve and parks. Impervious area includes roofs, driveways, tennis courts, sheds, sealed roads and footpaths. Roofs usually comprise the largest impervious area within individual allotments. Impervious area is either DCIA in the case of roof guttering conveying runoff directly to the street, or ICIA in the case of gutters connected to rainwater tanks or backyard ‘bubbler style’ outlets that discharge onto the lawn.

Property drainage subsequently flows into a street drainage system where the cumulative effect of the discharge from individual causes flows to be increased. Streets are usually drained by a combination of gutters, inlet pits, manholes and pipes, or using surface drainage through the use of swales and other WSUD measures. The use and application of these drainage elements is closely linked to the street and property design. Design of street drainage systems in Australia generally requires the capture of small, regular rain events (1 to 5 year) within the gutter and underground drainage system, and the conveyance of large, irregular events (50 to 100 year) within the entire road reserve (Figure 3.1).

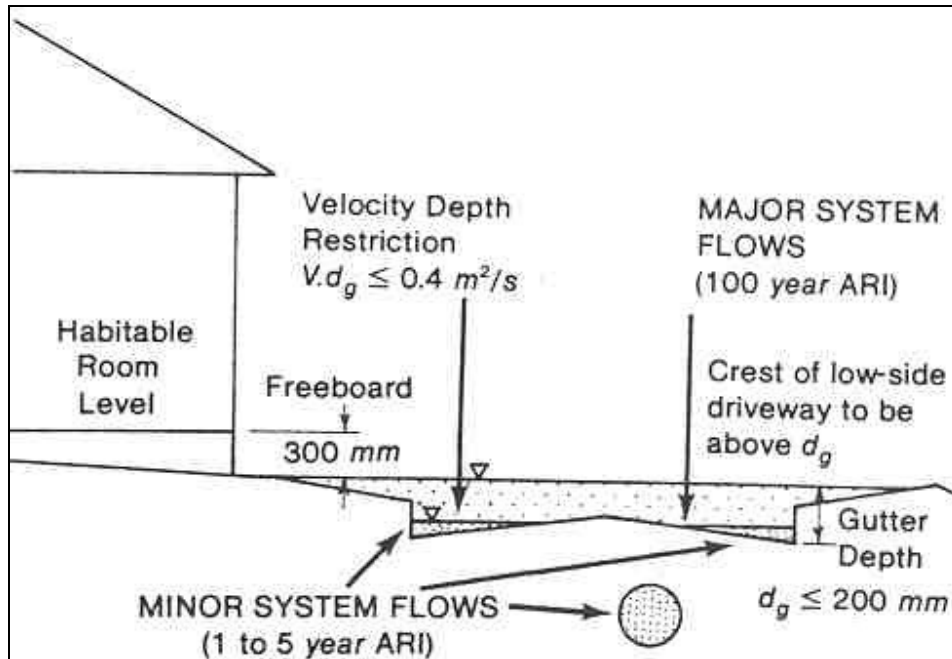


Figure 3.1 Design objectives of a street drainage system (USQ, 2003)

Trunk drainage in the form of underground pipes, open channels or a combination of both, convey the runoff from street drainage systems into the receiving environment. As with street drainage, the cumulative effects of a number of sub-areas contributing increases the size of the flowrates and thus the drainage infrastructure required.

Detention systems, otherwise known as detention basins or retarding basins are designed to store runoff volumes from a catchment for a pre-determined period of time. Detention basins are not to be confused with retention basins which are design to permanently hold water. These basins then release the runoff at a controlled rate that acts to limit the peak runoff in the downstream drainage system. The use of detention basins can be effective in areas where flooding immunity is a problem and can limit the size of trunk drainage required downstream and as a result provide cost savings. The use of detention basins is

now widely regarded as an ineffective use of urban space, and additional outcomes can be achieved while integrating community, open space, biodiversity and water quality objectives using the design principles associated with WSUD.

WSUD elements are secondary design elements integrated into the four components of urban drainage systems previously mentioned. The principles of WSUD are most successfully applied at the property drainage level as the cumulative effect of increased flowrates and pollutant loads further down the drainage system limits the opportunity for its effective use. As a result, applications of WSUD are typically divided into lot scale measures when referring to property drainage and 'end of pipe' measures when referring to street and trunk drainage (Healthy Waterways, 2006). WSUD elements generally involve a combination of retention, detention, infiltration and reuse and can include:

- Swales and buffer strips,
- Bioretention systems,
- Sediment basins,
- Wetlands,
- Infiltration measures,
- Storage and reuse.

3.3 Suspended Solids

Runoff from urban areas carries a wide variety of pollutants, including sediment, nutrients, organic matter, bacteria, oil and grease, toxic substances and heavy metals (NCDWQ, 2007). Suspended solids are small particulates present in water that remain in suspension due to colloidal mixing or the motion of the liquid and are usually measured by the total suspended solids (TSS) water quality analytical

measurement. Some pollutants such as phosphorous and heavy metals readily attach themselves to particulates, allowing suspended solids to act as a mobile substrate for other pollutants. TSS is not a good indicator of highly soluble pollutants from atmospheric sources such as nitrogen, which are of significant importance in urban areas.

The use of suspended solid loads as an estimate of pollution has been corroborated by numerous studies and has resulted in TSS being used as a widespread indicator of stormwater quality (Sartor et al., 1974; IEAust, 2003). However the ability of TSS to adequately measure suspended solid loads has been questioned (de Ridder et al., 2002; Brodie, 2006). The United States Geological Survey and the US Federal Highway Administration also mirror this view and have issued a policy establishing the suspended sediment concentration (SSC) as the preferred technique for runoff analysis (USGS, 2000).

This preference is largely due to the standardised nature of the SSC test and the fact that it limits bias. The TSS method involves the extraction of a sample from a container and subsequently passing the contents through a glass fibre filter that retains the solids. The mass of the retained solids is then weighed and extrapolated to the total volume to calculate the concentration. The SSC test involves calculating the mass of the entire sample negating the need for extrapolation (ASTM, 1999). This testing of the whole sample provides more accurate results especially if larger particles and fine sands are present. The presence of these particles can introduce bias if the TSS sub-sample is taken in such a way that discounts these particles, commonly through insufficient mixing times. This is particularly important if particle size distribution is an important component of the monitoring.

In order to determine pollutant loads, suspended solid concentrations obtained through either method can be multiplied by runoff. The simplest method for obtaining stormwater loads is through the use of the event mean concentration (EMC). The event mean concentration is defined as the total mass discharged divided by the total runoff volume for a defined event. This allows the mass loading to be calculated using the measured concentrations and runoff volume for individual storm events or on any defined time period.

3.4 Suspended Solids and Urban Areas

As with the majority of stormwater pollutants, suspended solid loads in receiving environments result from the accumulation and removal of particles on impervious surfaces such as streets, roofs and parking lots as well as localised soil erosion on pervious areas. As a result, the relationship between stormwater pollutants and urban areas is also applicable to suspended solid loads. Suspended solid loads are often highly correlated with the activities undertaken on the particular surface.

Suspended solids on urban surfaces undergo a cyclical process of accumulation and removal that is dependant on a number of variables. The process kinetics of accumulation and removal are often highly site specific and involve complex relationships that can be difficult to model in a changing urban environment (Ayoko, 2004). Simplified, accumulation generally occurs during periods of dry weather, and these accumulated particles are removed in periods of wet weather. Major sources of accumulation include atmospheric fallout, surface wear, soil erosion, and vehicle and tire wear.

It has been corroborated by numerous studies that the process of accumulation is not highly time dependent and occurs within a few days of wet weather (Vaze and Chiew, 2002; Deletic and Maskismovic, 1998). Brodie (2006) also discovered a 'rapid post-storm recovery' of particles on impervious surfaces and considered that the build-up of free particles reaches an equilibrium that is dependant on supply and removal process. These processes can include wind as well as activities involving cleaning surfaces such as roof, driveway and street cleaning. Numerous studies have conceded that either vehicle generated or external wind has a very significant effect on pollutant accumulation (Ball et al., 2002; Pitt et al., 2004; Vaze and Chiew, 2002). It is important to note that a number of other studies have found a relationship between the antecedent dry period (ADP) or time since last rainfall and particle load, suggesting a slower post-storm recovery (Ball et al., 1998). It has been noted by Pitt (2004) that an equilibrium accumulation condition depends on a number of environmental factors specific to the location. This suggests that post-storm recovery may occur fast or slow depending on the prevailing environmental conditions.

Accumulation has a tendency to reach an upper limit that is often dependant on the location and type of surface. This phenomenon has been noticed in a wide range of empirically based studies, which are generally highly specific to the particular environment. A typical study on the mass accumulation of TSS on eight highway sites in Southern California discovered a build-up of 0.544 g/m^3 per day (Jeong et al., 2006). Spatial variation has also been found when studying road surfaces, with over 90% of particles found within 30 cm of the kerb (Pitt et al., 2004)

As with accumulation of particles, the removal process of particles from impervious area is also quite complex. Removal is dependant on the available supply and size distribution of particulates, kinetic energy of the rainfall, capacity

of runoff to mobilise particles and the surface characteristics (Pitt et al., 2004; Ayoko et al., 2004). Brodie (2006) also discovered that the use of a rainfall detachment index (RDI) improved the relationship with suspended solid loads.

Runoff from surfaces frequented by vehicles has often been the major focus of a large number of studies on surface water quality. Although road surfaces generally comprise only 30% of the DCIA of an urban catchment they represent a major source of suspended solid loads (Ball, 2002). This stems from the view that a large portion of stormwater pollutants are generated through vehicular motion and other activities associated with the use of roads. In addition, small rainfall events generate sufficient depth of flow to mobilise particles on roads that are efficiently conveyed into the drainage system.

Several studies have attempted to develop indicators to describe and predict the relationship between vehicular use and runoff quality. Borroum et al. (2003) concluded that there is a link between annual average daily traffic and highway runoff constituents. However, it was noted that the influences of accumulation and removal as well as catchment characteristics outweighed any potential direct correlation between increases in traffic volumes and resultant increases in TSS concentration. These findings were mirrored by Tomlinson et al (1999) who could not find any direct relationship between TSS concentration and factors such as traffic volume, despite large difference in median concentrations (60 to 1925 mg/L). This variability is complicated by the transport of particles, with over 95% of the sediment on any given highway originating from sources other than the vehicles themselves (Shaheen, 1975).

Parking lots and driveways are also a surface frequented by vehicles, however measured TSS concentrations tend to be lower than that of highways and urban streets. In areas of heavy commercial traffic such as service stations, suspended

solid loads can be high. Rabanal and Gizzard (2000) monitored runoff from several commercial sites and discovered that while loads were variable, the highest were encountered at service stations when compared to restaurants and office parking lots. Results from Neary et al (2002) also proved that loads were extremely variable with smaller catchments providing the most irregularity.

Roofs comprise a significant portion of the impervious area of urban catchments however sediment concentrations are generally lower than roads due to differing processes of accumulation and removal. The majority of particles in roof runoff are fallout from atmospheric build-up and as a result the ambient concentration of atmospheric particles has a major influence, as does the roof surface. Brodie (2006) measured event mean concentrations of TSS at approximately 15 times smaller than road runoff. In a review of worldwide literature Duncan (1995) found TSS concentrations in urban areas approximately 10 times smaller for roofs than road runoff respectively.

Despite the large number of studies focussing on road runoff, pervious surfaces can contribute large volumes of sediment and are generally regarded as the largest source of sediment in waterways. This is highly dependant on vegetative cover and is largely due to localised soil erosion (Ball, 2002). Brodie (2006) discovered that for areas of bare soil, TSS concentrations were similar to that of road surfaces however runoff was less due to the pervious nature of the surface. This is due to the fact that bare soil detaches easily from a surface when subjected to rainfall, a characteristic that has been confirmed by numerous studies (Kinnel, 1997). Other studies have found a closer correlation between soil erosion and runoff compared to soil erosion and rainfall (Linhong, 2007). Grassed areas typically have the lowest sediment load due to the ability of the surface to trap particles and the higher energy required to detach particles. In

fact, grassed surfaces in the form of swales are commonly used as a stormwater treatment system to reduce sediment concentration to background levels.

3.5 Urban Stormwater Drainage Models

A variety of models exist for the prediction of the stormwater runoff from both a quality and quantity perspective. This review has been restricted to models that are used in the analysis of urban drainage systems and in particular the DRAINS modelling software that will be used in this project.

A variety of urban drainage computer models are used for the design and investigation of drainage systems both in Australia and internationally. A review undertaken by Dayaratne (2000) determined that the most popular computer models used for drainage design include:

- ILSAX / DRAINS (Watercom, 2000),
- RAFTS (WP Software, 1991),
- RAT-HGL (WP Software, 1992),
- CIVILCAD (Surveying and Engineering Software, 1997),
- RORB (Laurenson and Mein, 1990),
- SWMM (U.S. EPA, 1992),
- WBNM (Boyd et al., 2000),
- HEC-RAS (Hydraulic Engineering Centre, 2000),

Many of these computer models use hydrologic and hydraulic calculations involving loss modelling, and interchangeable pipe and overland flow routing. These models differ in their time structure and can be described as either event-based or continuous. Continuous models calculate losses through continual soil moisture and evaporation computations on a daily basis. Event-based models are

required to be constructed for a specific event, and generally do not take into account the effects of evaporation. Soil moisture accounting in event-based models is also limited, with inputs usually requiring some degree of user subjectiveness.

In addition to computer models, a procedure known as the statistical rational method (SRM) has widespread usage for the determination of peak discharge in simple situations. The SRM is known to incorrectly determine discharge in catchments where different land uses, irregular shaped catchments and inconsistent topography complicate calculations. This method also relies on a number of major assumptions that can limit its application.

DRAINS is a rainfall-runoff simulation program with uses in designing and analysing urban drainage systems. DRAINS combines hydrological and hydraulic calculations including routing in an event-based format. DRAINS can utilise either a modified version of the SRM that incorporates continuing soil losses during rainfall events or the ILSAX hydrologic model for calculation of discharge. In the ILSAX model a catchment is divided into a number of sub-catchments which are further subdivided into three sub-areas (Figure 3.2); a paved area directly connected to the drainage system (DCIA), pervious area, and a supplementary impervious area not directly connected to the drainage system (ICIA). This approach allows separate hydrological parameters to be applied to each surface with the resultant hydrograph from each sub-catchment a sum of the hydrograph in each sub-area (O'Loughlin and Stack, 2008).

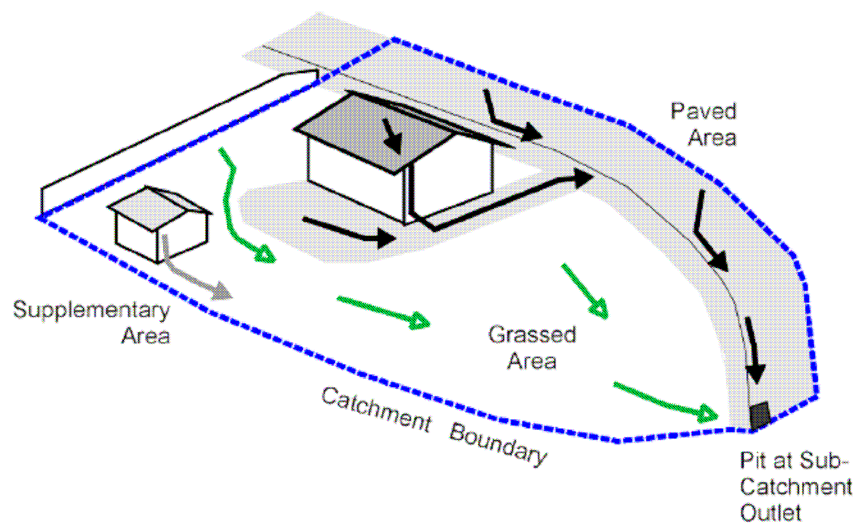


Figure 3.2 ILSAX catchment model land-use types (O'Loughlin and Stack, 2008)

To apply infiltration losses, the ILSAX model utilises an initial and continuing loss approach using Horton infiltration curves. The initial losses are specified by the modeller for each sub-area while the continuing losses for pervious area are calculated from the infiltration curves. These curves are based on a starting antecedent moisture condition (AMC) that is determined from the amount of rainfall in the past five days. The infiltration capacity is then a function of the time after the start of the rainfall event (Figure 3.3). The Horton curves have been developed by the U.S. Department of Agriculture and tested widely in North America. The determination of soil type is the subjective choice of the modeller and it has been suggested that the use of U.S. soil classifications to characterise Australian catchments is somewhat untested (Dayaratne, 2000).

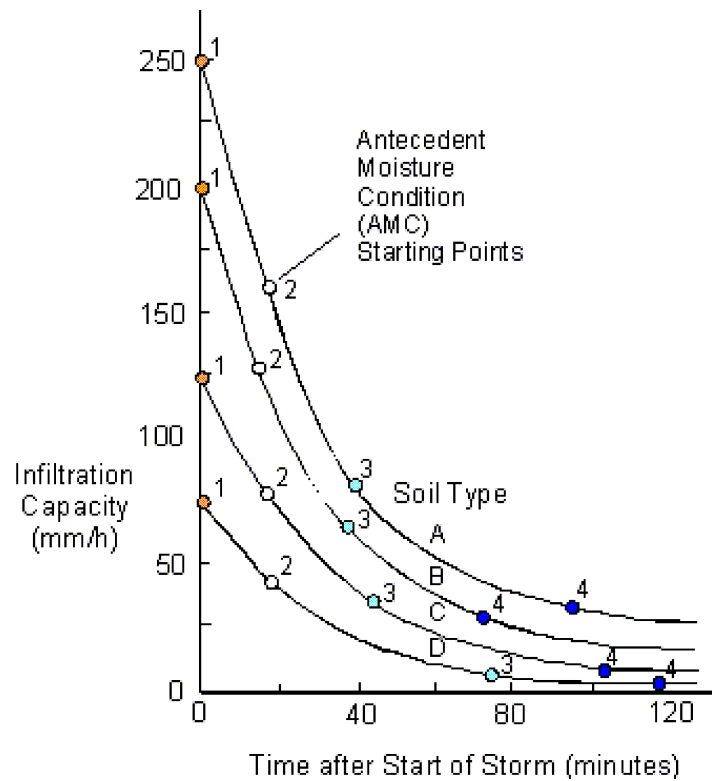


Figure 3.3 Horton Infiltration Curves (O'Loughlin and Stack, 2008)

The summing and routing of individual sub-areas is then combined with hydraulic computations from which lag times are determined and the discharge at individual nodes calculated (Figure 3.4). The routing through pit and pipe systems occurs in conjunction with surface overflows and additional input from adjacent sub-areas. These complex interactions can result in very different results when comparing storms with different durations and rainfall intensities. DRAINS also incorporates a link with HEC-RAS software to model flow relationships through open channel systems before discharging into the receiving environment. (O'Loughlin and Stack, 2008).

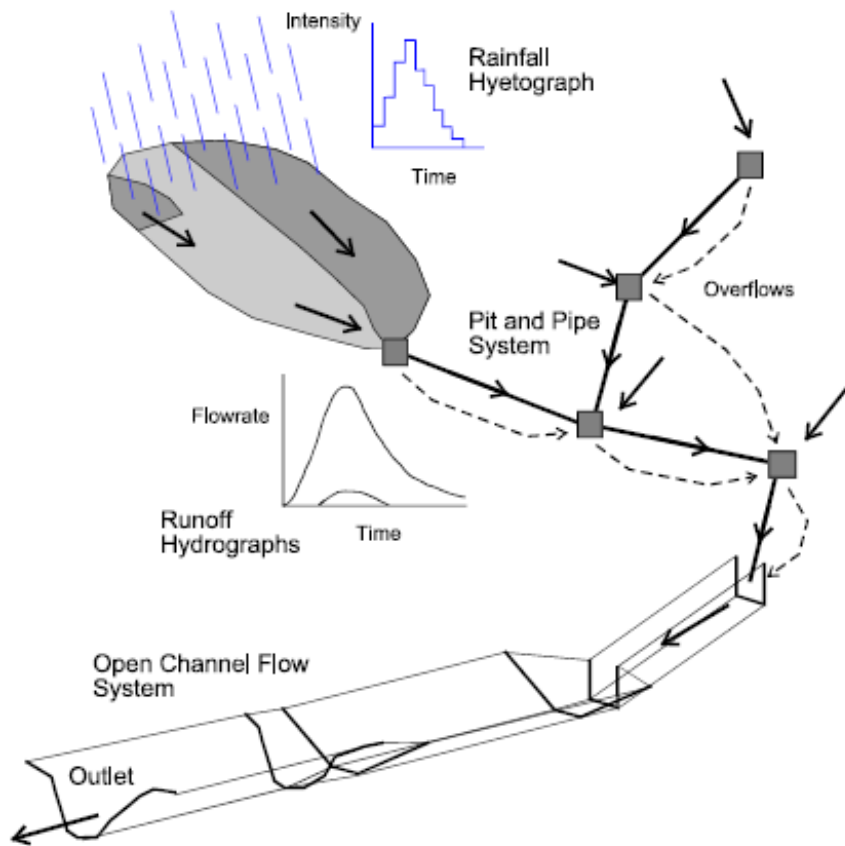


Figure 3.4 Representation of complete DRAINS modelling process (O'Loughlin and Stack, 2008)

CHAPTER 4

DATA COLLECTION AND ANALYSIS

4.0 DATA COLLECTION AND ANALYSIS

4.1 Drainage System

Data available to describe the drainage system has been collected from BCC spatial information software known as iBimap, 'as constructed' plans as well as on-site site visits. Information was obtained from iBimap and compared with available plans before a number of site visits were conducted to confirm the accuracy of this data. Council iBimap records along with plans were found to be in error and measurements were taken to correct the information.

Given the size of the Hoyland Street catchment, the drainage is not overly complex and typical of a pit and pipe drainage system. On-grade and sag inlets capture runoff which is conveyed in pipes through a series of pits that subsequently discharge into the bioretention system. Overland flow paths from each catchment area do not discharge into the bioretention system when pit capacities are reached. The entire drainage system has been constructed in 2002 as part of the aforementioned Hoyland Street connection linking Gympie Road and Bracken Ridge Road.

The iBimap software displays stormwater infrastructure visually and has the option of linking to background metadata where a variety of detailed information is available. This process involved in entering stormwater infrastructure into the iBimap system creates inconsistencies in how gully pits are linked to the trunk drainage system. The result of this is that for gully pits are shown as linked directly to the closest adjacent stormwater pipe. In addition, no metadata is available for gully pit connections such as size of pipe or invert levels. Despite

this shortcoming, metadata includes a variety of information, however data pertaining directly to drainage system analysis includes:

- Eastings and Northings of pits and chambers,
- Surface levels and invert levels of pits and chambers,
- Chamber and pit sizes and types,
- Pipe lengths,
- Downstream and upstream pipe invert levels,
- Pipe sizes and types, and
- Number of pipe barrels

As constructed plans of the Hoyland Street connection include detailed drainage construction plans that have been amended following construction. Surface and piped drainage plans as well as catchment plans are available in Appendix C. A review of these plans identified the discrepancy between the orientation of gully pits and connecting pipes with iBimap. This was evident in the upper portion of HSA catchment where two on-grade inlets at the corner of Coriander and Kluver Street were displayed as being directed in opposing directions. A number of pipe and pit invert levels as shown in iBimap were inconsistent with the detailed plans.

Site visits conducted in May and June 2007 concluded that both sources of information were incorrect and that the two gullies drained into a 375 mm diameter reinforced concrete pipe (RCP) toward the North, external to the Hoyland Street catchment. Discrepancies between pit and pip levels were also checked which confirmed the accuracy of the drainage plans. A number of discrepancies in the surface drainage regime as indicated on the plans were also noted including road crossfall and locations of road crowns. These were noted and new measurements taken using a digital level.

Figure 4.1 below, indicates the location of all amended stormwater infrastructure within and external to the Hoyland Street catchment (Appendix D). Manholes are represented by circles, gully pits are represented by squares and the extent of the catchment area is represented by the blue line. It can be seen that the majority of the catchment runoff is captured by gully pits on and adjacent to Hoyland Street and conveyed into a branched trunk drainage system before discharging into the bioretention system through HSA. HSB collects a small portion of the road runoff downstream of the HSA network while HSC collects a large portion of the road runoff south of the bioretention system through a single gully pit.

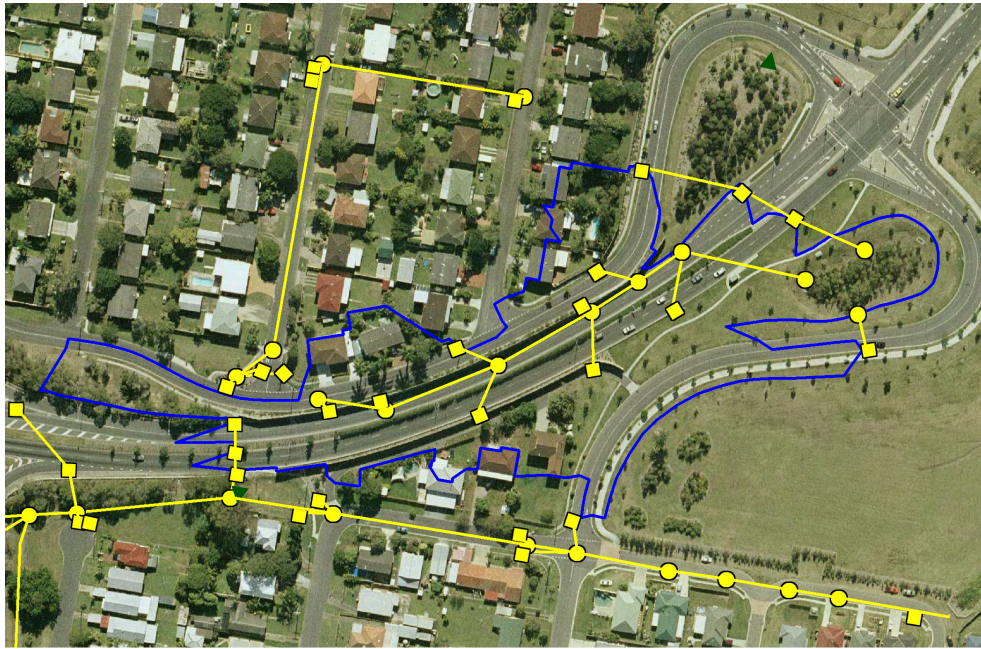


Figure 4.1 Location of Stormwater Infrastructure within the Hoyland Street catchment

4.2 Catchment Characteristics

The vast majority of the Hoyland Street catchment is comprised of road runoff which discharges through piped drainage directly into the Hoyland Street system. A portion of rainfall also enters the bioretention system directly as well as sheet flow runoff from the grassed area surrounding the bioretention system. Figure 4.2 depicts the HSA, HSB and HSC catchments discharging into the Hoyland Street bioretention system (reproduced in Appendix E).



Figure 4.2 Catchments discharging into the Hoyland Street bioretention system

Catchment areas have been delineated using MapInfo version 8.0 GIS software using a combination of aerial photography, contours, detail plans and airborne laser survey (ALS) data. Site visits also proved beneficial to determine the extent

of directly connected roof area, supplementary area as well as other lot based connections. Initial rough catchments were defined using high-resolution 2005 aerial photography and 0.5 m contours. These catchments were subsequently refined to determine hydrological characteristics for DRAINS and surface characteristics for the mass balance model. This was achieved with the use of detail plans as well as ALS data. ALS data has been obtained from scanning conducted from 15 to 20 June 2002 and surface levels to a 1 mm resolution (± 200 mm) are directly available at 1 m intervals. The use of this data proved necessary in the delineation of the boundary between catchments on very flat ground adjacent to the bioretention system and to the south of the catchment. The detailed 'as constructed' plans provided valuable information on the road grading and location of crowns. The 2002 and 2005 data sets have been assumed as applicable to the catchment conditions over the entire monitoring period.

The HSA catchment at 1.034 ha is the largest catchment discharging into the bioretention system. DCIA, pervious and supplementary areas comprise approximately 54%, 43% and 3% respectively. While not the largest percentage, this represents the greatest area of pervious and supplementary area in all catchments. The effect of this pervious and supplementary area is the delay of the runoff contribution from these areas. In many rainfall events it has been discovered that this results in a second peak, where the DCIA produces an initial runoff response that is followed by the other contributions well after the peak. The large elongated catchment shape also tended to result in an extended falling limb of measured hydrographs.

Table 4.1 Hoyland Street catchment hydrological characteristics

Catchment	DCIA	Pervious	Supplementary	TOTAL
	Area	Area	Area	Total Area
	ha	ha	ha	ha
HSA	0.564 54.56 %	0.440 42.55 %	0.030 2.89 %	1.034
HSB	0.147 64.60 %	0.074 32.64 %	0.006 2.76 %	0.228
HSC	0.102 34.35 %	0.150 50.41 %	0.045 15.24 %	0.297
HSO	0.000 0.00 %	0.249 96.39 %	0.009 3.61 %	0.258
TOTAL	0.813 44.76 %	0.912 50.24 %	0.091 5.00 %	1.816

HSB catchment represents the smallest catchment at 0.228 ha with the largest portion of impervious area. DCIA, pervious and supplementary areas comprise approximately 64%, 33% and 3% respectively. The high imperviousness of the catchment and proximity to the bioretention basin results in a fast runoff response. In the majority of consistent runoff events, HSB recorded the earliest response time and time to peak runoff. The relatively uniform, rectangular catchment shape as well as the high imperviousness also tended to result in a faster, almost linear falling limb of the hydrograph.

HSC catchment is slightly larger than HSB at 0.297 ha with the majority of the catchment as pervious area. DCIA, pervious and supplementary areas comprise approximately 34%, 51% and 15% respectively. As a result HSC has the largest proportion of pervious and supplementary area of all catchments, due to its proximity to adjacent parkland and bikeways. Recorded hydrographs tend to exhibit a more traditional ‘bell curve’ shape when compared to HSA and HSB. This may be due to the crescent shape and flat grade of the catchment providing

additional lag time. The flat grade may also be the contributing factor behind the comparatively slow response times evident in measured hydrographs despite the proximity of the catchment to the bioretention system and the very short length of connecting pipe.

4.3 Rainfall Data

Hyetographs for rainfall events within the BCC local government area are available for locations citywide due to an extensive network of pluviometers combined with a real-time data management system. As previously explained, rainfall data is available from 6 pluviometers that are within 7 km of the Hoyland Street bioretention system (Table 4.2). Unfortunately data is not available for the entire monitoring period at all pluviometer locations.

All pluviometers are of a tipping bucket design with a 1 mm resolution allowing small rainfall events to be recorded. The closest rainfall gauge (BDR712) has not collected data for the entire monitoring period. Subsequent investigations indicated that the data obtained from this gauge was questionable and was taken offline in 2003 due to error with the gauge itself. Plots of the remaining rainfall gauges when compared to three samples of runoff indicated that pluviometer BDT839 and MBR752 provided the best correlation with runoff response. Pluviometer BDR839 located at the Bracken Ridge reservoir, 1.7 km from the catchment was subsequently chosen due to its closer proximity to the catchment. A total of 16 storms were available that produced catchment runoff.

Table 4.2 Pluviometers within 7km of the Bioretention System (Brisbane City Council)

Location	Pluviometer	Distance to Site (km)	Collection Period	
			From	To
Bald Hills Ck at Bracken Ridge Rd, Bracken Ridge	BDR712	1152	7/06/1994	21/11/2003
Bracken Ridge Reservoir, Jude St Bracken Ridge	BDR839	1772	21/02/2005	Present
Moreton Bay at Brighton Bowls Club	MBR752	4761	1/12/1999	Present
Albany Ck at Pinnaroo Cemetery,	A_R842	4102	22/03/2005	Present
Cabbage Tree Ck at Braun St, Deagon	C_R560	4642	11/06/1994	Present
Little Cabbage Tree Ck at Aspley Reservoir, Aspley	LCR566	6264	22/06/1994	Present

Given the size of the catchment and the distance of rainfall gauges from the centroid of the catchment, difficulties arise in the applicability of the measured rainfall data to the modelling of runoff. Ideally a pluviometer should be located as close as possible to the centroid of a catchment to enable meaningful correlation with the timing and volume of runoff. This is largely due to the spatial variation in rainfall events and the relatively small size of the catchment.

Due to the questionable nature of the applicability of the rainfall data to the Hoyland Street catchment, other techniques have been employed to ‘weed out’ inconsistent data and only use event with an acceptable correlation with rainfall. As a first pass, the rainfall contour maps previously mentioned were used where available to determine any noticeable discrepancies in the spatial variation of rainfall in proximity to the gauges and catchment. This resulted in a total of five events being removed due to obvious variability in the rainfall patterns. While this test proved inherently subjective, a method was required to quickly eliminate inconsistent data. Subsequent plots comparing two of these storms to measured runoff confirmed that the volume and timing of these events did not correlate at all to runoff.

Events suitable for the modelling and calibration of the DRAINS model should exhibit a high degree of correlation with the timing and intensity of rainfall. In order to maintain data integrity and it was necessary to analyse the rainfall patterns and available runoff hydrographs for each specific event. This process involved the graphical representation of discharge (L/min) and rainfall (mm) for storms and individual burst events. Hydrographs for HAS, HSB and HSC inlet pipes as well as HSOa and HSOB outlet pipes were plotted from the start of the rainfall to the final recession of the longest recorded hydrograph. This intensive screening process provided the ability to select events that indicated a direct and quantifiable correlation with measured runoff. Additional data collected to improve event selection and subsequent DRAINS analysis included the:

- Time period,
- Total precipitation,
- Precipitation in past five days,
- Precipitation in past twelve hours, and
- AMC condition.

Figure 4.3 displays the output of this process for Storm 6, burst event 4. The complete rainfall-runoff analysis graphs for individual events are available in Appendix B. It was discovered that only two events were suitable for calibration with the DRAINS due to the large data inconsistencies previously described; storm 2 (event 2) and storm 3 (event 1). The vast majority of the events exhibited runoff hydrographs that did not correlate at all to rainfall, indicating that rainfall was not occurring in the catchment at the same time as gauge BDR839. HSC was found to be the outlet with minimal error due to negative flowrates and in some situation HSB also provided acceptable hydrograph responses. The exercise did provide insight into each typical catchment response with HSA providing a dual

peak hydrograph for most events indicating a definite lag between the pervious and impervious area contribution times.

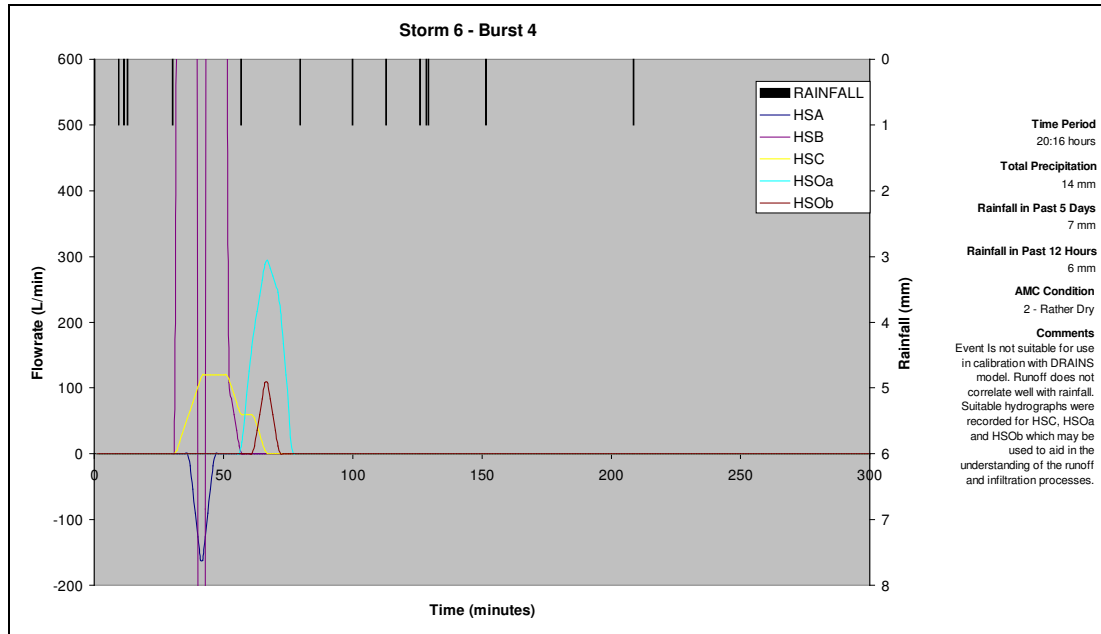


Figure 4.3 Measured rainfall-runoff screening graph for Storm 6, Event 4.

For application of the mass balance model a larger number of events proved acceptable, since event averages are required as opposed to whole of event relationships. A good correlation between measured rainfall volume and runoff volume should be evident for the specific storm event. Since the specific hydrograph response is not as important a total of 11 events (Appendix B) were suitable for application to the mass balance model.

4.3.1 Storm 1

Storm 2 occurred from 12:31am to 6:40am on 12/03/2006 to 6:40pm and consisted of a single burst event. This event is not suitable for use in calibration

with DRAINS model. A correlation with the rainfall is evident however the timing indicates gauge and monitoring data inconsistencies. Hydrograph shapes are consistent with an expected response however the relative magnitude of each response is inconsistent with catchment characteristics. The measured runoff shows a response in HSC catchment in excess of the contribution in HSA and HSB, which is unlikely to be correct.

4.3.2 Storm 2

Storm 2 occurred from 1:44am 23/03/2006 to 10:21pm 24/03/2006 and consisted of a total of 4 events with a single suitable modelling event observed. Event 1 is not suitable for use in calibration of DRAINS model. This is due to the low runoff response that that would be expected from a dry catchment. The data also features negative flowrates for HSB and HSA. Event 2 is suitable for use in calibration with DRAINS model. A high correlation between rainfall and runoff is evident. Suitable hydrographs were recorded for HSC, HSOa and HSOB. The data also features negative flowrates for HSA and HSB. Event 3 is not suitable for use in calibration with DRAINS model. Correlation between rainfall and runoff is evident however not consistent. Event 4 is not suitable for use in calibration with DRAINS model. A high rainfall-runoff correlation is apparent, however wildly exaggerated hydrographs and negative flowrates are also features of this data.

4.3.3 Storm 3

Storm 3 occurred from 10:25am to 3:43pm on 31/03/2006 and consisted of a single burst event, which also provided a suitable modelling event. Event 1 is suitable for use in calibration of DRAINS model. There is a high correlation between rainfall and runoff. A suitable hydrograph was recorded for HSC

however wildly exaggerated hydrographs and negative flowrates are also features of this data.

4.3.4 Storm 4

Storm 4 occurred from 6:00pm on 04/04/2006 to 4:18am on 06/04/2006 and consisted of 2 burst events, both of which were not suitable for calibration of the DRAINS model. Event 1 is not suitable for use in calibration with DRAINS model however can provide insight into the time of concentration for low flows. Although HSB is exaggerated, HSA appears relatively accurate and exhibits a high degree of correlation with the rainfall. Event 2 is not suitable for calibration. Flowrates are only available for HSA and are obviously inaccurate.

4.3.5 Storm 5

Storm 5 occurred from 5:02pm to 11:42pm on 30/04/2006 and consisted of 2 burst events, both of which were not suitable for calibration of the DRAINS model. Event 1 is not suitable for use in calibration with DRAINS model however can provide insight into the time of concentration for low flows. Although HSB is exaggerated, HSA appears relatively accurate and exhibits a high degree of correlation with the rainfall. Event 2 is not suitable for use in calibration with DRAINS model. While the runoff appears to correlate well to the first two events, a third rainfall event does not produce any runoff. The data may be used to aid in the understanding of the runoff response for catchment C and the initial response for catchment A.

4.3.6 Storm 6

Storm 6 occurred from 4:25pm on 09/06/2006 to 11:41pm 11/06/2006 and consisted of 4 burst events, none of which were not suitable for calibration of the DRAINS model. Events 1 and 2 are not suitable for use in calibration with DRAINS model. Runoff correlates very poorly with rainfall. Event 3 is not suitable for use in calibration with DRAINS model. Runoff appears to occur before the rainfall is registered suggesting a poor correlation. Hydrographs for HSC and HSOa appear to be consistent with an expected response. Event 4 is not suitable for use in calibration with DRAINS model. Runoff does not correlate well with rainfall. Suitable hydrographs were recorded for HSC, HSOa and HSOB, which may be used to aid in the understanding of the runoff and infiltration processes.

4.3.7 Storm 7

Storm 7 occurred from 1:13pm to 11:17 pm on 26/06/2006 and consisted of a single burst event. This event is not suitable for use in calibration with DRAINS model. A very poor correlation between rainfall and runoff is evident

4.3.8 Storm 8

Storm 8 occurred from 10:26pm on 29/06/2006 to 12:22am 30/06/2006 and consisted of a single burst event. This event is not suitable for use in calibration with DRAINS model. No runoff has been recorded despite rainfall occurring at the gauge. This was perhaps due to a dry catchment condition limiting runoff.

4.3.9 Storm 9

Storm 9 occurred from 1:50am to 6:42am on 16/07/2006 and consisted of a single burst event. This event is not suitable for use in calibration with DRAINS model. Poor correlation with rainfall is evident and wildly varying negative and positive flowrates indicate instrument error.

4.3.10 Storm 10

Storm 10 occurred from 1:41am to 6:32am on 25/07/2006 and consisted of a single burst event. This event is not suitable for use in calibration with DRAINS model. Poor correlation with rainfall is evident. This is coupled with negative flowrates for both HSA and HSB. Instrument error has caused no suitable hydrograph to be recorded for HSC, HSOa and HSOB. The two distinct responses for HSA indicate separate pervious and impervious area responses and may provide information regarding the pervious area time of concentration for input into the DRAINS model.

4.3.11 Storm 11

Storm 11 occurred from 12:49am on 28/07/2006 to 12:01am on 29/07/2006 and consisted of 2 burst events, none of which were not suitable for calibration of the DRAINS model. Event 1 is not suitable for use in calibration with DRAINS model. Poor correlation with rainfall is evident. Runoff has only been measured at HSA, where negative flowrates are a feature. Event 2 Event is also not suitable for use in calibration with DRAINS model. Very poor correlation with the start of the rainfall event is evident. Runoff has only been measured at HSA, where the hydrograph shape is not indicative of a typical catchment response.

4.4 Monitored TSS Data

Suspended solid concentrations and loads for a large number of storm events between 2002 and 2006 are available for the bioretention system and have been testing using the TSS method. Monitoring ceased in 2006 due to the inconsistencies in flow measuring data previously discussed resulting in an inability to calculate loads. Data quality of TSS concentration samples however is expected to be good and will be used for the selected events.

Discrete flow-weighted water quality samples have been collected using ‘Sigma 900 Max’ automatic samplers. Samples were collected at each inlet pipe (HSA, HSB, and HSC) as well as one of the outlet pipes (HSO). This single monitoring of the two outlet pipes made the assumption that equal flows would occur in both pipes. The sampler is triggered by flows of $0.01\text{m}^3/\text{s}$ for the inlet pipes and $0.05\text{m}^3/\text{s}$ for the outlet pipe. Samples of 500mL were gathered for every consecutive 12.5m^3 of runoff. As a result the sampling time and quantity varies with the inflow volume, and is unique for each storm event. Table 4.3 below displays the sampling configuration and characteristics for the bioretention system.

Table 4.3 Hoyland Street bioretention system sampling locations and characteristics

Site ID	Full Site Name and Sampler Location	Sampler Type	Trigger Volume (m ³ /s)	Volume between Collections (m ³)
HSA	Western inlet pipe draining Catchment A	SIGMA 900MAX with integral flow meter	0.01	12.5
HSB	Northern inlet pipe draining Catchment B	SIGMA 900MAX with integral flow meter	0.01	12.5
HSC	Southern inlet pipe draining Catchment C	SIGMA 900MAX with integral flow meter	0.01	12.5
HSO	Outlet pipe	SIGMA 900MAX with integral flow meter	0.05	12.5
HSL	Level Sensor, adjacent to outlet pipe.	Greenspan PS310 level logger	N/A	N/A

Table 4.4 below indicated the average TSS concentrations from each catchment over the extent of the monitoring period. Table 4.5 indicates TSS event mean concentrations for available events. HSA exhibits the highest mean TSS concentration of 74.6 mg/L which is followed by HSC with 56 mg/L and HSB with 50 mg/L. This distribution of concentrations is consistent with the surface types in each catchment as well as the dominant activity of each surface.

Table 4.4 Hoyland Street TSS inflow concentration (average over monitoring period)

	Site	TSS mg/L
HSA	Mean	74.639
	Median	58.214
	C.V. (StdDev/Mean)	22.125
	Minimum	137.400
	Maximum	0.605
	10%ile	34.013
	90%ile	136.440
HSB	Mean	50.065
	Median	42.498
	C.V. (StdDev/Mean)	16.667
	Minimum	160.000
	Maximum	0.847
	10%ile	18.227
	90%ile	78.640
HSC	Mean	55.992
	Median	46.500
	C.V. (StdDev/Mean)	14.800
	Minimum	98.000
	Maximum	0.514
	10%ile	28.705
	90%ile	96.200

The significance of Hoyland Street as a regional link between Gympie Road and Bracken Ridge Road results in high traffic volumes. In addition to traffic volumes, the steep grade of Hoyland Street as traffic approaches the intersection is likely to result in increased brake dust and tire wear accumulation in HSA catchment. The location of a sound barrier on both sides of this narrow ramp is likely to limit the ability for wind to remove accumulated particles. Traffic also needs to slow and turn 90 degrees on Parer Street approaching Hoyland Street in catchment HSC. The relatively straight sections of road draining into HSB may slow or limit the extent of pollutant accumulation, despite this catchment having the largest percentage of impervious area. Given that HSA catchment is at a steep grade, this may also provide an additional removal mechanism for particles when

comparing to the flat catchments of HSB and HSC. As a result the relative distribution of TSS concentration appears to be consistent with surface type as well as use.

Table 4.5 Monitored rainfall events and TSS event mean concentrations

	Inlet Pipe	TSS mg/L
8/12/06	HSA	69
	HSB	76
	HSC	60
4/01/06	HSA	190
	HSB	130
	HSC	220
12/02/06	HSA	170
	HSB	50
	HSC	180
17/02/06	HSA	69
	HSB	76
	HSC	60
2/03/06	HSA	100
	HSB	320
	HSC	140
5/03/06	HSA	31
	HSB	32
	HSC	15
23/03/06	HSA	190
	HSB	260
	HSC	170

CHAPTER 5

MODEL CALIBRATION AND APPLICATION

5.0 MODEL CALIBRATION AND APPLICATION

5.1 DRAINS MODEL

5.1.1 Model Structure

To set up the DRAINS model structure, catchment subdivision is necessary into the pervious, DCIA and ICIA sub-areas as previously explained. This catchment subdivision occurs at the catchment draining into each gully pit. A number of catchment, A through to N, were delineated and each sub-area plotted. These surfaces can be seen in Appendix I.

Flow paths as well as estimates for the DRAINS parameters then require input into the model. In addition, pipe and gully characteristics as well as rainfall characteristics and soil types as well as the AMC are required. The resultant characteristics output file is shown in Appendix G. The display of the set up DRAINS model structure following a model run is shown in Appendix H.

5.1.2 Sensitivity Analysis

A major limit of the DRAINS modelling software is that it does not have the ability to calibrate a modelled hydrograph based on measured data. The ability to calibrate the model to the measured data to obtain accurate event volumes is essential in determining the event pollutant loads. As such, a subjective trial and error approach was undertaken, where a number of variables were modified

and the results recorded. As a result of this analysis the most sensitive variables for a number of criteria were obtained and were altered to provide a better fit of the model.

Due to the nature of trial and error calibration a comparison between the goodness of fit of measured and modelled hydrographs will invariably lead to the model having a tendency toward a particular hydrograph aspect. These can include peak discharge, hydrograph shape, hydrograph recession and runoff volume.

An additional degree of subjectiveness involves the limit of the sensitivity analysis. Since accurate measured hydrographs are only available for HSC (Deletic and Fletcher, 2006) and due to the limit of available events that correlate with rainfall the sensitivity analysis can only be applied to catchment HSC. As a result, the outcome of the analysis will require the selection of parameters not only for catchment HSC but also for catchment HSA and HSB. Consequently, this method assumes that the parameters that require calibration in HSC are also the parameters that require calibration in the other two catchments as well. While this approach is far from ideal, it is necessary due to the aforementioned inconsistencies in measured runoff data. This sensitivity analysis was applied to burst 2 of storm 2 as it proved the most ideal event for calibration due to a high correlation with rainfall and runoff.

Table 5.1 DRAINS parameters for catchment HSC

Catchment HSC					
Storm 2: Burst Event 2					
Total Area		ha	0.297		
			Paved	Supplementary	Grassed
Percentage of Area	%	34.35	15.24	50.41	
Additional Time	min	0	5	15	
Flow Path Length	m	150	2	29	
Flow Path Slope	%	1	1.7	1.7	
Flow Path Roughness		0.015	0.015	0.13	
Area Depression Storage	mm	1.5	1	3	
Lag Time	min	10	-	-	

Pressure Loss Coefficient (Ku)	4
--------------------------------	---

Soil Type	4
AMC	2
Storm	2
Burst	2
Intensity averaging	5 min

Table 5.1 above displays the parameters required by DRAINS to model the catchment. Parameters that can be modified and will be subjected to the sensitivity analysis are shown in bold red. The sensitivity analysis required successive runs of DRAINS and the resultant hydrograph output was recorded. A total of 73 runs were performed analysing 15 parameters with a minimum of five runs for each parameter to graphically represent trends. The sensitivity of each parameter to changes in peak runoff, total volume, time to peak, and correlation (R), which attempts to measure hydrograph shape, were recorded. A search and move approach was trailed for each initial adjustment to determine an appropriate direction of movement. Hydrograph aspects for each parameter were also graphed if they proved sensitive. A discussion on each parameter

below is supplemented with the summary results table and graphed relationships in Appendix J and Appendix K respectively.

Impervious Area Lag Time

Impervious area lag time showed a decreasing correlation and decreasing peak discharge with an associated increase in value. Time to peak also increased with additional lag time as would be expected. The parameter did not have any impact on runoff volumes. This indicates that any modification to the existing value of 0 is likely to result in a less accurate model. This would be expected as there are very few obstructions to flow in the kerb and channel in the catchment.

Impervious Area Flow Path Slope

The impervious area lag time exhibited a decreasing time to peak and an increase in peak discharge with an associated increase in value. This response is expected as an increase in the grade of an impervious area with minimal surface roughness will increase velocities. The parameter did not have any impact on runoff volumes.

Impervious Area Roughness

The impervious area roughness value which relates directly to the Manning's roughness value provided a decrease in peak discharge with an increase in value. Time to peak also increased which indicates that the modification to the roughness coefficient has the opposite response to the flow path slope previously discussed. The parameter had no impact on runoff volumes.

Impervious Area Depression Storage

This depression storage parameter displayed a rapid linear decrease in runoff volume as a result of increasing the size of the storage. This is typical of a model response whereby more runoff is being detained resulting in less reaching the catchment outlet. No change in peak runoff or time to peak was noticed however correlation remained almost unchanged indicating that the modification to this parameter does not appreciably change the hydrograph shape.

Supplementary Area Additional Time

Modification to the supplementary area additional time resulted in changes to all hydrograph features. An increase in this value decreased total runoff and peak discharge and increased time to peak. Despite this a slight increase in correlation was apparent for reasons unknown.

Supplementary Area Flow Path Slope and Roughness

Modifications to both these parameters did not appreciably change the hydrograph features. A very slight increase in time to peak was noticeable with a decrease in slope and an increase in roughness, similar to the impervious parameters.

Supplementary Area Depression Storage

This storage parameter, as expected, resulted in a decrease in total runoff volume with an increase in value. The modification to this parameter did not affect other hydrograph features and correlation remained almost unchanged.

Supplementary Area Lag Time

The lag time factor decreased peak runoff and increased time to peak when it was increased. As this variable, results in modifications to the time component only, no change was evident in total runoff volumes.

Grassed Area Additional Time

The grassed area additional time parameter proved highly sensitive to modification. An increase in time produced a large decrease in total runoff and peak discharge. This is undoubtedly due to increased contact time on the pervious surface resulting in an increase in infiltration.

Grassed Area Flow Path Slope and Roughness

The response of this variable was similar to the effect on impervious area however the magnitude was more pronounced. Total runoff as well as peak discharge was increased with a decrease in surface roughness or an increase in grade. Conversely, time to peak decreased with an increase in surface roughness and a decrease in grade. This may be due to the cumulative effect of overlapping hydrographs.

Grassed Area Depression Storage

Depression storage in pervious area proved highly sensitive, with a rapid linear increase in total runoff and peak discharge with a decrease in storage size. The depression storage factor had no appreciable impact on time to peak, and exhibited a mild increase in correlation with an increase in storage size.

Pit Loss Coefficient

As expected the pit loss coefficient had no impact on any hydrograph parameters. This is likely to be due to the fact that there is only one pit in

the HSC catchment. The sensitivity of the model to a catchment such as HSA with over 15 pits may prove different.

Soil Type

Modifications to the soil type had negligible impact on the model for soil types 1 to 3. However the use of soil type 4 dramatically increased total runoff. As the soil type is already at 4 this may have no impact on the calibration process.

5.1.3 Calibration and Testing

The most important parameter to achieve is the mass balance of runoff volume. It would be ideal to also correlate time to peak, peak discharge and overall hydrograph shape however that is not the aim of this exercise as only total loads are required. The following parameters were found to affect the mass balance of flows in order of sensitivity:

1. Soil Type
2. Grassed Area Depression Storage
3. Grassed Area Additional Time
4. Grassed Area Flow Path Roughness
5. Paved Area Depression Storage
6. Grassed Area Flow Path Slope
7. Supplementary Area Additional Time
8. Supplementary Area Depression Storage

The model initially predicted a mass balance of 22.7m^3 for Storm 2 Burst 2, while the measured hydrograph indicated a mass balance of 26m^3 . Thus any modification to the variables needs to result in an associated increase in the

total runoff. While soil type was the most sensitive variable, any further change in the soil type would lead to a further decrease in total runoff (since the soil type is already at 4). A modification to the grassed area flow path roughness or slope will also result in a modification to the peak discharge as well as time to peak.

Since Grassed, paved and supplementary area depression storages has no appreciable impact on other features of the hydrograph, these variables will be modified to achieve mass balance of flows. New values include 2 mm, 0.5 mm and 0.5 mm respectively. In addition, the grassed area additional time has been reduced from 15 minutes to 10 minutes. A re-run of the model following the modification of these parameters produced a total runoff volume of 24.7 m³ suggesting an accuracy of 5%. Figure 5.2 displays the calibrated hydrograph for catchment HSC.

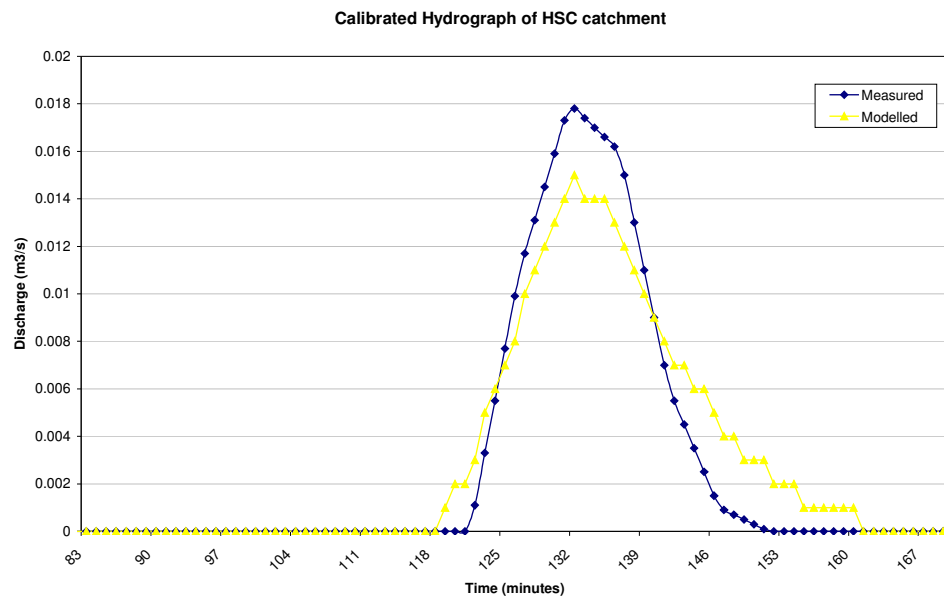


Figure 5.1 Measured and modelled hydrographs for HSC catchment following calibration

As a result of this process the HSA and HSB catchments were modified to match the new parameters accordingly with a few exceptions. The upper portion of catchment HSA in sub-catchment A has retained a grassed area additional time of 15 minutes due to the vegetated nature of the area. This differs markedly from the well-mown grassed area that was contributing to HSC. Grassed and supplementary area additional times have also retained high values where a prolonged flow path through backyards or across lawns indicates the necessity for a higher value. Unfortunately with now residential area contributing to HSC the effects of these parameters could not be tested.

5.1.4 Modelled Runoff and Pollutant Loads

In order to determine runoff volume the DRAINS model was run for the seven available monitored events described in the previous chapter with event specific rainfall. The discharges were collated and are presented below in Table 5.2.

Table 5.2 Measured TSS concentration combined with modelled runoff to calculate TSS load

	Inlet Pipe	TSS Concentration mg/L	Modelled Runoff m ³	TSS Load g
8/12/06	HSA	69	4.60	317.40
	HSB	76	1.30	98.80
	HSC	60	0.90	54.00
4/01/06	HSA	190	18.70	3553.00
	HSB	130	5.20	676.00
	HSC	220	3.60	792.00
12/02/06	HSA	170	22.30	3791.00
	HSB	50	6.10	305.00
	HSC	180	4.20	756.00
17/02/06	HSA	69	0.90	62.10
	HSB	76	0.20	15.20
	HSC	60	0.20	12.00
2/03/06	HSA	100	10.50	1050.00
	HSB	320	2.90	928.00
	HSC	140	2.00	280.00
5/03/06	HSA	31	7.10	220.10
	HSB	32	2.90	92.80
	HSC	15	2.00	30.00
23/03/06	HSA	190	14.30	2717.00
	HSB	260	2.70	702.00
	HSC	170	3.90	663.00

5.1 MASS BALANCE MODEL

5.2.1 Model Application

To apply Brodie's (2006) mass balance model it is necessary to determine the surface characteristics of the catchment and measured rainfall characteristics. As mentioned, the surface characteristics to be applied to the mass balance model differ to the application of the DRAINS model in that surface types are required as opposed to hydrological areas. As a result a different division of catchments and determination of surface types based on pollutant generation

characteristics is required. Obviously road areas have been modelled as road surfaces and roof areas have been modelled as roof surfaces. However grassed areas as indicated by Brodie (2006) have been also assumed as any pervious surface in the catchment including gardens, backyards, trees and median strips. In addition driveways and footpaths have been modelled as carpark area. No bare soil was observed in aerial photography or during site visits and as such has not been included in the mass balance model. The proportion of each surface type in the major catchments is shown below in Table 5.3.

Table 5.3 Surface characteristics of the Hoyland Street catchment for input into the mass balance model.

Catchment	Road	Carpark	Roof	Grass	TOTAL
	Area	Area	Area	Area	Total Area
	ha	ha	ha	ha	ha
HSA	0.536 51.86 %	0.029 2.83 %	0.029 2.76 %	0.440 42.55 %	1.03
HSB	0.147 64.60 %	0.001 0.49 %	0.005 2.27 %	0.074 32.64 %	0.23
HSC	0.102 34.35 %	0.045 15.24 %	0.000 0.00 %	0.150 50.41 %	0.30
HSO	0.000 0.00 %	0.009 3.61 %	0.000 0.00 %	0.249 96.39 %	0.26
TOTAL	0.79 43.23 %	0.08 4.68 %	0.03 1.86 %	0.91 50.24 %	1.82

In addition to determining the catchment surface characteristics a number of other variables specific to each individual rainfall event are required. These rainfall parameters include:

- Total antecedent rainfall (mm);
- Antecedent rainfall period (hours);
- Storm duration (hours);
- Rainfall depth (mm);
- Mean intensity (mm/hour);
- Peak 6 minute intensity (mm/hour);
- Interburst period (hours);
- Square of the peak 6 minute intensity ($\text{mm}^2/\text{hour}^2$);
- Sum of the square of all 6 minute intensities ($\text{mm}^2/\text{hour}^2$);

A fundamental basis of the mass balance model is that the main drivers for particle washoff are rain power and transport of particles along the lateral drain. Transport along the lateral drain is assumed a function of the peak intensity while particle washoff is assumed a function of the sum of the square of all intensities and the rainfall depth. The rainfall depth is also used to calculate the discharge from each surface. The total and time between antecedent rainfall is used to calculate the dry weather accumulation of particles on each surface. Wet weather accumulation is calculated from the length of the interburst period (the period between peak intensities in an individual storm). A hyetograph of each rainfall event has been graphed in order to determine each of these variables. The hyetograph for storm of 05/03/2006 has been reproduced in Figure 5.2 below, while the hyetographs for all storms are available in Appendix L.

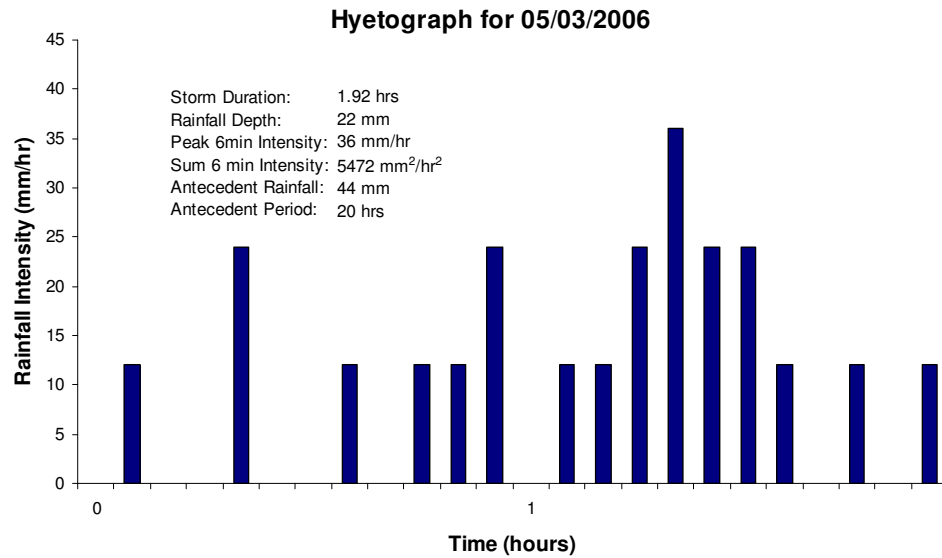


Figure 5.2 Hyetograph and mass balance model parameters for storm 05/03/2006

All accumulation, transport and washoff parameters that were calibrated by Brodie (2006) to Toowoomba conditions have been assumed as applicable to the Hoyland Street catchment. The simple linear approximation for grassed surfaces where the load is equal to 40 times the runoff depth has also been used. The use of these pre-calibrated parameters is expected to introduce inaccuracy into the EMC predictions of the mass balance model, however it is not the purpose of this dissertation to calibrate the mass balance model to the conditions of the Hoyland Street Catchment.

5.2.2 Modelled Event Mean Concentration and Runoff

Application of the mass balance model revealed that NCP EMCs as predicted by the mass balance model did not correlate well with the measured EMCs. The results also revealed that the surface discharge as predicted by the mass balance

model did not correlate well with the subcatchment discharge produced by the calibrated DRAINS model. Mass balance model spreadsheets for catchments HSA, HSB and HSC are available in Appendix M.

Figure 5.3 below illustrates the poor correlation between modelled and predicted EMCs. All data points are spread horizontally or vertically from the ideal dotted 1:1 line. This indicates that the mass balance model may be systematically over and underestimating the NCP EMC. For event with a small TSS EMC a large NCP EMC is produced, and vice versa. It appears that either the mass balance model is not estimating the EMC, data collection for the EMC was inadequate, or there is some bias in the measured EMC.

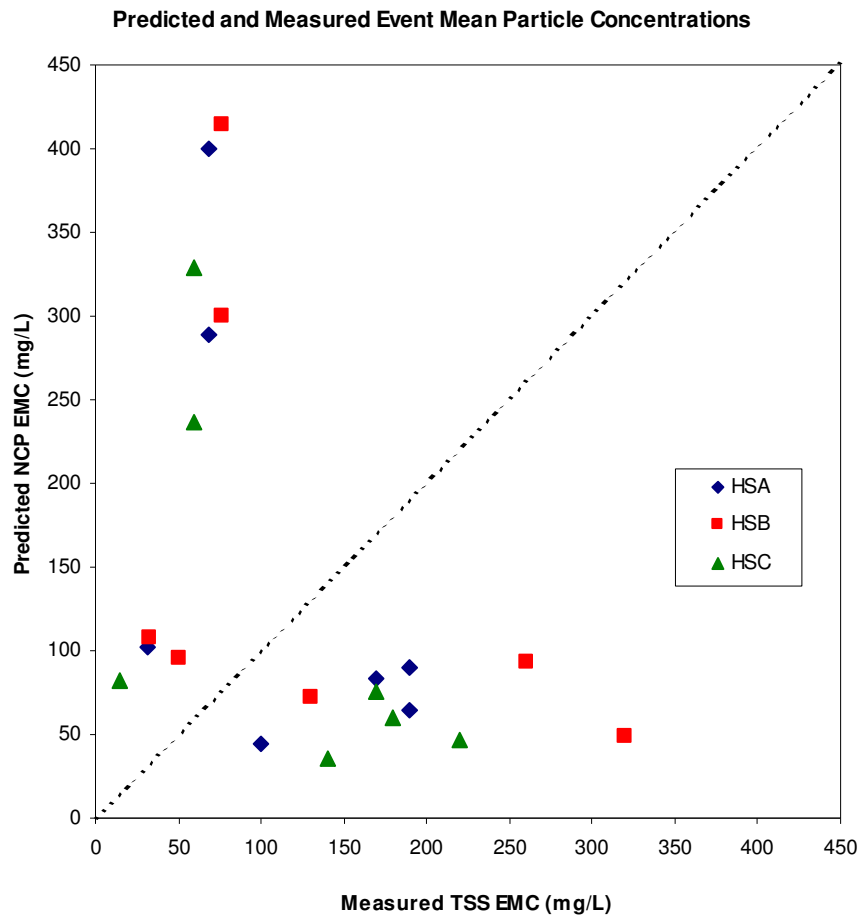


Figure 5.3 Event means for modelled NCP and measured TSS concentrations

In addition to the inaccurate prediction of particle concentrations, total event runoff volume was dramatically overestimated using the simple hydrologic calculations of the mass balance model. Figure 5.4 below illustrates this overestimation. This has resulted in a similar overestimation of event NCP load as can be seen in Figure 5.5. Given that the same rainfall data was used, the systematic overestimation of runoff when compared to the calibrated DRAINS model may represent an error in the assumptions underlying the mass balance hydrologic calculations or in the DRAINS modelling process.

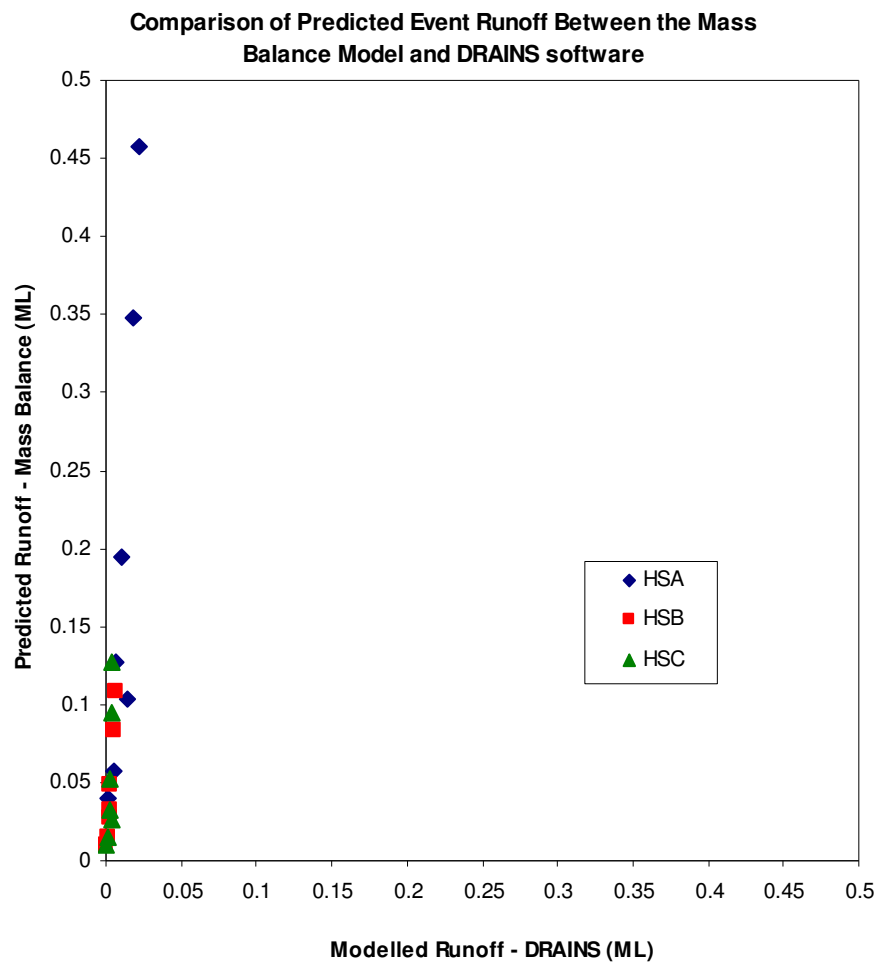


Figure 5.4 Event means for runoff volume produced by the mass balance model and the DRAINS software

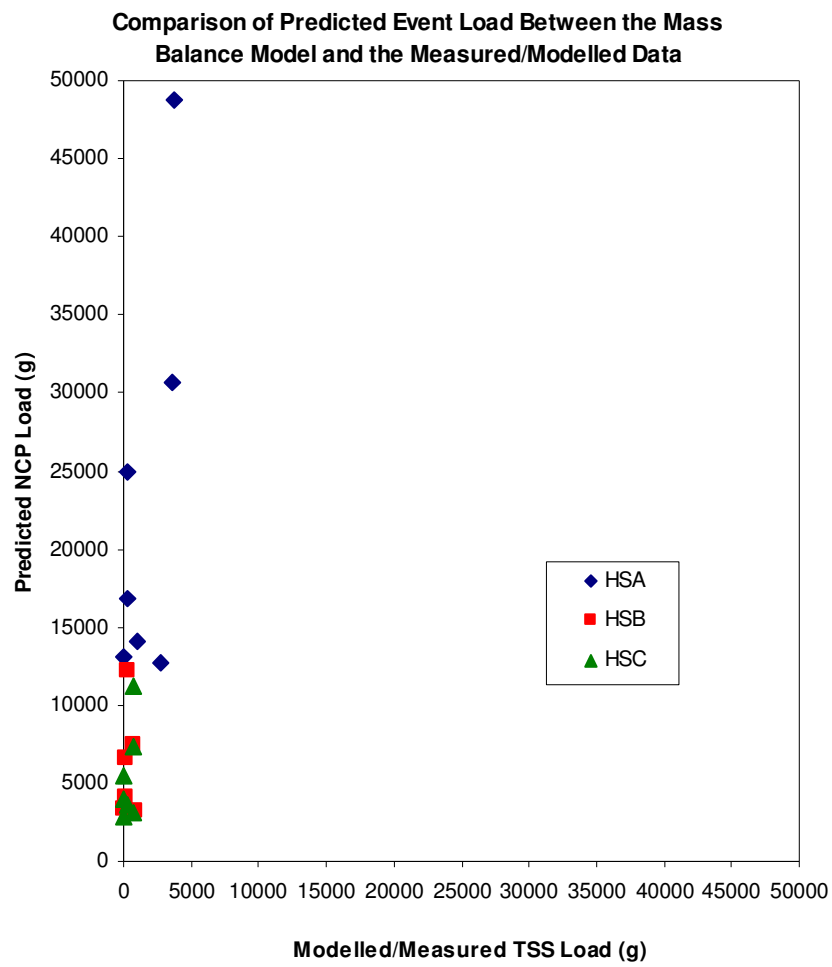


Figure 5.5 Event means for modelled NCP and measured/modelled TSS loads

5.2.3 Error Analysis

The mass balance model is calibrated using accumulation, transport and washoff parameters derived from experiments conducted by Brodie (2006) in Toowoomba. While it is not a part of this dissertation to calibrate the mass balance model, given the inherent data limitations and inaccuracies, some analysis of the sensitivity of these parameters is warranted. Figure 5.3 and 5.4 only provide a visual guide to the matching of modelled and measured data.

Calculated parameters from storm events also provide a source for error. The determination of the magnitude of this error may aid in understanding the source of error in the data. Given the fine scale determination of catchment surface characteristics and the likelihood of a direct linear correlation when modified, changes to surface areas were not included in the analysis.

The error analysis involved modification to the mass balance model parameters in order to determine the associated magnitude of error in predicted and measured EMCs. Each parameter was multiplied individually using a factor of 0.5 and 3 while other parameters were kept constant. Increasing and decreasing the size of the value aided in the determination of a direction of greatest sensitivity and whether the relationship was linear. As defined below, the two error statistics used are the cumulative EMC error (CEMCE) and the maximum EMC error (MEMCE).

$$\text{CEMCE} = \frac{\sum \text{EMC}_P - \sum \text{EMC}_M}{\sum \text{EMC}_M}$$

$$\text{MEMCE} = \text{Absolute Max } \frac{\text{EMC}_P - \text{EMC}_M}{\text{EMC}_M}$$

Where: EMC_P = Predicted event mean concentration

EMC_M = Measured event mean concentration

This error analysis was completed for each catchment and parameter individually. The detailed results from this analysis can be seen in Appendix N, however a summary is shown in Figure 5.6 below and indicates a large variation in the change in total error due to the application of the multiplication factors. Rainfall depth proved the most sensitive parameter and resulted in a median change in total error of 267% for cumulative error and 78% for maximum error. Maximum free particle load prior to the storm was the second most sensitive, with a median change in total error of 119% and 36% respectively. The highly sensitive nature of these parameters could account for some of the error present in the results. The parameters tested are listed below in order of the cumulative EMC error displayed (insensitive parameters excluded):

1. Rainfall Depth;
2. Maximum Free Particle Load Prior to Storm;
3. Particle Washoff Multiplier for Grassed Areas;
4. Maximum Detained Particle Load Prior to Storm;
5. Dry Weather Accumulation Rate of Detained Particles Before the Storm;
6. Antecedent Rainfall Period;

7. Interburst Period;
8. Wet Weather Accumulation Rate of Free Particles During the Storm;
9. Peak 6 min Intensity;
10. Sum of the Square of all 6 min Intensities;
11. Storm Duration;
12. Maximum Free Particle Load During Storm;
13. Dry Weather Loss Rate of Retained Particles in the Drain Before the Storm;

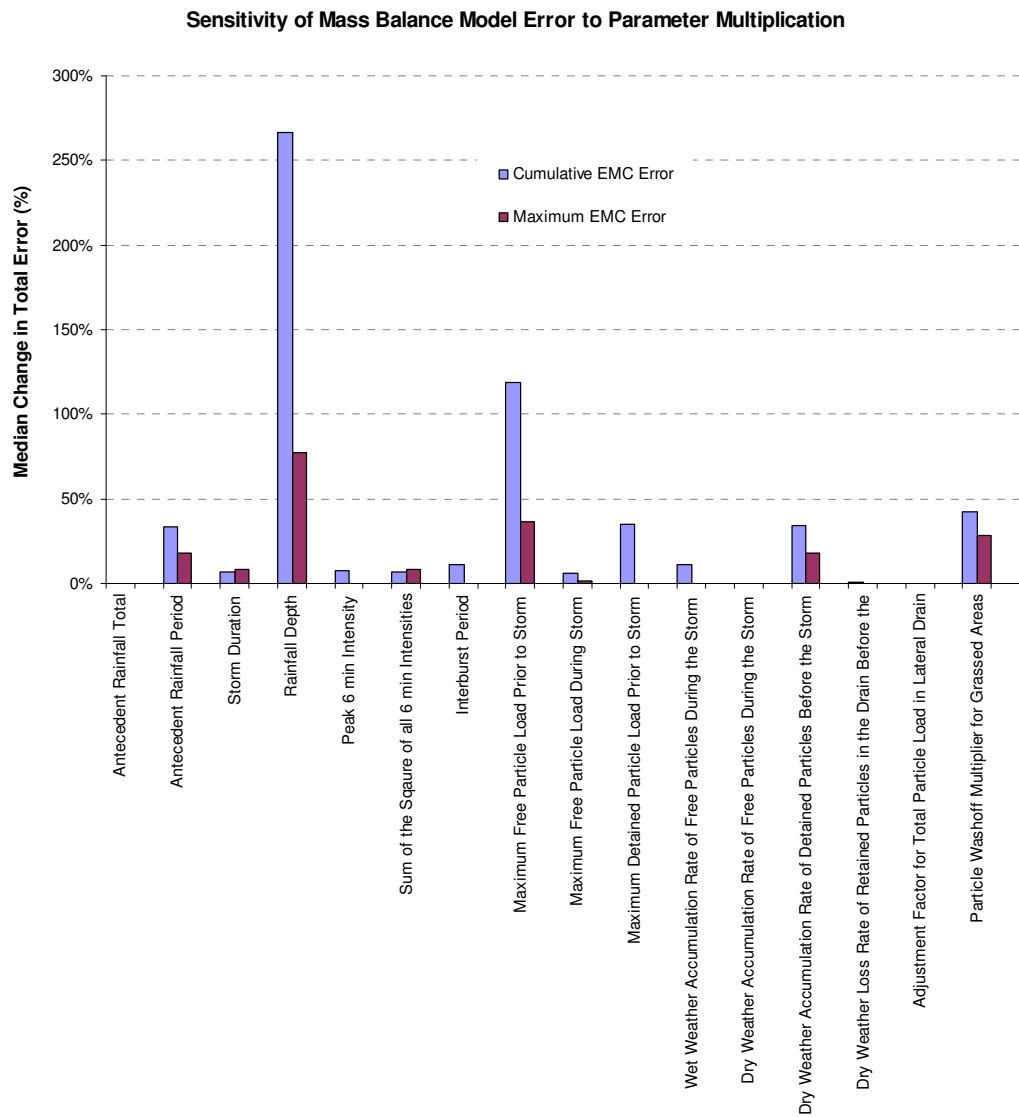


Figure 5.6 Results from error analysis of mass balance model

CHAPTER 6

DISCUSSION

6.0 DISCUSSION

6.1 DISCUSSION

Major sources of error in available data have resulted in a very poor prediction of event mean concentrations of non-coarse particles when compared to measured data. This, combined with a poor prediction of runoff has led to an inability to predict event loads of non-coarse particles. As a result an error analysis was undertaken which revealed that the major cause of the error might have been due to inaccurate rainfall depth and unavailable data for the maximum free particle load prior the storm event.

The difficulty in obtaining accurate and relevant rainfall data proved a time-intensive exercise that limited available events and may have introduced error into the models. The pluviometric gauge used to obtain rainfall data was located 1.7 km from the catchment itself. Given the small size of the catchment it was likely in many instances that runoff was not occurring in the catchment while rainfall was registering at the gauge. The distinct lack of rainfall events that correlated well with recorded runoff confirmed this assumption.

Measured event mean concentrations of suspended solids were determined using the TSS method, while the mass balance model predicted suspended solid concentrations in non-coarse particles using the SSC method. Literature suggests that the SSC method provides a more accurate measurement of suspended solid concentrations and the two can differ markedly in some instances. Initial error was anticipated as a result of the contrast between techniques, however the extent of the error will not be known.

The availability of measured runoff volumes also proved unreliable due to a poorly functioning monitoring system. This necessitated the need for rainfall runoff-modelling using DRAINS hydraulic and hydrologic software. Due the inconsistencies with the measured runoff hydrographs in HSA and HSB, catchment HSC was used to calibrate the DRAINS model. Due to poor correlation with rainfall only two events were available for calibration with the DRAINS model.

The use of the parameters calibrated from HSC during a single storm event may have also introduced significant error into the runoff volume calculations. HSC catchment differs markedly from other catchments in that the entire catchment drains to a single pit and into the bioretention system. HSC catchment also consists mostly of road surface of constant grade and no roof area. The pervious area in the catchment drains directly to the road with a small time of concentration. This is in contrast HSA and HSB, where there is a significant lag time between impervious and pervious area runoff to the point that two separate peaks are often evident.

Selection of surface types for input into the mass balance model may have also introduced error. As there are no surfaces in the mass balance model to describe footpaths or densely vegetated areas such as gardens and trees, carpark surfaces and grassed areas were selected. In the Hoyland Street catchment these surfaces are also indirectly connected to the drainage system flowing over grassed lawns or along property boundaries before discharging into the stormwater system. As a result the mass balance model is likely to have overestimated the pollutant generation from these surfaces. A number of ICIA surfaces represented as garden sheds and roofs draining into ‘bubbler style’ outlets onto grassed surfaces were represented as roofed area by the mass balance model. This may have also been a

source or error in the runoff volume calculation as the mass balance model represents these surfaces as draining directly into the stormwater system.

6.2 FURTHER APPLICATION OF MASS BALANCE MODEL

As shown in Chapter 6, the most sensitive parameter to error in predicting the EMC from each catchment using the mass balance model proved to be the rainfall depth. In addition, the antecedent rainfall period proved somewhat sensitive with a mean cumulative error of 34% given a 50% movement in value. This highlights the need for an accurate rainfall data series when using the mass balance model to predict event mean concentration.

While Brodie (2006) confirmed the accuracy of the mass balance model in comparison to other new and well-established techniques, the practical limitations of data availability to limit model error may prevent more widespread use. While established networks of rainfall gauges are prevalent in many major cities, they are seldom within 2km of each other, generally due to the historical need for calibration of models in much larger catchments. The time-intensive screening process required to 'weed out' events that do not correlate with runoff may prove a barrier, given the widespread industry and local government acceptance of far more easily applied yet inaccurate models (eg. MUSIC). Many of these models have been calibrated and are recommended for larger catchments. As the mass balance model using surface specific parameters has the potential to provide ease of use and greater accuracy at the smaller scale, these barriers to use may be especially applicable.

The particle accumulation and washoff parameters may also represent an additional obstruction to widespread use due to the uncertainty in applying the parameters to different environmental conditions in different catchments. The error analysis confirmed that the maximum free particle load prior to the storm proved highly sensitive to error with a 50% increase in the variable producing a 117% increase in mean cumulative error. Other variables proved somewhat sensitive including the maximum detained particle load prior to the storm, the accumulation rate of detained particles prior to the storm, and the particle washoff multiplier for grassed areas. A mean cumulative error of 35%, 34% and 43% respectively was observed. Given the number of these variables and their sensitivity, calibration for potential users would prove difficult. In addition, no guidance is provided in the selection of these parameters. Testing of the mass balance model on additional catchments, and calibrating to determine appropriate values for parameters would prove wholly beneficial.

The accurate calculation of event mean and annual loads can prove an indispensable planning, design and maintenance tool. The runoff volume determination through hydrologic computations is limited to simple linear functions in the mass balance model. While this may be accurate for impervious surfaces, the relationship between rainfall and runoff in pervious surfaces is generally more complex. However, the application of the mass balance model to smaller mostly impervious catchments may invalidate this assumption. Integration with existing hydrologic software may provide ease of use and greater accuracy. The prevalence of already established hydraulic runoff routing models for urban stormwater planning in major cities may also provide runoff volumes. A number of design storms could be run through the hydraulic model to determine a curve of rainfall versus runoff volume, and provide a relationship directly based on rainfall. The hydrologic calculations performed by the model could be expanded to improve runoff volume prediction and as a result load prediction, however this could be balanced against the necessity to do so.

A wide range of surface types and configurations are present in urban and suburban areas, each with differing pollutant accumulation and washoff characteristics. The choice of surface type in the mass balance model is limited to five major surface types each of which is assumed to drain directly to the stormwater system. The reality of urban surfaces may include footpaths, brick and porous paving, vegetated surfaces and tiled roofs which may drain onto a combination of each other before reaching the stormwater system. Additional monitoring of different surface types and configurations could improve the accuracy and applicability of the mass balance model, however increasing complexity for limited increases in accuracy may invalidate the need.

The mass balance model has been proven to more accurately predict event mean concentrations and loads than a number of models available and in general use

(Brodie, 2006). The inability to correlate the measured and predicted NCP concentrations and loads is likely to be a result of inaccurate data and the unavailability of particle washoff and accumulation parameters specific to the Hoyland Street catchment. An error analysis undertaken confirms that these are the largest sources of cumulative and maximum error. However as a result of the application, insight has been gathered into the application of the mass balance model and potential areas for improvement.

CHAPTER 7

CONCLUSION AND RECOMMENDATIONS

6.0 CONCLUSION AND RECOMMENDATIONS

6.1 CONCLUSION

A major component of this dissertation was comparing the measured EMC with the predictions of the mass balance model to determine the efficacy of the model in calculating suspended solid concentrations and loads. This was achieved, however the result was a very poor correlation between data sets. This was primarily due to limitations with measured rainfall and runoff data. The use of particle accumulation and washoff parameters calibrated to Toowoomba may have also increased the error. As a result a major exercise of this dissertation was to:

- Limit already inherent error in the data by minimising any introduced error;
- Conduct an error analysis to determine likely causes of error;
- Provide feedback on the application and limitations of the mass balance model.

While the ability of the mass balance model to predict suspended solid concentrations and loads could not be fully tested, several limitations to the widespread use of the mass balance model were discussed. These limitations are focussed on the ease of application and calibration of the model. The major limitation to further use of the model to determine suspended solid EMCs involves the lack of available rainfall data that correlates well with runoff. Incorrect screening of rainfall events can introduce large cumulative errors into the model results.

6.2 RECOMMENDATIONS

Recommendations pertaining to the future monitoring of WSUD devices:

- Pluviometric gauges should be located within the catchment or as close as possible to the centroid of the catchment.
- Consideration should be given to the monitoring of WSUD devices in the design phase to limit the potential for inaccurate measurements.

Recommendations pertaining to the future use of the mass balance model:

- Surface types for input into the mass balance model should be selected carefully and take into account the nature of the surface type and its runoff configuration. Further monitoring of NCP EMCs on different surface types and configurations may be warranted.
- For the calculation of event specific loads and concentrations, rainfall gauges should be as close as possible to the centroid of the catchment to be modelled. If this is not possible a detailed analysis of the correlation between rainfall and runoff should be undertaken.
- Consideration should be given to the results of the error analysis, particularly the ability of inaccurate rainfall measurements to introduce error.
- The sensitive nature of the pollutant accumulation and washoff characteristics should be noted. Further monitoring may prove necessary to determine the variance of these parameters with different environmental conditions.
- Guidance should be provided on the selection of appropriate particle accumulation and washoff parameters for different urban surfaces.

- For complex catchments considerable pervious area, consideration should be given for the use of hydrologic/hydraulic software models for the determination of event runoff volume.

CHAPTER 8

BIBLIOGRAPHY

8.0 BIBLIOGRAPHY

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APPENDICES

FACULTY OF ENGINEERING AND SURVEYING

ENG4111/4112 Research Project **PROJECT SPECIFICATION**

FOR: KIEREN DAVIS

TOPIC: Application of an Urban Runoff Model using Surface Specific Parameters

SUPERVISOR: Ian Brodie

PROJECT AIM: Apply a mass balance urban runoff model to a specific catchment/s based on surface type and compare and evaluate its effectiveness in predicting the loads of suspended solids.

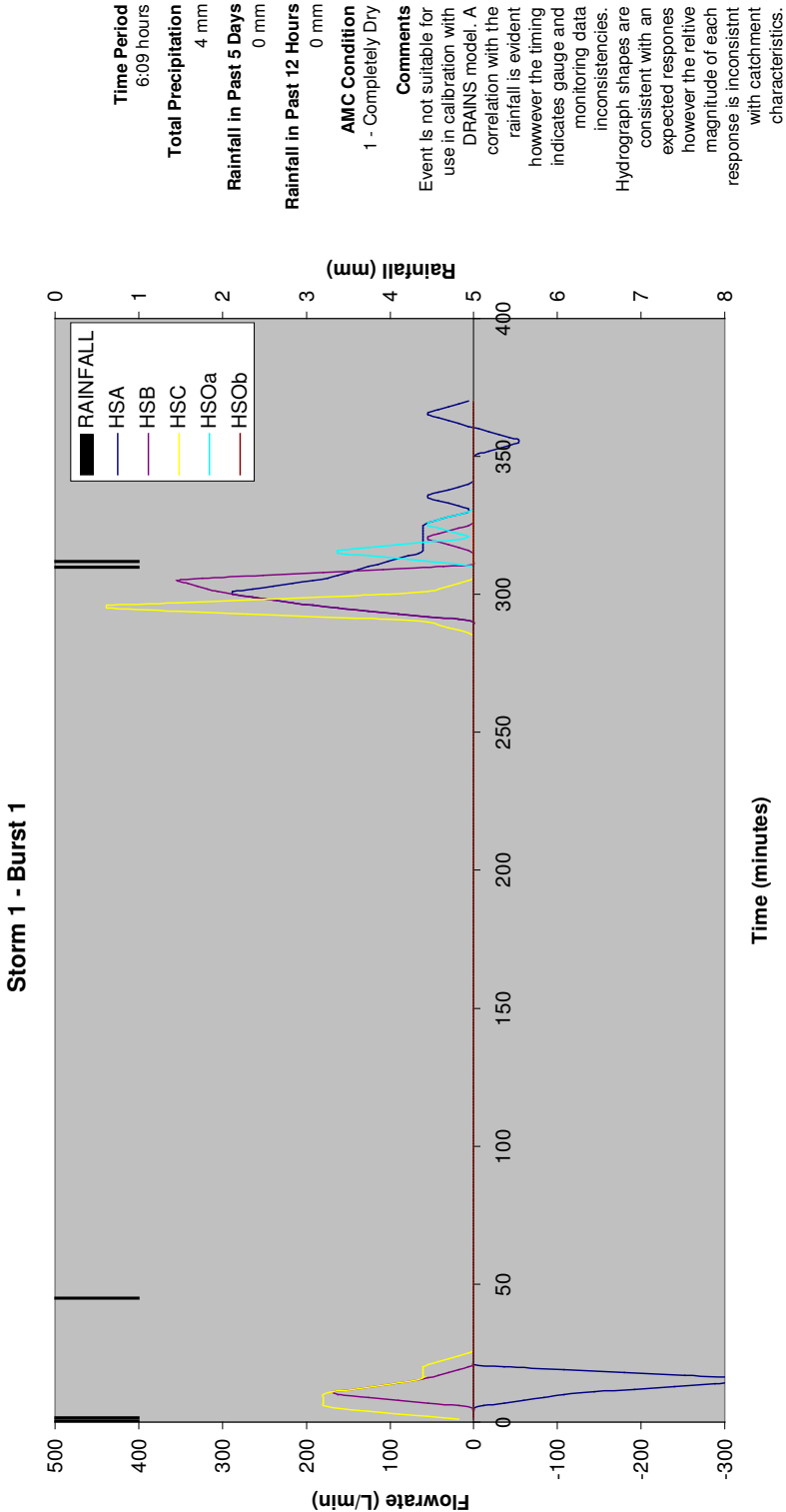
PROGRAMME: (Issue B, 22 March 2007)

1. Conduct a literature review on the prediction of suspended solids in stormwater using urban surface data.
2. Gather rainfall and water quality data relating to a specific catchment/s.
3. Determine the surface characteristics of the catchment.
4. Customise and test the Mass Balance Model spreadsheet.
5. Calibrate the model for a number of design storms
6. Compare and verify the model results with the real-life specific catchment data.
7. Write and submit a dissertation on the application and effectiveness of the model in predicting the quality of urban stormwater runoff.

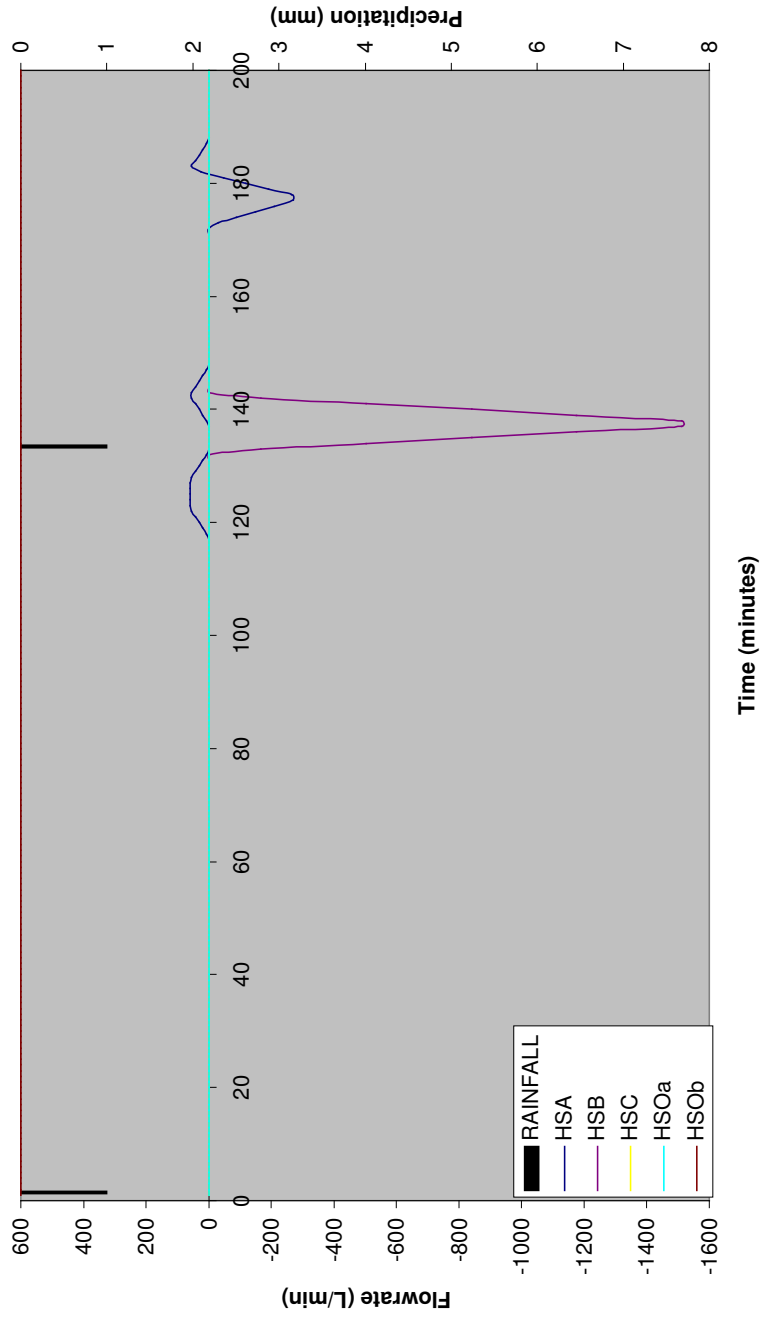
AGREED _____ (student) _____ (supervisor)
 Date: / / 2007 Date: / / 2007

Co-examiner: _____

APPENDIX B: Rainfall Runoff Analysis Graphs



Storm 2 - Burst 1



Time Period
3:06 Hours

Total Precipitation
2 mm

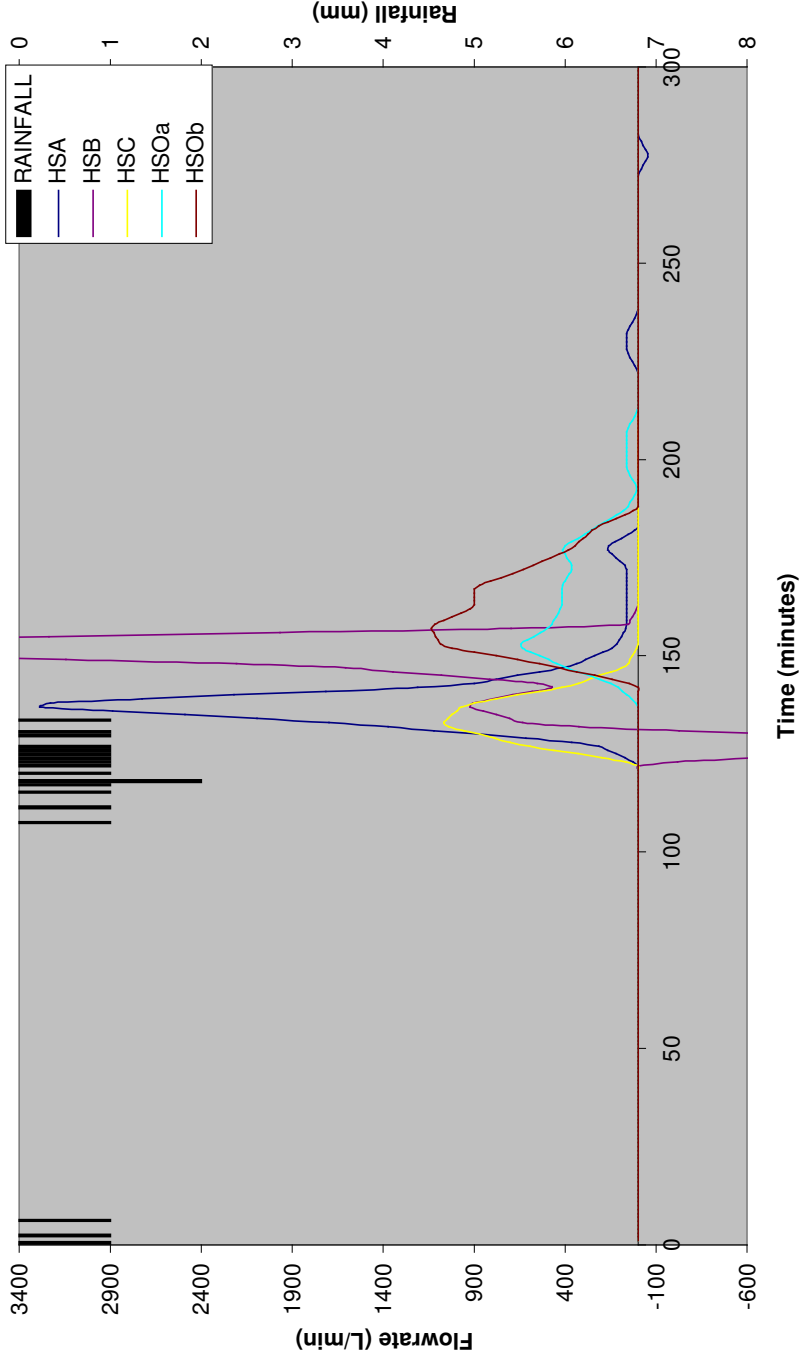
Rainfall in Past 5 Days
0 mm

Rainfall in Past 12 Hours
0 mm

AMC Condition
1 - Completely Dry

Comments
Event is not suitable for use in calibration of DRAINS model. This is due to the low runoff response which that would be expected from a dry catchment. The data also features negative flowrates for HSB and HSA.

Storm 2 - Burst 2



Time Period
4:42 hours

Total Precipitation
19 mm

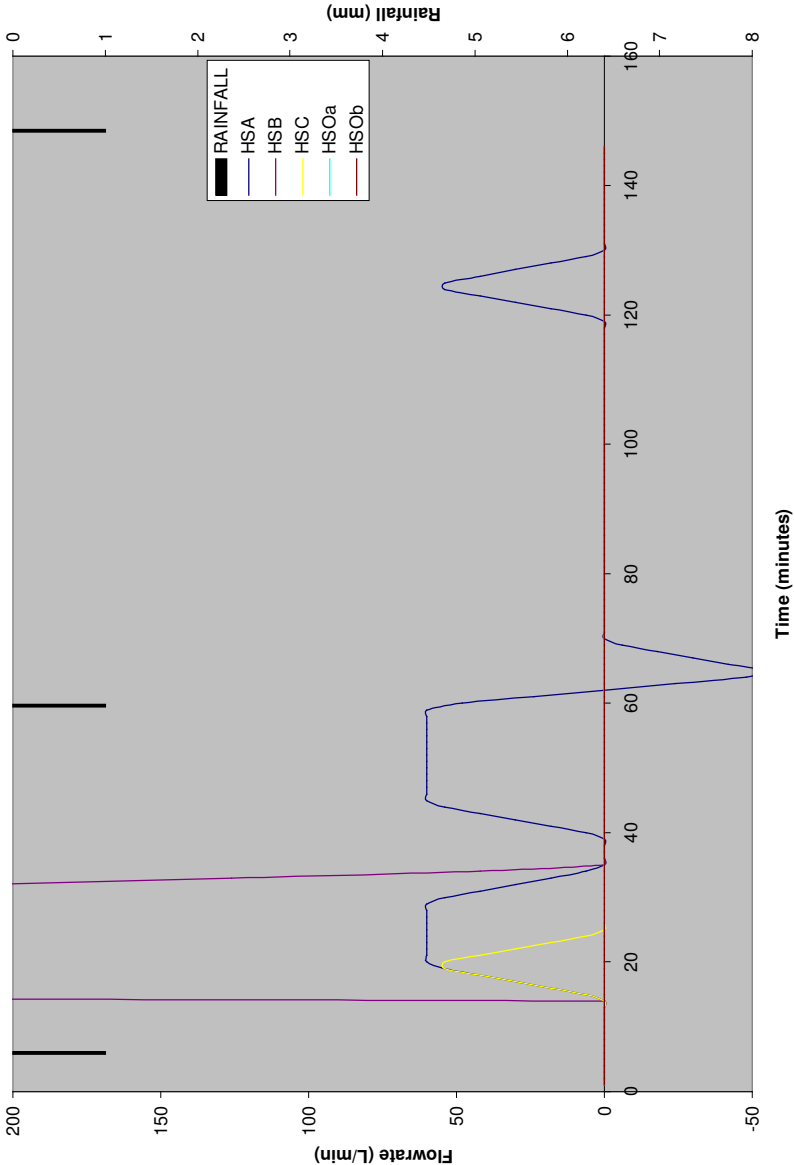
Rainfall in Past 5 Days
2 mm

Rainfall in Past 12 Hours
0 mm

AMC Condition
2 - Rather Dry

Comments
Event is suitable for use in calibration with DRAINS model. A high correlation between rainfall and runoff is evident. Suitable hydrographs were recorded for HSC, HSOa and HSOB. The data also features negative flowrates for HSA and HSB.

Storm 2 - Burst 3



Time Period
2:03 hours

Total Precipitation
3 mm

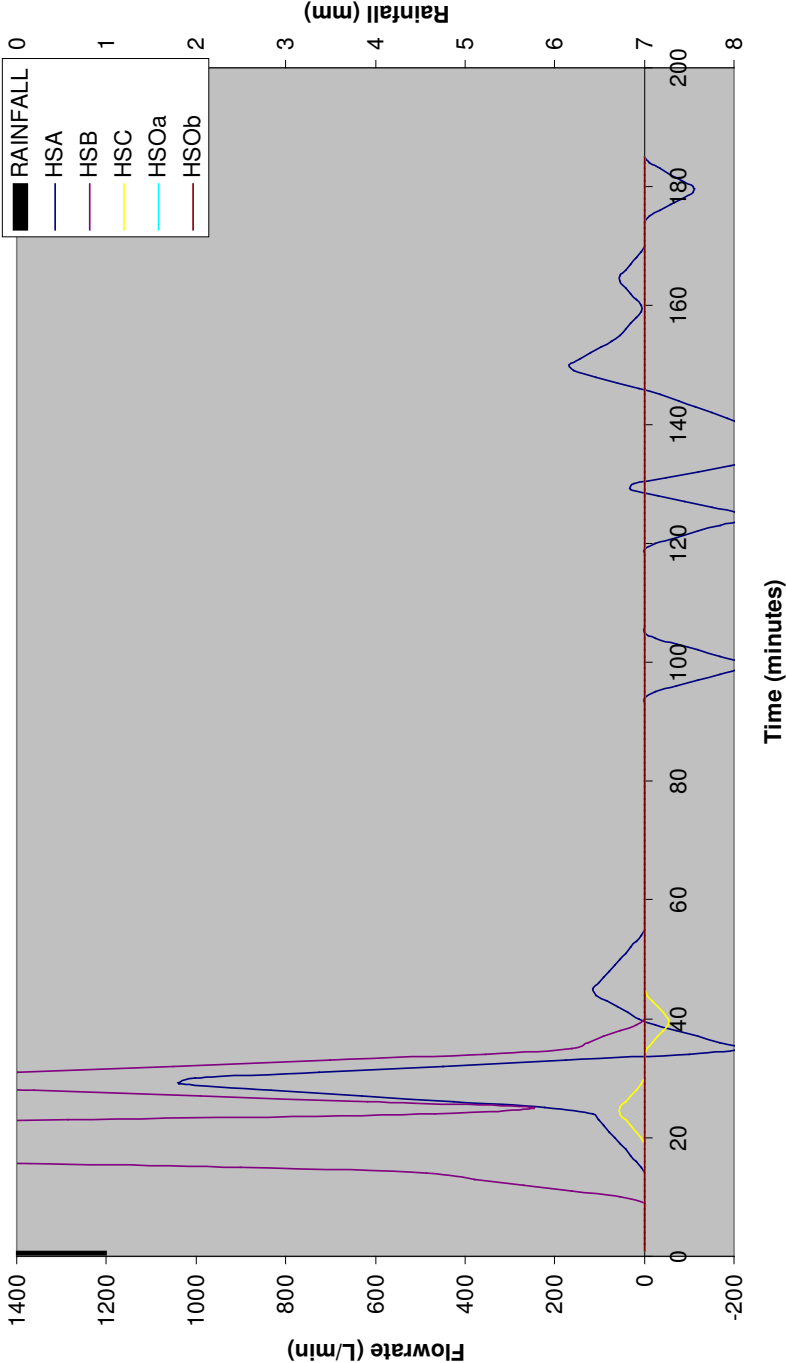
Rainfall in Past 5
21 mm

Rainfall in Past 12
19 mm

AMC Condition
3 - Rather Wet

Comments
Event is not suitable for use in calibration with DRAINS model. Correlation between rainfall and runoff is evident however not consistent.

Storm 2 - Burst 4



Time Period
3:04 hours

Total Precipitation
1 mm

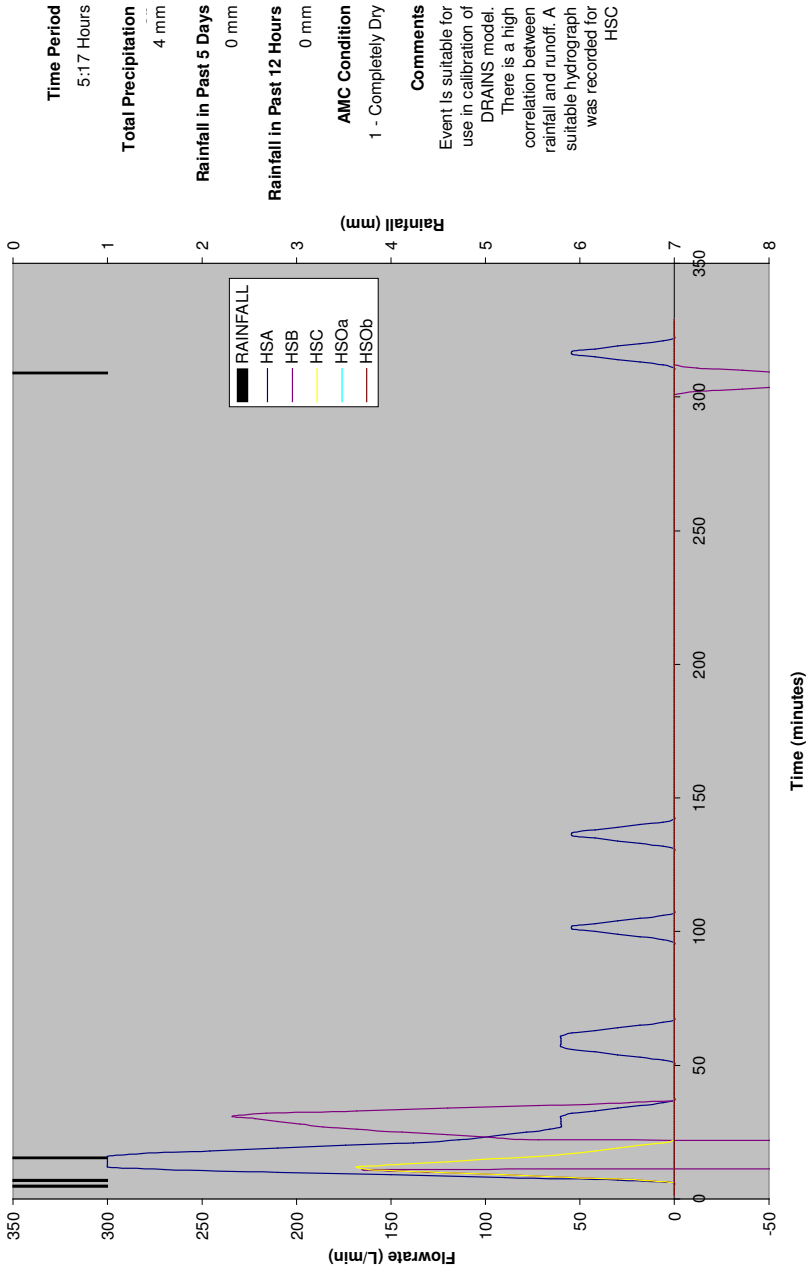
Rainfall in Past 5 Days
25 mm

Rainfall in Past 12 Hours
3 mm

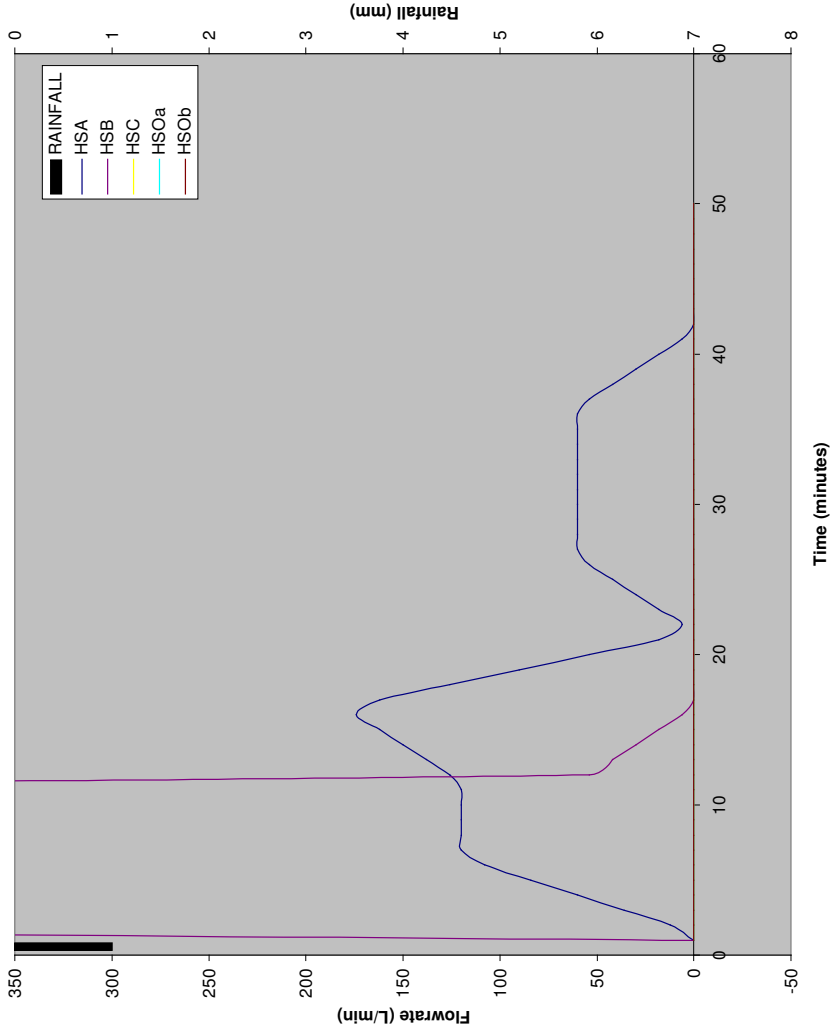
AMC Condition
4 - Saturated

Comments
Event is not suitable for use in calibration with DRAINS model. A high rainfall-runoff correlation is apparent, however widely exaggerated hydrographs and negative flowrates are also features of this data

Storm 3 - Burst 1



Storm 4 - Burst 1



Time Period
0:40 hours

Total Precipitation
1 mm

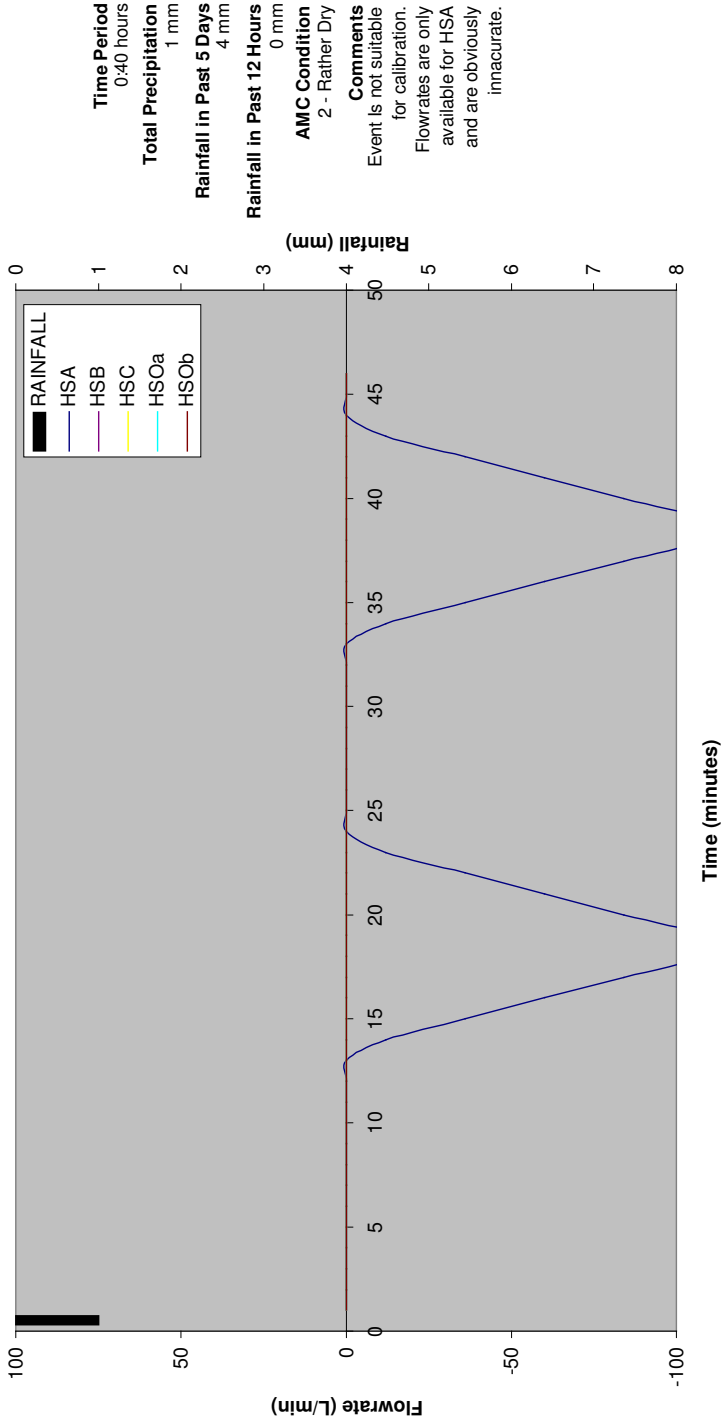
Rainfall in Past 5
4 mm

Rainfall in Past 12
0 mm

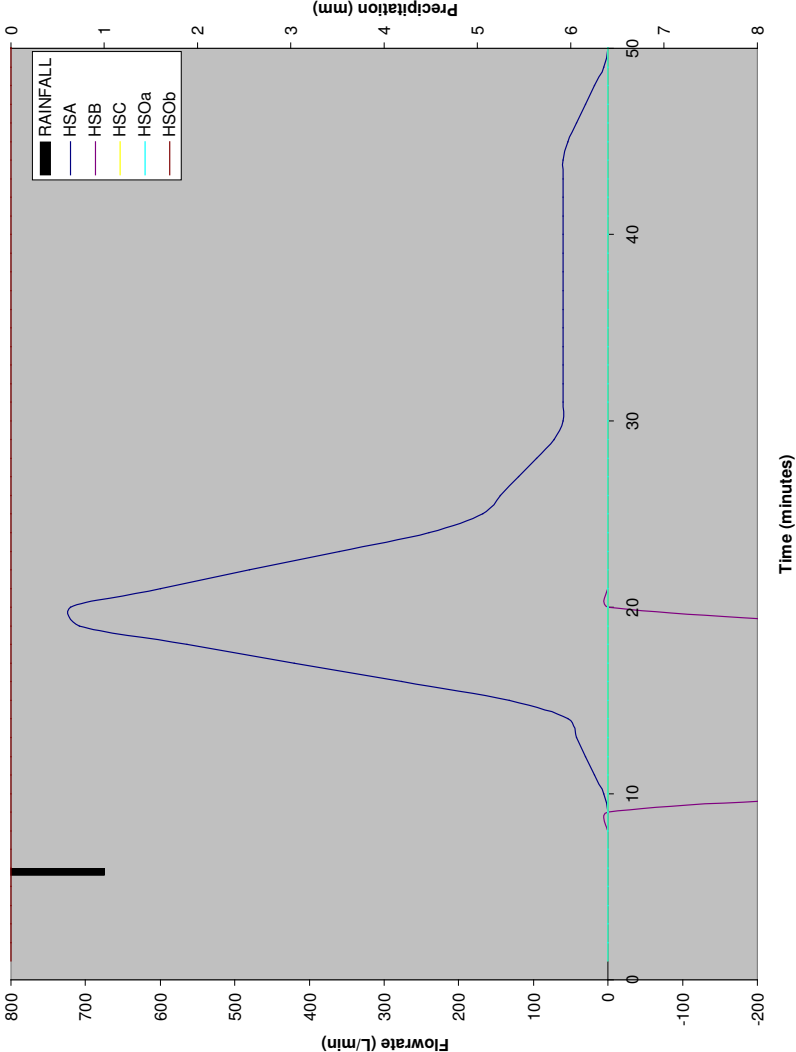
AMC Condition
2 - Rather Dry

Comments
Event is not suitable for use in calibration with DRAINS model however can provide insight into the time of concentration for low flows. Although HSB is exaggerated, HSA appears relatively accurate and exhibits a high degree of correlation with the rainfall.

Storm 4 - Burst 2



Storm 5 - Burst 1



Time Period
0:44 Hours

Total Precipitation
1 mm

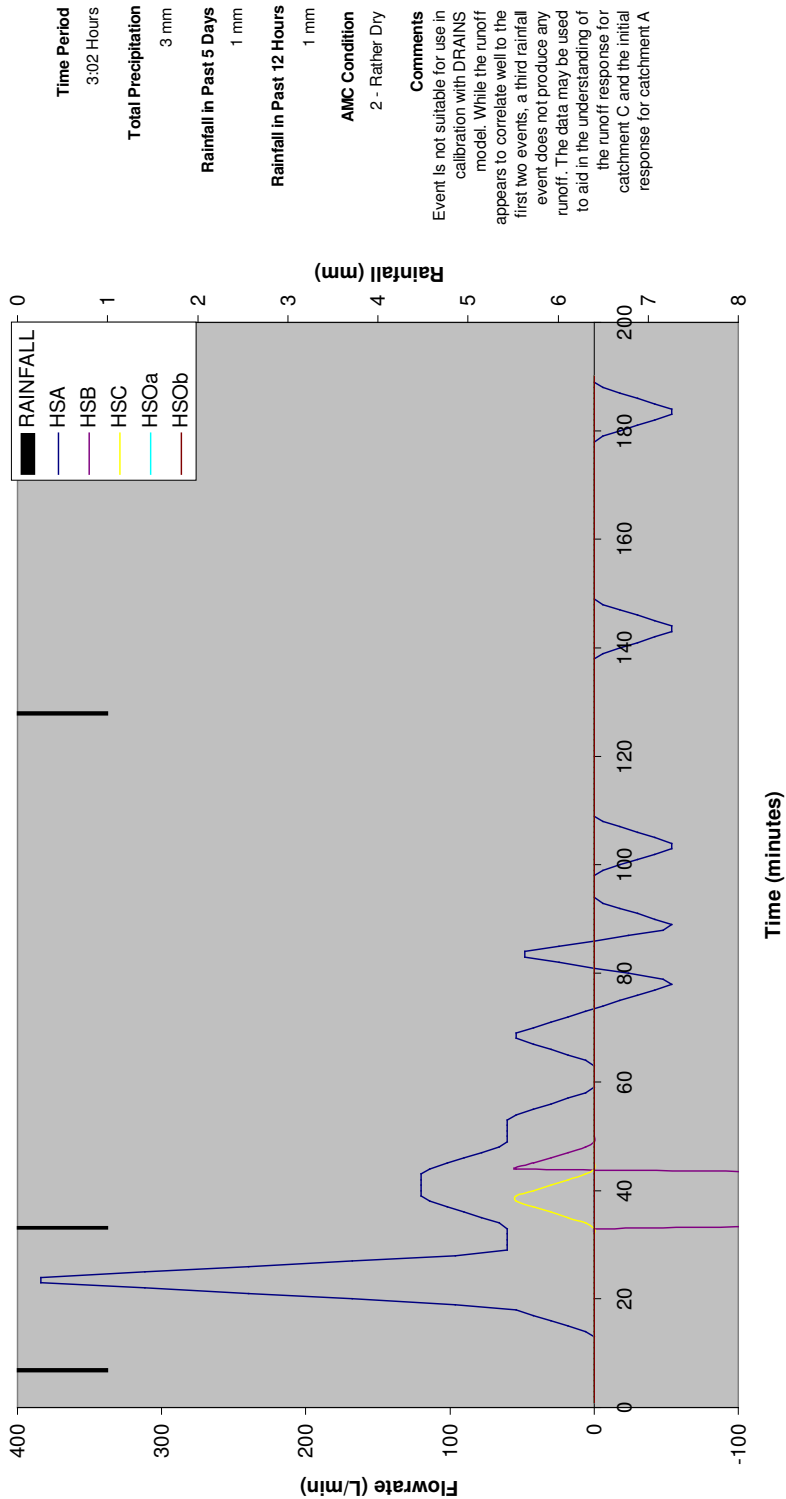
Rainfall in Past 5 Days
0 mm

Rainfall in Past 12 Hours
0 mm

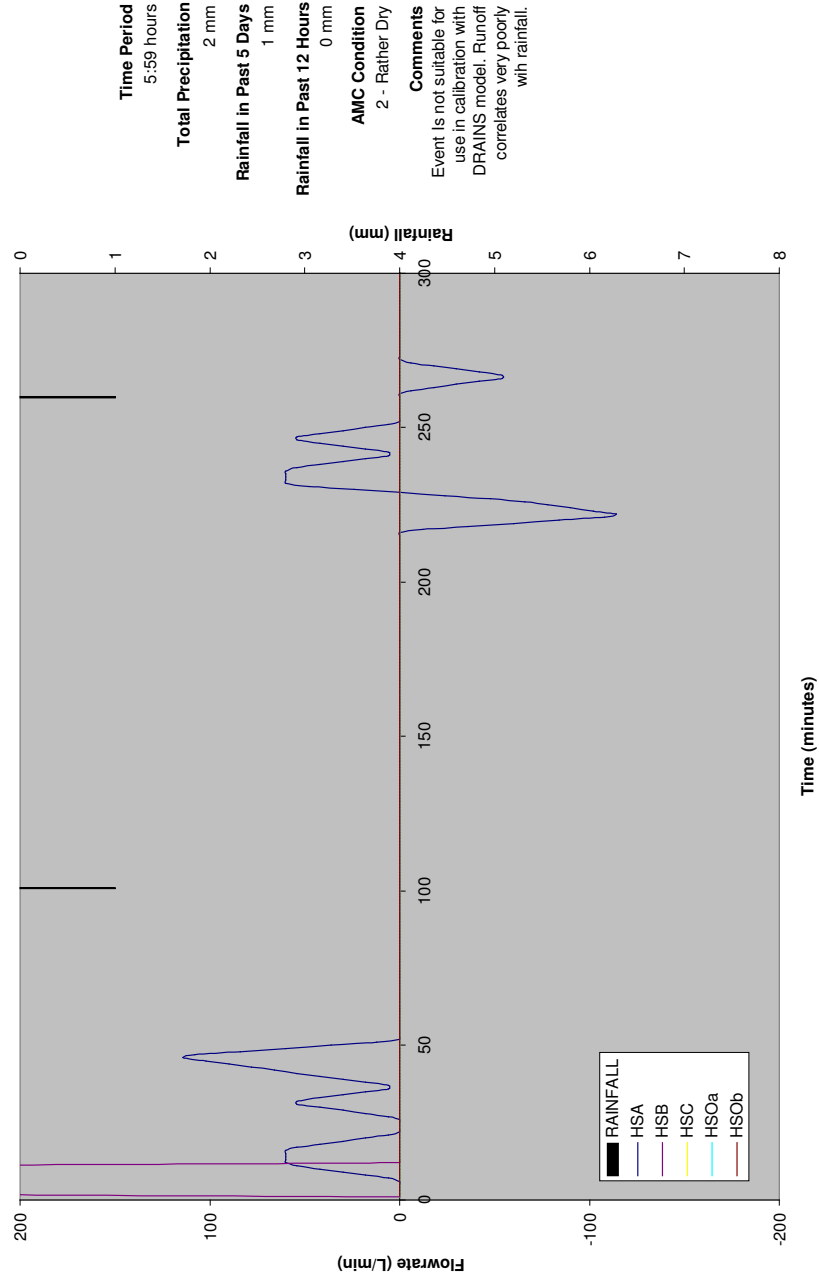
AMC Condition
1 - Completely Dry

Comments
Event is not suitable for use in calibration with DRAINS model however can provide insight into the time of concentration for low flows. Although HSB is exaggerated, HSA appears relatively accurate and exhibits a high degree of correlation with the rainfall.

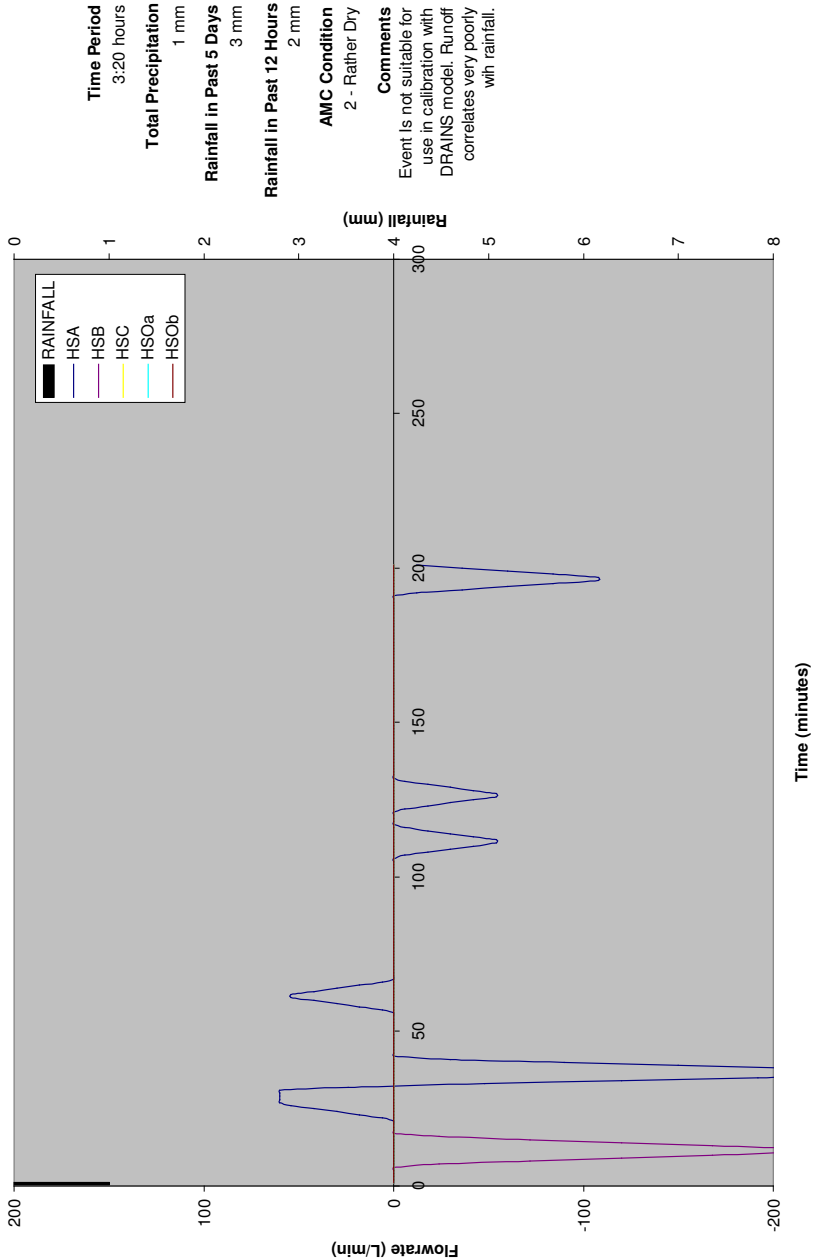
Storm 5 - Burst 2



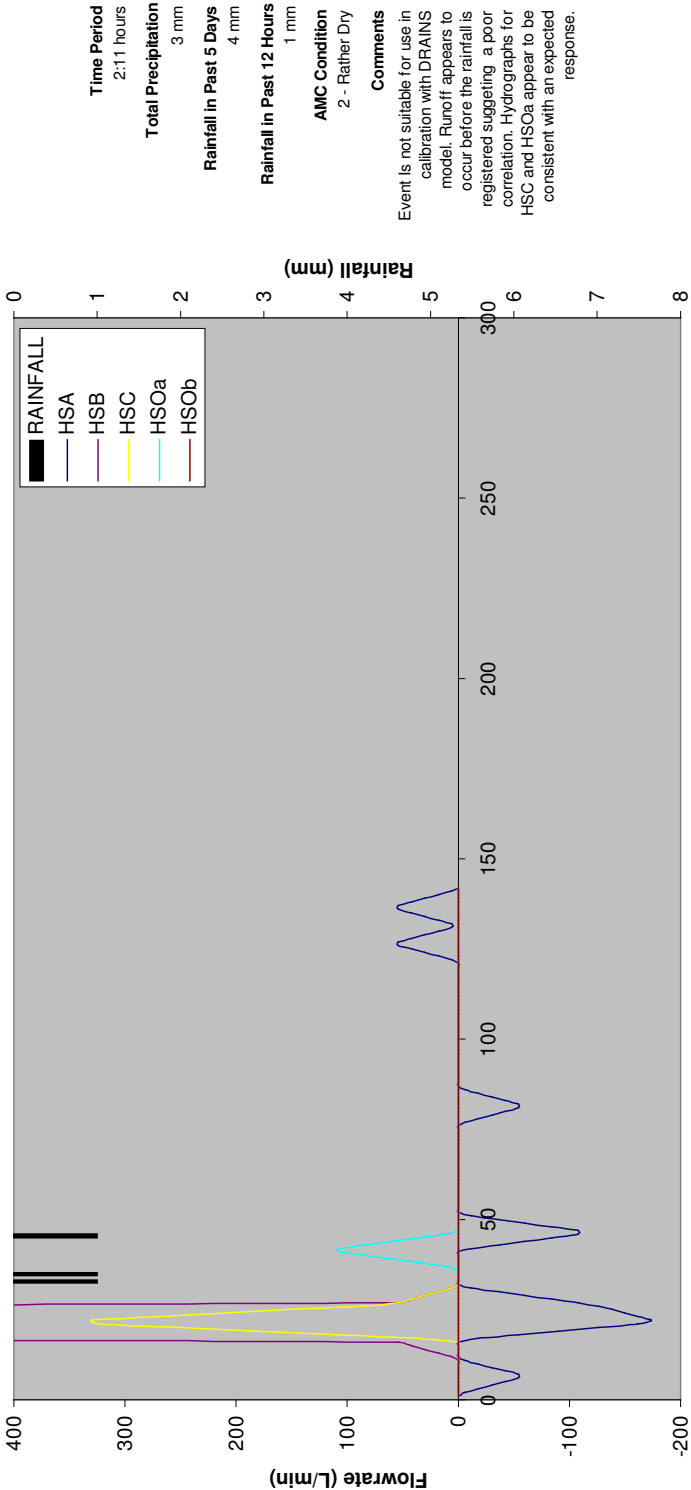
Storm 6 - Burst 1



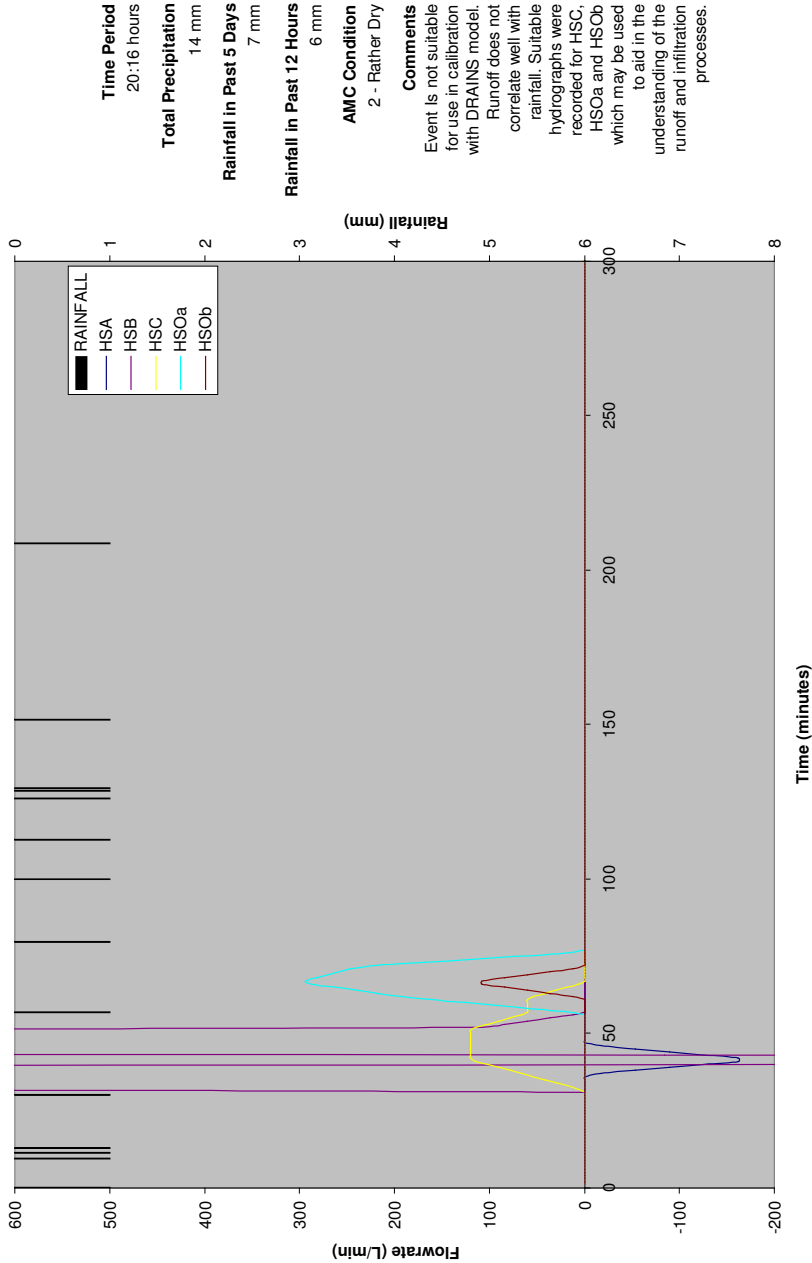
Storm 6 - Burst 2



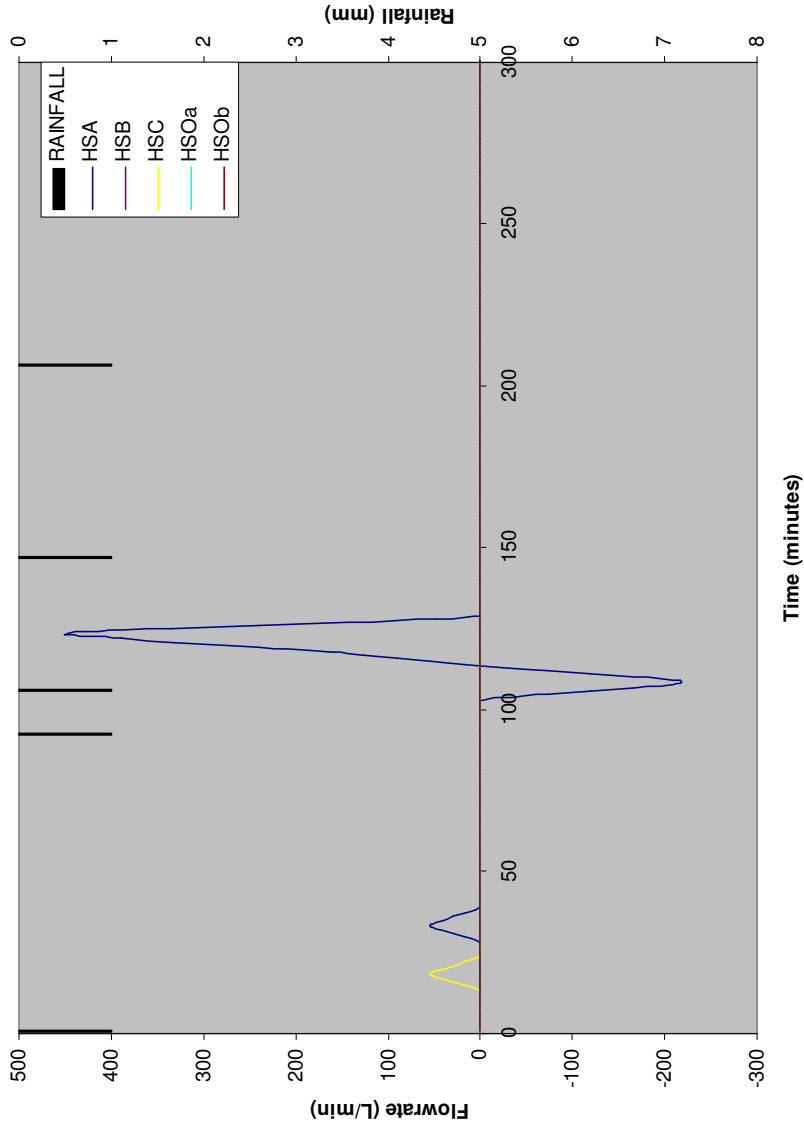
Storm 6 - Burst 3



Storm 6 - Burst 4



Storm 7 - Burst 1



Time Period
19:13 hours

Total Precipitation
5 mm

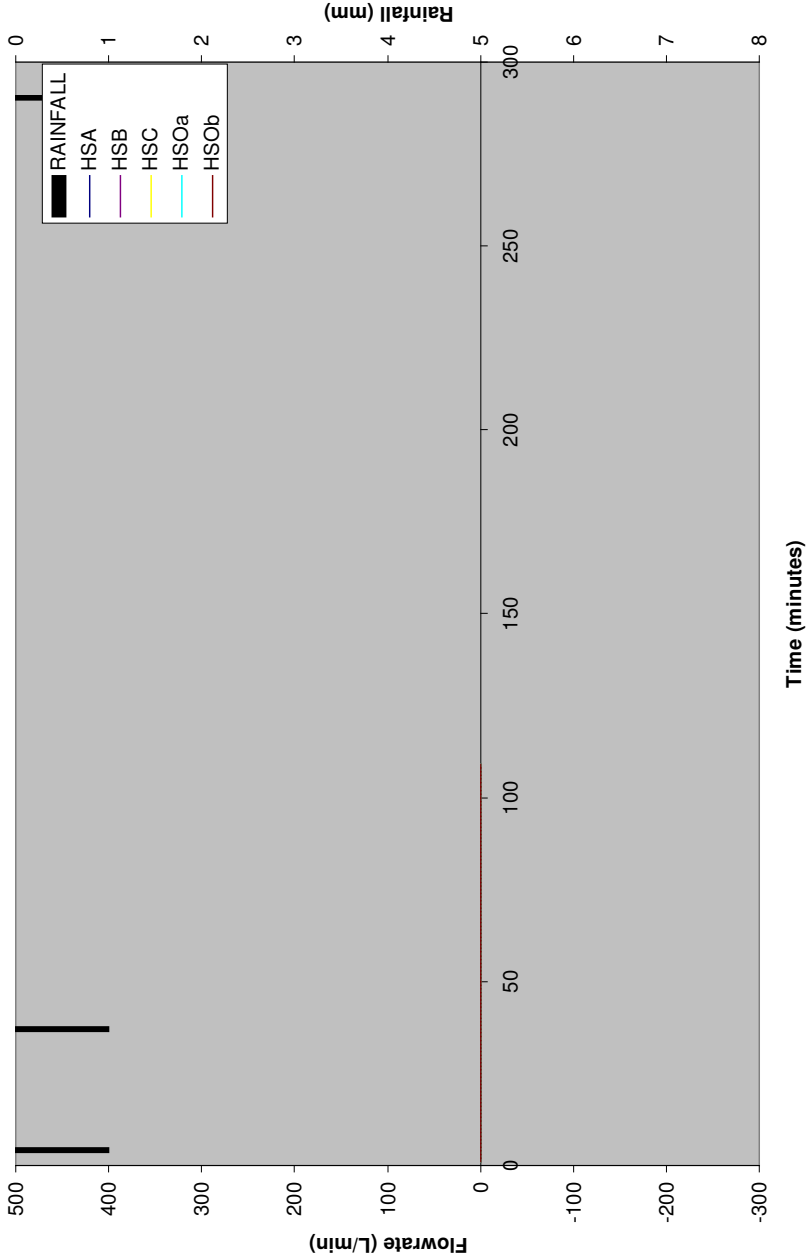
Rainfall in Past 5 Days
1 mm

Rainfall in Past 12 Hours
0 mm

AMC Condition
2 - Rather Dry

Comments
Event is not suitable for use in calibration with DRAINS model. A very poor correlation between rainfall and runoff is evident.

Storm 8 - Burst 1



Time Period
22:36 hours

Total Precipitation
3 mm

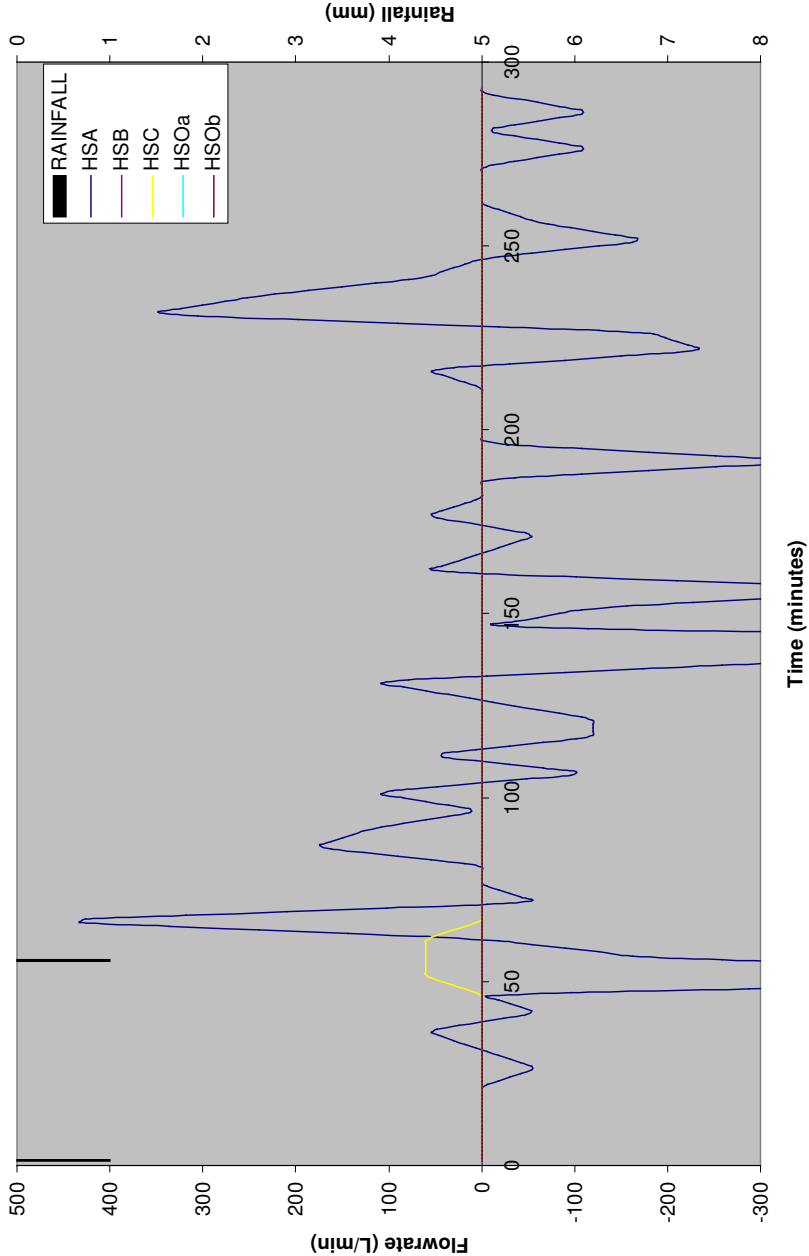
Rainfall in Past 5 Days
0 mm

Rainfall in Past 12 Hours
0 mm

AMC Condition
1 - Completely Dry

Comments
Event is not suitable for use in calibration with DRAINS model. No runoff has been recorded despite rainfall occurring at the gauge. This was perhaps due to a dry catchment condition limiting runoff.

Storm 9 - Burst 1



Time Period
4:52 hours

Total Precipitation
2 mm

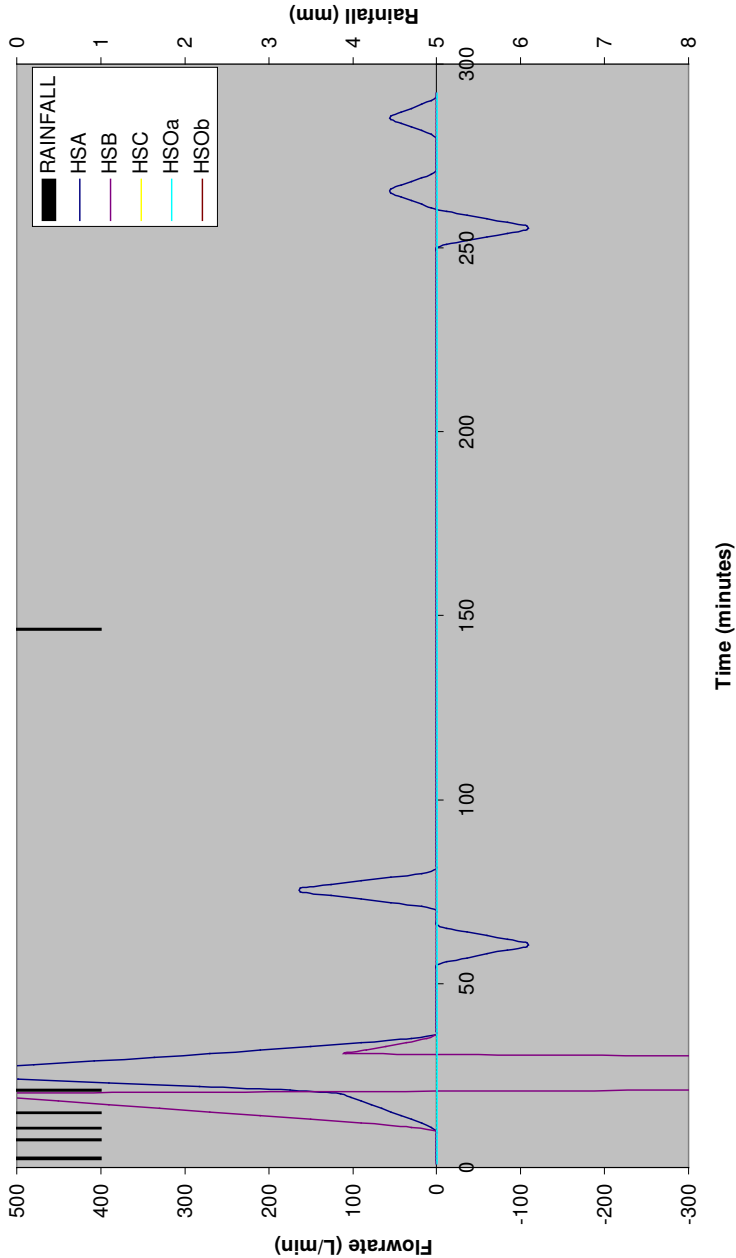
Rainfall in Past 5 Days
0 mm

Rainfall in Past 12 Hours
0 mm

AMC Condition
1 - Completely Dry

Comments
Event is not suitable for use in calibration with DRAINS model. Poor correlation with rainfall is evident and wildly varying negative and positive flowrates indicate instrument error.

Storm 10 - Burst 1



Time Period
4:51 hours

Total Precipitation
6 mm

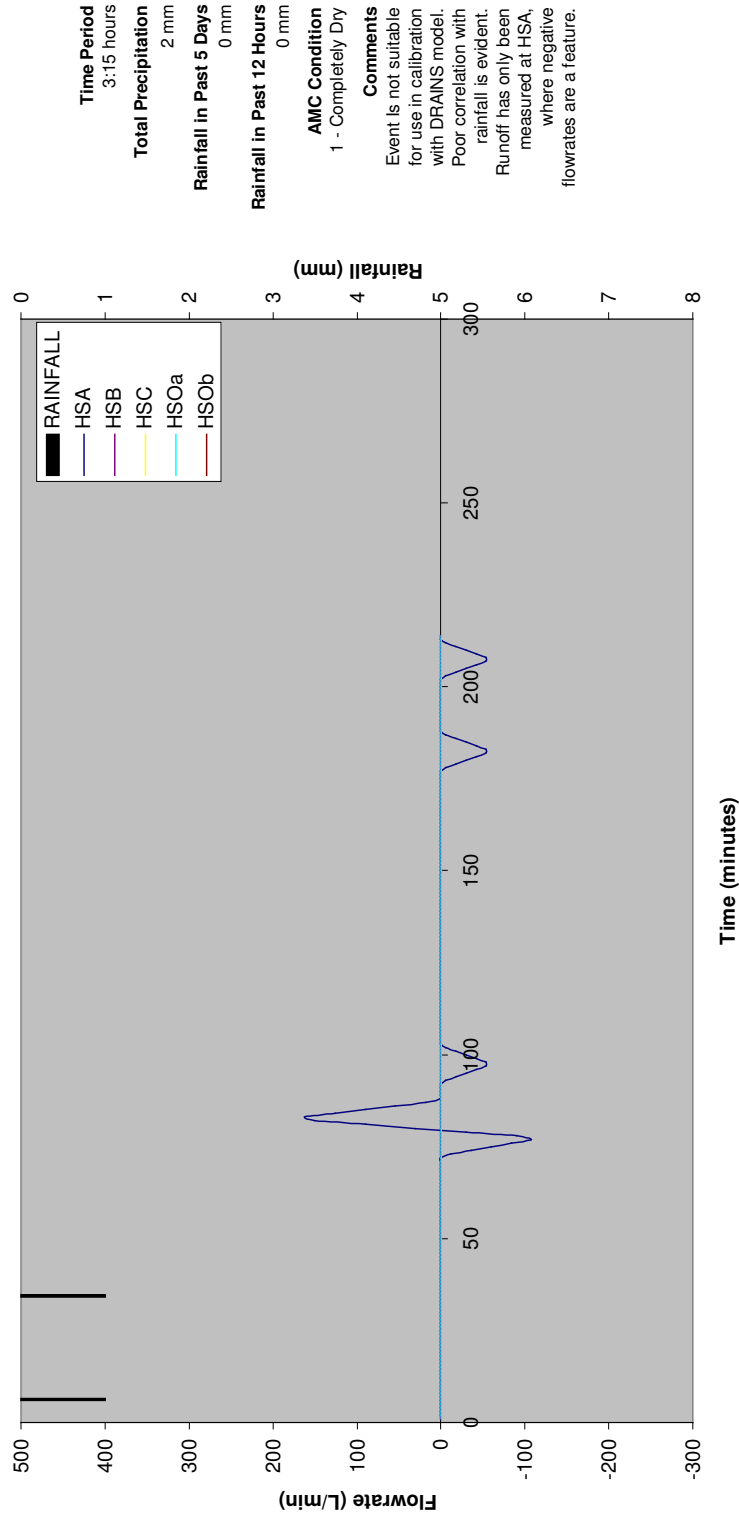
Rainfall in Past 5 Days
0 mm

Rainfall in Past 12 Hours
0 mm

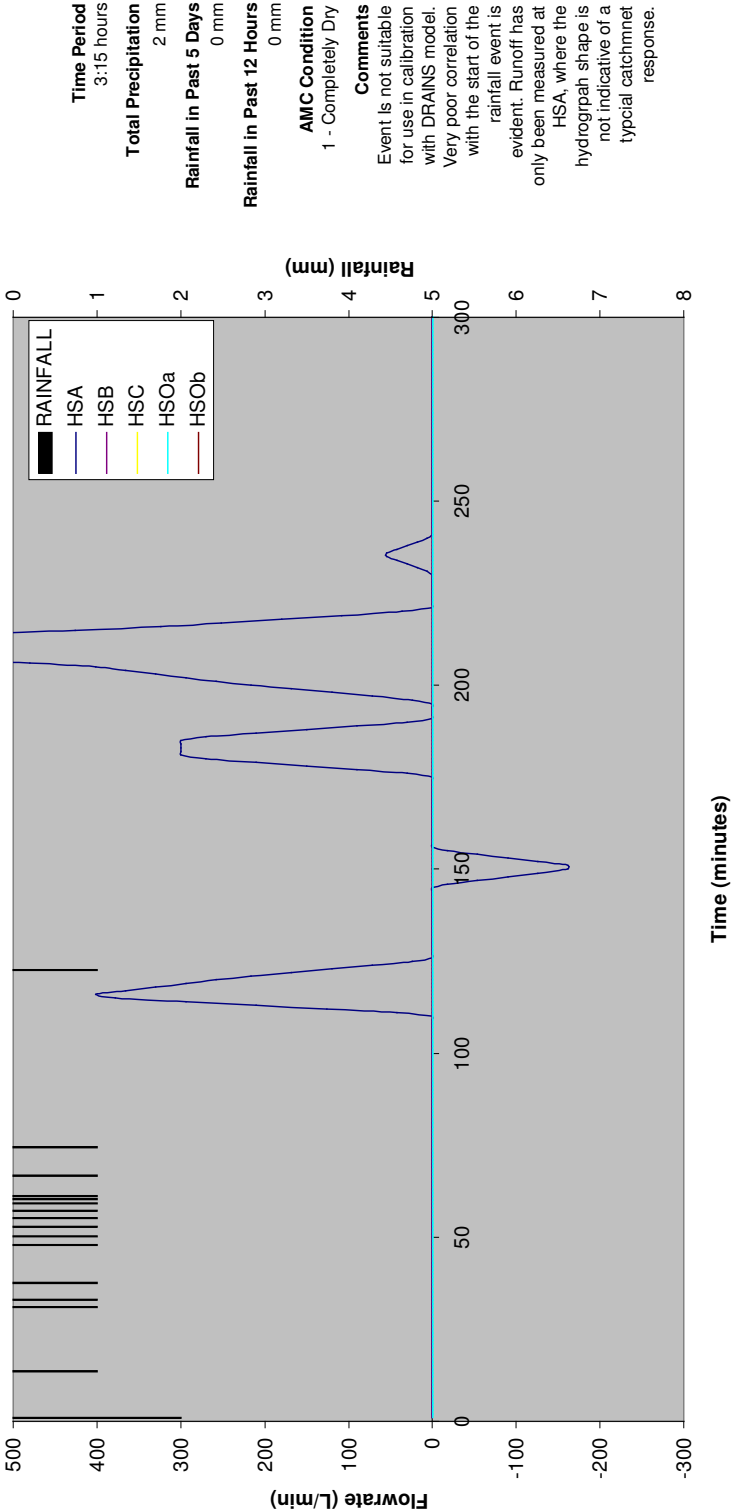
AMC Condition
1 - Completely Dry

Comments
Event is not suitable for use in calibration with DRAINS model. Poor correlation with rainfall is evident. This is coupled with negative flowrates for both HAS and HSB. Instrument error has caused no suitable hydrograph to be recorded for HSC, HSOa and HSOB.

Storm 11 - Burst 1



Storm 11 - Burst 2



Time Period
3:15 hours

Total Precipitation
2 mm

Rainfall in Past 5 Days
0 mm

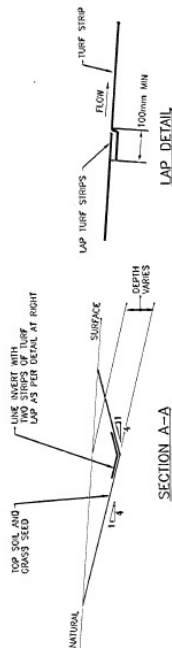
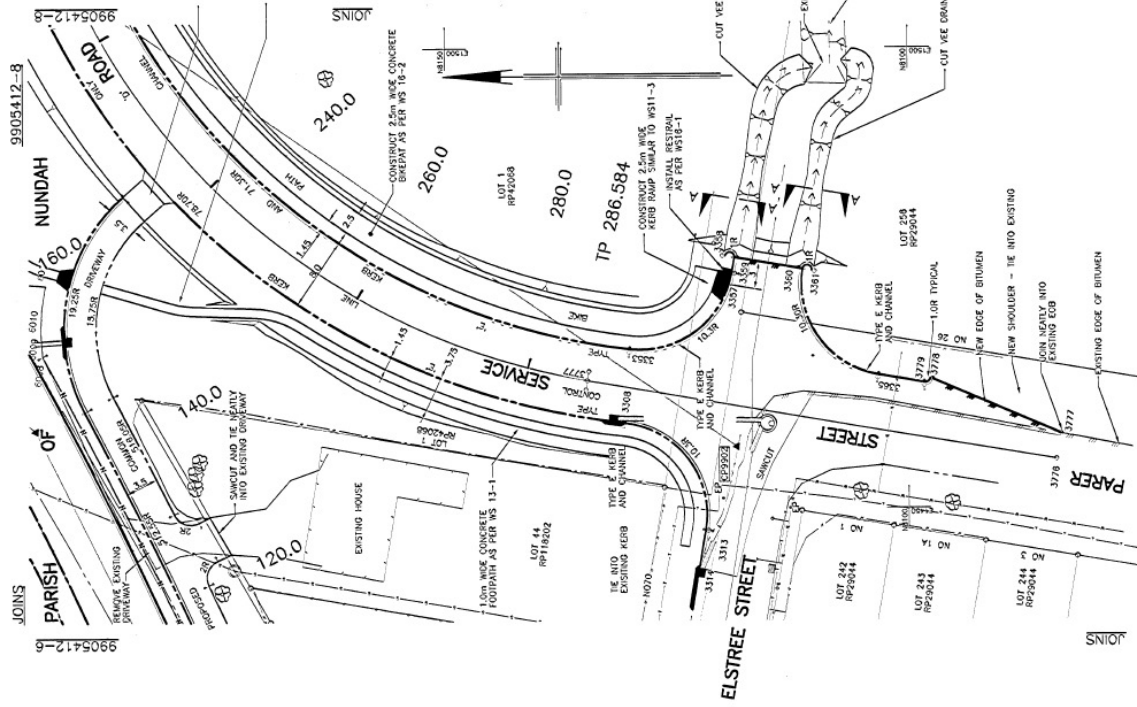
Rainfall in Past 12 Hours
0 mm

AMC Condition
1 - Completely Dry

Comments
Event is not suitable for use in calibration with DRAINS model. Very poor correlation with the start of the rainfall event is evident. Runoff has only been measured at HSA, where the hydrograph shape is not indicative of a typical catchment response.

APPENDIX C: Hoyland Street Connection Drawings

(irrelevant sections omitted)



PLAN



W11733

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CONTRACT	SHEET	SHEET	SHEET	SHEET	SHEET	SHEET	SHEET	SHEET	SHEET

BRISBANE CITY COUNCIL
CITY DESIGN
ROAD AND TRAFFIC DESIGN
DRAWING SHEET 7 OF 13
DRAWING NO. 9905412
SHEET NO. 7
A

JOB
ROADWORKS
AND DRAINAGE
HOTLAND STREET CONNECTION
BRACKEN RIDGE
GYMPIE ROAD TO
BRACKEN RIDGE ROAD



FILE NO.	M.P.	1/2000
SERVICES FILE NO.	R.A.L.	1/2000
ACQUISITION FILE NO.	M.A.M.	1/2000
SHEP FILE NO.	M.A.M.	1/2000

DESIGN	CHECKED	M.P.	1/2000
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CHECKED	M.A.M.	1/2000	1/2000

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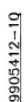
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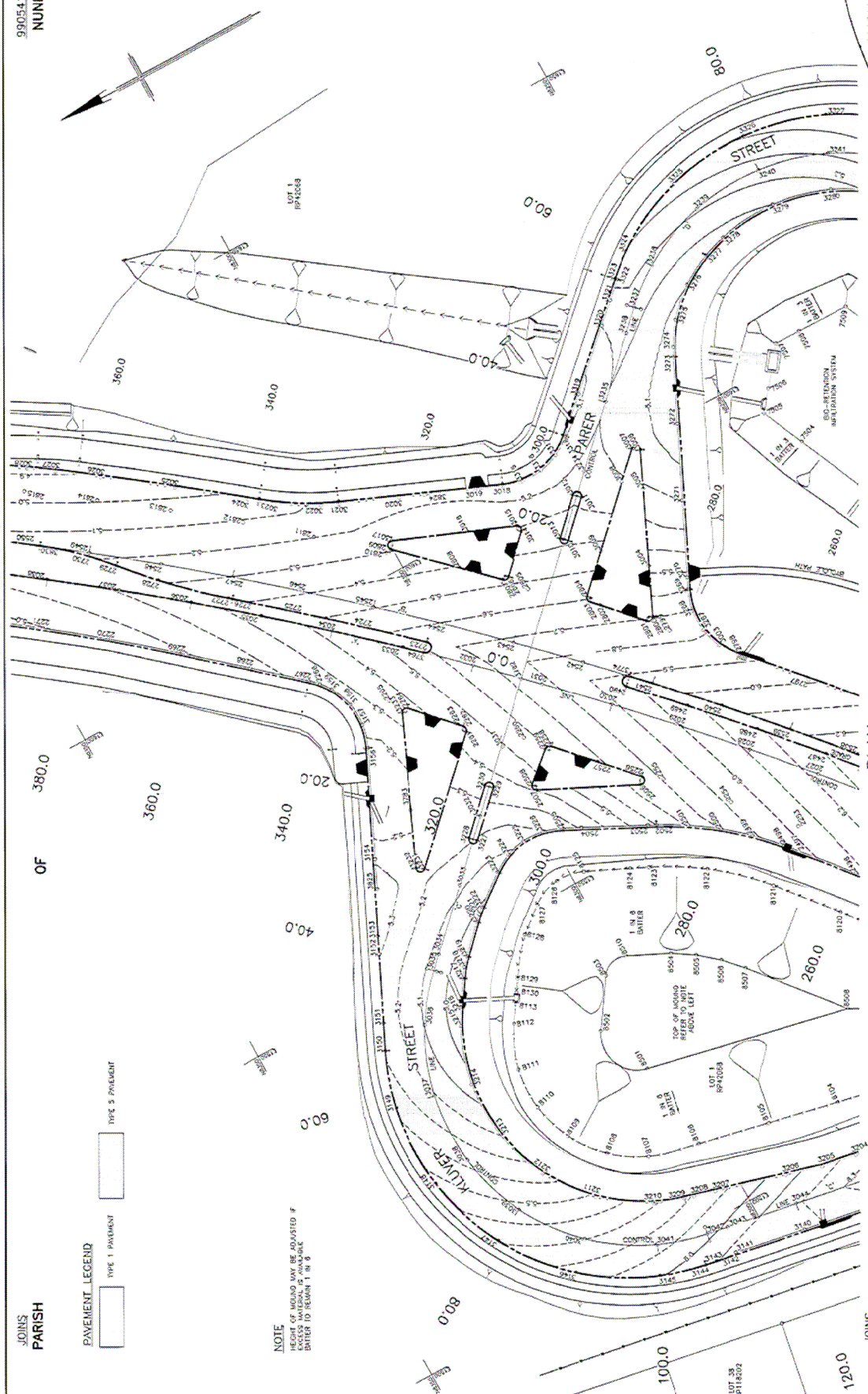
W11733

AMENDMENT		DATE	CHECKED	AUTHOR	AS CONST.	MICRO	COMM'L SURVEYOR IN CHARGE, CONCERNING FIELD BOOK 9104, FIRMNAME 1395-0 REFERENCE TO BM. SURVEYED IN GILGUTH PT. NO. 3199 SEC. 10, T. 23 N. R. 19 E. S. 28 S. 10, T. 23 N. R. 19 E. S. 28		REMARKS CONCERNING THIS SURVEY  J. B. Smith JUNE 1954 IMPRINT COVER - 500 US DOLLAR NOTE 500B SERIAL - 240 053 241 500B DOWNTOWN TRAFFIC - 240 053 241 500B DOWNTOWN TRAFFIC - 240 053 241		CADW REFERENCE 9905412-208-03-049 PLAN CUSTOMER REGISTERED MICROFILM		DESIGN DESIGN CHECKED DRAWN TRACED CHECKED M.L.P. R.L.L. M.M. M.M. P.L.		FILE NO. SERIALIZED FILE NO. ACQUISITION FILE NO. SER. FILE NO. 1/2000 1/2000 1/2000 1/2000 1/2000		 ROADWORKS AND DRAINAGE HOYTAND STREET CONNECTION BRACKEN RIDGE GYMPIE ROAD TO BRACKEN RIDGE ROAD		DRAWING TITLE LAYOUT PLAN SHEET 4 OF 13		BRISBANE CITY COUNCIL CITY DESIGN ROAD AND TRAFFIC DESIGN DRAWING SHEET 9 OF 113 DRAWING NO. 9905412 9 A	
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JOINS
PARISH

TYPE 1 PAVEMENT	TYPE 5 PAVEMENT

NOTE: HEIGHT OF MOUND MAY BE ADJUSTED IF EXCESS MATERIAL IS AVAILABLE
BATTERY TO REMAIN 1 IN. Ø



NOTE
DESIGN CONTOURS AT 0.10M INTERVALS

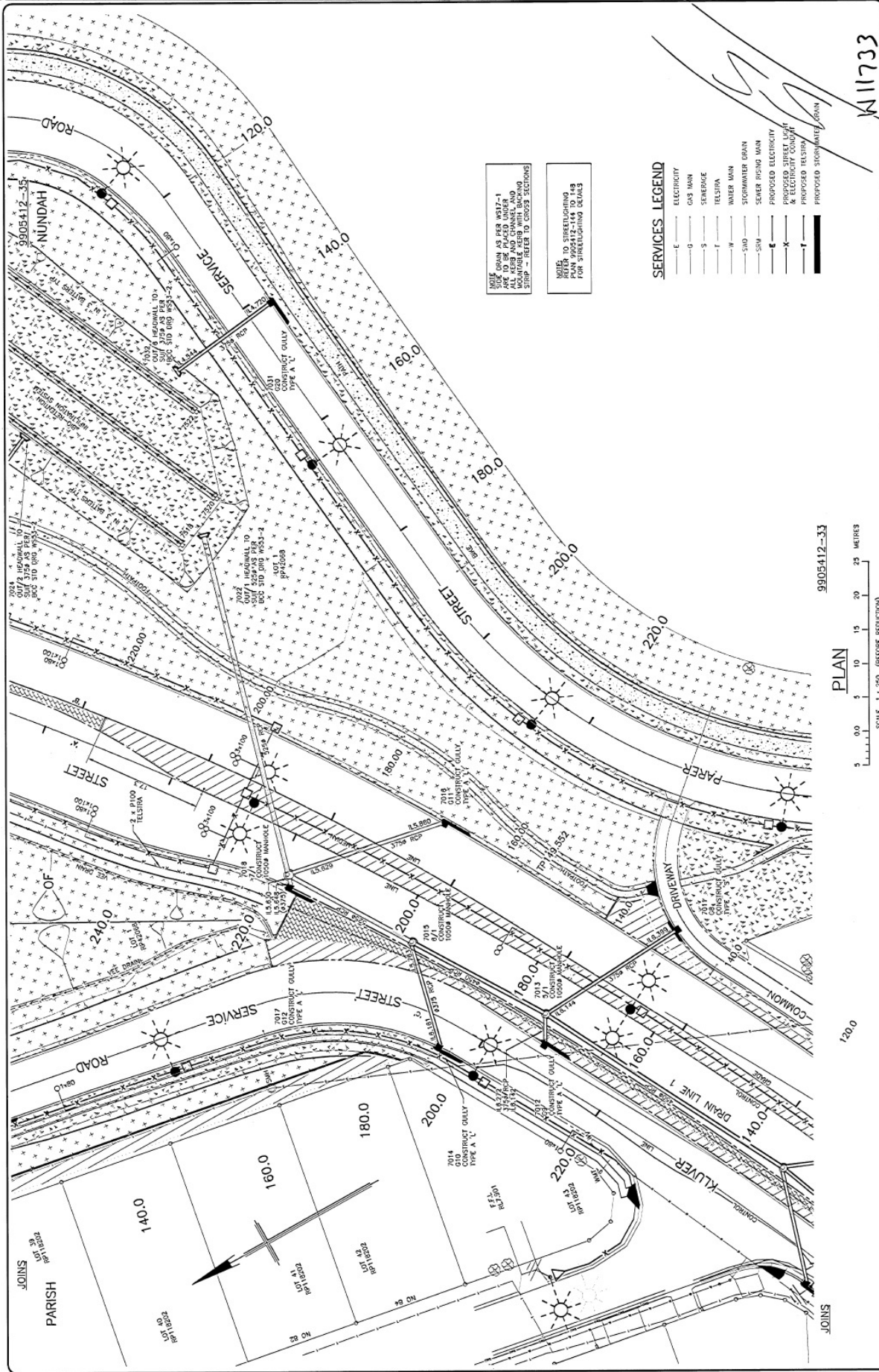
PLAN

SCALE 1 : 250 (BEFORE REDUCTION)

0 5 10 15 20 25 METRES

PLAN

[illegible]



NOTE:
SEE DRAIN AS PER WS17-1
ARE TO BE PLACED UNDER
THE ROAD SURFACE WITH A
LOCATABLE KEB WITH BACKING
STRIP - REFER TO CROSS SECTIONS

NOTE:
REFER TO STRENGTHENING
DETAILS FOR STRENGTHENING
FOR STRENGTHENING DETAILS

- SERVICES LEGEND**
- E — ELECTRICITY
 - G — GAS MAIN
 - S — SEWERAGE
 - T — TELEPHONE
 - W — WATER MAIN
 - SW — STORMWATER DRAIN
 - SS — SEWER RISING MAIN
 - E — PROPOSED ELECTRICITY
 - K — PROPOSED STREET LIGHT
 - T — PROPOSED TELEPHONE
 - — PROPOSED STORMWATER DRAIN



BRISBANE CITY COUNCIL CITY DESIGN ROAD AND TRAFFIC DESIGN DRAWING NO. 9905412-33 SHEET NO. 34 AMEND. A	
DRAWING TITLE ROADWORKS AND DRAINAGE SERVICES AND SURFACE TREATMENT PLAN SHEET 3 OF 13	
PROJECT HOVLAND STREET CONNECTION GYMPIE ROAD TO BRACKEN RIDGE ROAD	
DATE 18.03.08	
CONTRACTOR A. CONSTRUCTION ISSUE	
FIELD BOOK 9104 FILENAME 1305-0 REFERENCE P. 104 DATUM 1.00 DATE 18.03.08 BY J. L. L. CHKD J. L. L. PROJECT HOVLAND STREET CONNECTION PROJECT NO. 9905412-33 PROJECT NAME HOVLAND STREET CONNECTION PROJECT LOCATION HOVLAND STREET CONNECTION PROJECT SCALE 1:250 PROJECT DATE 18.03.08	
DESIGN M.P. 1/2000 CHECKED R.L. 1/2000 DESIGN M.M. 1/2000 TRACED M.M. 1/2000 CHECKED M.M. 1/2000	
CAD REFERENCE 9905412-33-01-04 PLAN CUSTOMER REGISTERED MICROFILM CHECKED	
FORM ACCEPTED AND APPROVED DRAWING NO. 9905412-33-01-04 PROJECT NO. 9905412-33 PROJECT NAME HOVLAND STREET CONNECTION PROJECT LOCATION HOVLAND STREET CONNECTION PROJECT SCALE 1:250 PROJECT DATE 18.03.08	

9905412-36

JOINS
PARISH

SERVICES LEGEND

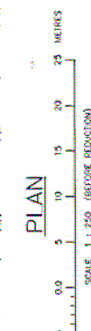
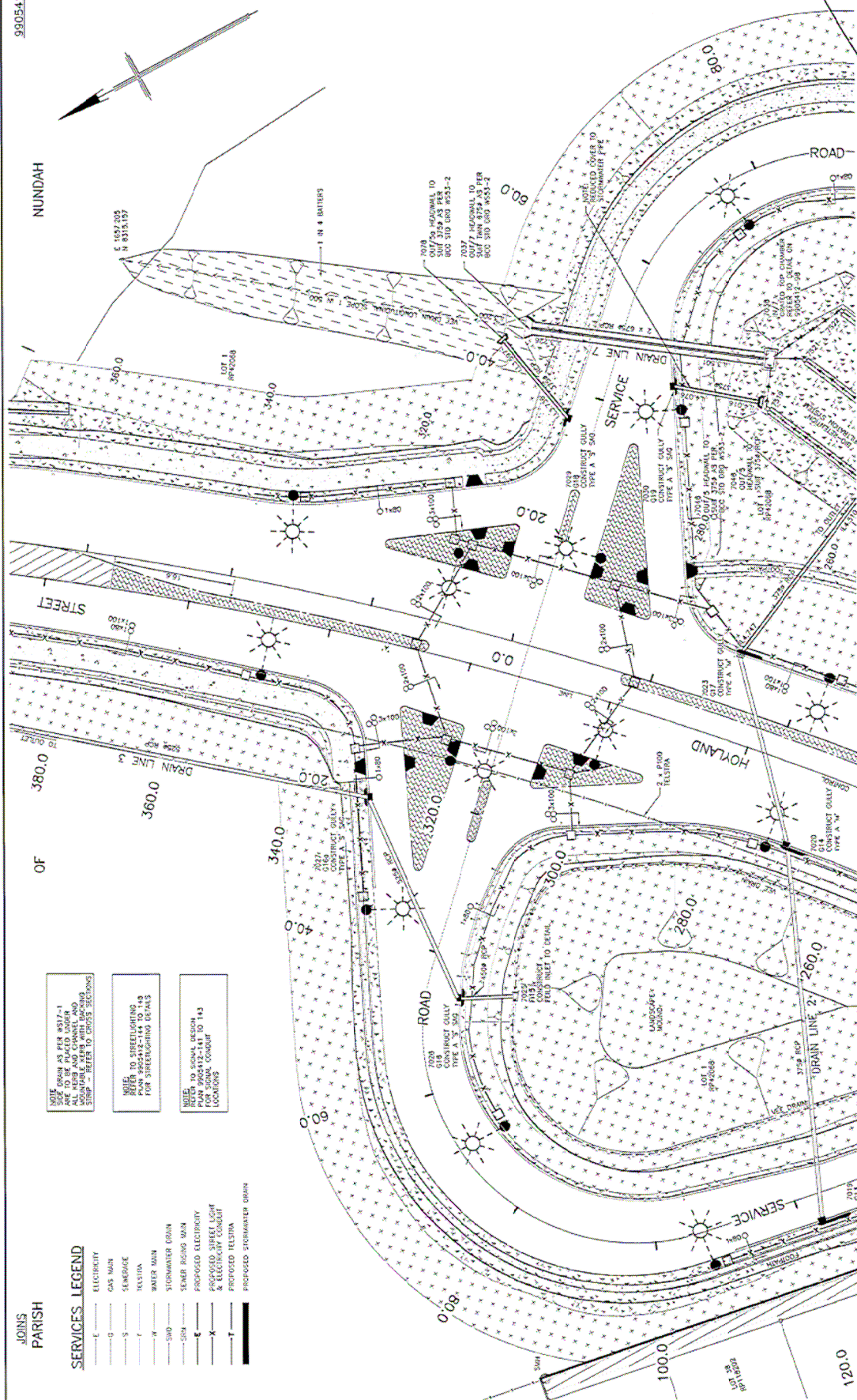
- ELECTRICITY
- GAS MAIN
- SEWERAGE
- TELSTRA
- WATER MAIN
- STORMWATER DRAIN
- SEWER RISING MAIN
- PROPOSED ELECTRICITY
- PROPOSED STREET LIGHT & ELECTRICITY CONDUIT
- PROPOSED TELSTRA
- PROPOSED STORMWATER DRAIN

NOTE: SIZE DRAIN AS PER WS77-1
ALL RISES AND CHANNELS TO BE
CONSTRUCTED TO THE STREET
STOP - REFER TO CROSS SECTIONS

NOTE: REFER TO STREET LIGHTING
PLAN 9905412-143
FOR STREET LIGHTING DETAILS

NOTE: TO SIGNAL DESIGN
PLAN 9905412-141 TO 143
FOR SIGNAL CONDUIT
LOCATIONS

NUNDDAH



PLAN

9905412-34

W11733

AMENDMENT										DATE	CHECKED	AUTHOR	AS CONST.	MICRO	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. 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BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY - NO USE CHANGINGS	FIELD BOOK	9104	REFERENCE - 5 BM	DATE - 1/19/95	DRAWN - 1/19/95	BY - J. N. BLOOM	SCALE - 1" = 40' (SEE SHEET 35 OF 113)	SHEET NO.	9104-328	BL. 1002	CONTRACTOR'S USE ONLY
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SETOUT CO-ORDINATES

LEVEL SETOUT CO-ORDINATES

LEVEL SETOUT CO-ORDINATES

LEVEL SETOUT CO-ORDINATES

LEVEL SETOUT CO-ORDINATES

SETOUT CO-ORDINATES

POINT #	EASTING	NORTHING
5501	1307.751	8173.818
5502	1309.373	8172.918
5503	1310.995	8172.018
5504	1312.617	8171.118
5505	1314.239	8170.218
5506	1315.861	8169.318
5507	1317.483	8168.418
5508	1319.105	8167.518
5509	1320.727	8166.618
5510	1322.349	8165.718
5511	1323.971	8164.818

POINT #	EASTING	NORTHING
6000	1333.545	8170.817
6001	1335.167	8169.917
6002	1336.789	8169.017
6003	1338.411	8168.117
6004	1340.033	8167.217
6005	1341.655	8166.317
6006	1343.277	8165.417
6007	1344.899	8164.517
6008	1346.521	8163.617
6009	1348.143	8162.717
6010	1349.765	8161.817
6011	1351.387	8160.917
6012	1353.009	8160.017
6013	1354.631	8159.117
6014	1356.253	8158.217
6015	1357.875	8157.317
6016	1359.497	8156.417
6017	1361.119	8155.517
6018	1362.741	8154.617
6019	1364.363	8153.717
6020	1365.985	8152.817
6021	1367.607	8151.917
6022	1369.229	8151.017
6023	1370.851	8150.117
6024	1372.473	8149.217
6025	1374.095	8148.317

NOTE: SETOUT TO TOE OF WALL

BIO-RETENSION INFILTRATION SYSTEM

POINT #	EASTING	NORTHING	ELEVATION
7500	1592.872	8228.870	4.900
7501	1594.494	8227.970	4.900
7502	1596.116	8227.070	4.900
7503	1597.738	8226.170	4.900
7504	1599.360	8225.270	4.900
7505	1600.982	8224.370	4.900
7506	1602.604	8223.470	4.900
7507	1604.226	8222.570	4.900
7508	1605.848	8221.670	4.900
7509	1607.470	8220.770	4.900
7510	1609.092	8219.870	4.900
7511	1610.714	8218.970	4.900
7512	1612.336	8218.070	4.900
7513	1613.958	8217.170	4.900
7514	1615.580	8216.270	4.900
7515	1617.202	8215.370	4.900
7516	1618.824	8214.470	4.900
7517	1620.446	8213.570	4.900
7518	1622.068	8212.670	4.900
7519	1623.690	8211.770	4.900
7520	1625.312	8210.870	4.900
7521	1626.934	8209.970	4.900
7522	1628.556	8209.070	4.900
7523	1630.178	8208.170	4.900

LEVEL SETOUT CO-ORDINATES

POINT #	EASTING	NORTHING	ELEVATION
8500	1818.870	8235.358	8.317
8501	1820.492	8234.458	8.317
8502	1822.114	8233.558	8.317
8503	1823.736	8232.658	8.317
8504	1825.358	8231.758	8.317
8505	1826.980	8230.858	8.317
8506	1828.602	8229.958	8.317
8507	1830.224	8229.058	8.317
8508	1831.846	8228.158	8.317
8509	1833.468	8227.258	8.317
8510	1835.090	8226.358	8.317
8511	1836.712	8225.458	8.317
8512	1838.334	8224.558	8.317
8513	1839.956	8223.658	8.317
8514	1841.578	8222.758	8.317
8515	1843.200	8221.858	8.317
8516	1844.822	8220.958	8.317
8517	1846.444	8220.058	8.317
8518	1848.066	8219.158	8.317
8519	1849.688	8218.258	8.317
8520	1851.310	8217.358	8.317
8521	1852.932	8216.458	8.317
8522	1854.554	8215.558	8.317
8523	1856.176	8214.658	8.317
8524	1857.798	8213.758	8.317
8525	1859.420	8212.858	8.317
8526	1861.042	8211.958	8.317
8527	1862.664	8211.058	8.317
8528	1864.286	8210.158	8.317
8529	1865.908	8209.258	8.317
8530	1867.530	8208.358	8.317
8531	1869.152	8207.458	8.317
8532	1870.774	8206.558	8.317
8533	1872.396	8205.658	8.317
8534	1874.018	8204.758	8.317
8535	1875.640	8203.858	8.317
8536	1877.262	8202.958	8.317
8537	1878.884	8202.058	8.317
8538	1880.506	8201.158	8.317
8539	1882.128	8200.258	8.317
8540	1883.750	8199.358	8.317
8541	1885.372	8198.458	8.317
8542	1886.994	8197.558	8.317
8543	1888.616	8196.658	8.317
8544	1890.238	8195.758	8.317
8545	1891.860	8194.858	8.317
8546	1893.482	8193.958	8.317
8547	1895.104	8193.058	8.317
8548	1896.726	8192.158	8.317
8549	1898.348	8191.258	8.317
8550	1900.000	8190.358	8.317
8551	1901.622	8189.458	8.317
8552	1903.244	8188.558	8.317
8553	1904.866	8187.658	8.317
8554	1906.488	8186.758	8.317
8555	1908.110	8185.858	8.317
8556	1909.732	8184.958	8.317
8557	1911.354	8184.058	8.317
8558	1912.976	8183.158	8.317
8559	1914.598	8182.258	8.317
8560	1916.220	8181.358	8.317
8561	1917.842	8180.458	8.317
8562	1919.464	8179.558	8.317
8563	1921.086	8178.658	8.317
8564	1922.708	8177.758	8.317
8565	1924.330	8176.858	8.317
8566	1925.952	8175.958	8.317
8567	1927.574	8175.058	8.317
8568	1929.196	8174.158	8.317
8569	1930.818	8173.258	8.317
8570	1932.440	8172.358	8.317
8571	1934.062	8171.458	8.317
8572	1935.684	8170.558	8.317
8573	1937.306	8169.658	8.317
8574	1938.928	8168.758	8.317
8575	1940.550	8167.858	8.317
8576	1942.172	8166.958	8.317
8577	1943.794	8166.058	8.317
8578	1945.416	8165.158	8.317
8579	1947.038	8164.258	8.317
8580	1948.660	8163.358	8.317
8581	1950.282	8162.458	8.317
8582	1951.904	8161.558	8.317
8583	1953.526	8160.658	8.317
8584	1955.148	8159.758	8.317
8585	1956.770	8158.858	8.317
8586	1958.392	8157.958	8.317
8587	1960.014	8157.058	8.317
8588	1961.636	8156.158	8.317
8589	1963.258	8155.258	8.317
8590	1964.880	8154.358	8.317
8591	1966.502	8153.458	8.317
8592	1968.124	8152.558	8.317
8593	1969.746	8151.658	8.317
8594	1971.368	8150.758	8.317
8595	1972.990	8149.858	8.317
8596	1974.612	8148.958	8.317
8597	1976.234	8148.058	8.317
8598	1977.856	8147.158	8.317
8599	1979.478	8146.258	8.317
8600	1981.100	8145.358	8.317
8601	1982.722	8144.458	8.317
8602	1984.344	8143.558	8.317
8603	1985.966	8142.658	8.317
8604	1987.588	8141.758	8.317
8605	1989.210	8140.858	8.317
8606	1990.832	8139.958	8.317
8607	1992.454	8139.058	8.317
8608	1994.076	8138.158	8.317
8609	1995.698	8137.258	8.317
8610	1997.320	8136.358	8.317
8611	1998.942	8135.458	8.317
8612	2000.564	8134.558	8.317
8613	2002.186	8133.658	8.317
8614	2003.808	8132.758	8.317
8615	2005.430	8131.858	8.317
8616	2007.052	8130.958	8.317
8617	2008.674	8130.058	8.317
8618	2010.296	8129.158	8.317
8619	2011.918	8128.258	8.317
8620	2013.540	8127.358	8.317
8621	2015.162	8126.458	8.317
8622	2016.784	8125.558	8.317
8623	2018.406	8124.658	8.317
8624	2020.028	8123.758	8.317
8625	2021.650	8122.858	8.317
8626	2023.272	8121.958	8.317
8627	2024.894	8121.058	8.317
8628	2026.516	8120.158	8.317
8629	2028.138	8119.258	8.317
8630	2029.760	8118.358	8.317
8631	2031.382	8117.458	8.317
8632	2033.004	8116.558	8.317
8633	2034.626	8115.658	8.317
8634	2036.248	8114.758	8.317
8635	2037.870	8113.858	8.317
8636	2039.492	8112.958	8.317
8637	2041.114	8112.058	8.317
8638	2042.736	8111.158	8.317
8639	2044.358	8110.258	8.317
8640	2045.980	8109.358	8.317
8641	2047.602	8108.458	8.317
8642	2049.224	8107.558	8.317
8643	2050.846	8106.658	8.317
8644	2052.468	8105.758	8.317
8645	2054.090	8104.858	8.317
8646	2055.712	8103.958	8.317
8647	2057.334	8103.058	8.317
8648	2058.956	8102.158	8.317
8649	2060.578	8101.258	8.317
8650	2062.200	8100.358	8.317
8651	2063.822	8099.458	8.317
8652	2065.444	8098.558	8.317
8653	2067.066	8097.658	8.317
8654	2068.688	8096.758	8.317
8655	2070.310	8095.858	8.317
8656	2071.932	8094.958	8.317
8657	2073.554	8094.058	8.317
8658	2075.176	8093.158	8.317
8659	2076.798	8092.258	8.317
8660	2078.420	8091.358	8.317
8661	2080.042	8090.458	8.317
8662	2081.664	8089.558	8.317
8663	2083.286	8088.658	8.317
8664	2084.908	8087.758	8.317
8665	2086.530	8086.858	8.317
8666	2088.152	8085.958	8.317
8667	2089.774	8085.058	8.317
8668	2091.396	8084.158	8.317
8669	2093.018	8083.258	8.317
8670	2094.640	8082.358	8.317
8671	2096.262	8081.458	8.317
8672	2097.884	8080.558	8.317
8673	2100.006	8079.658	8.317
8674	2101.628	8078.758	8.317
8675	2103.250	8077.858	8.317
8676	2104.872	8076.958	8.317
8677	2106.494	8076.058	8.317
8678	2108.116	8075.158	8.317
8679	2109.738	8074.258	8.317

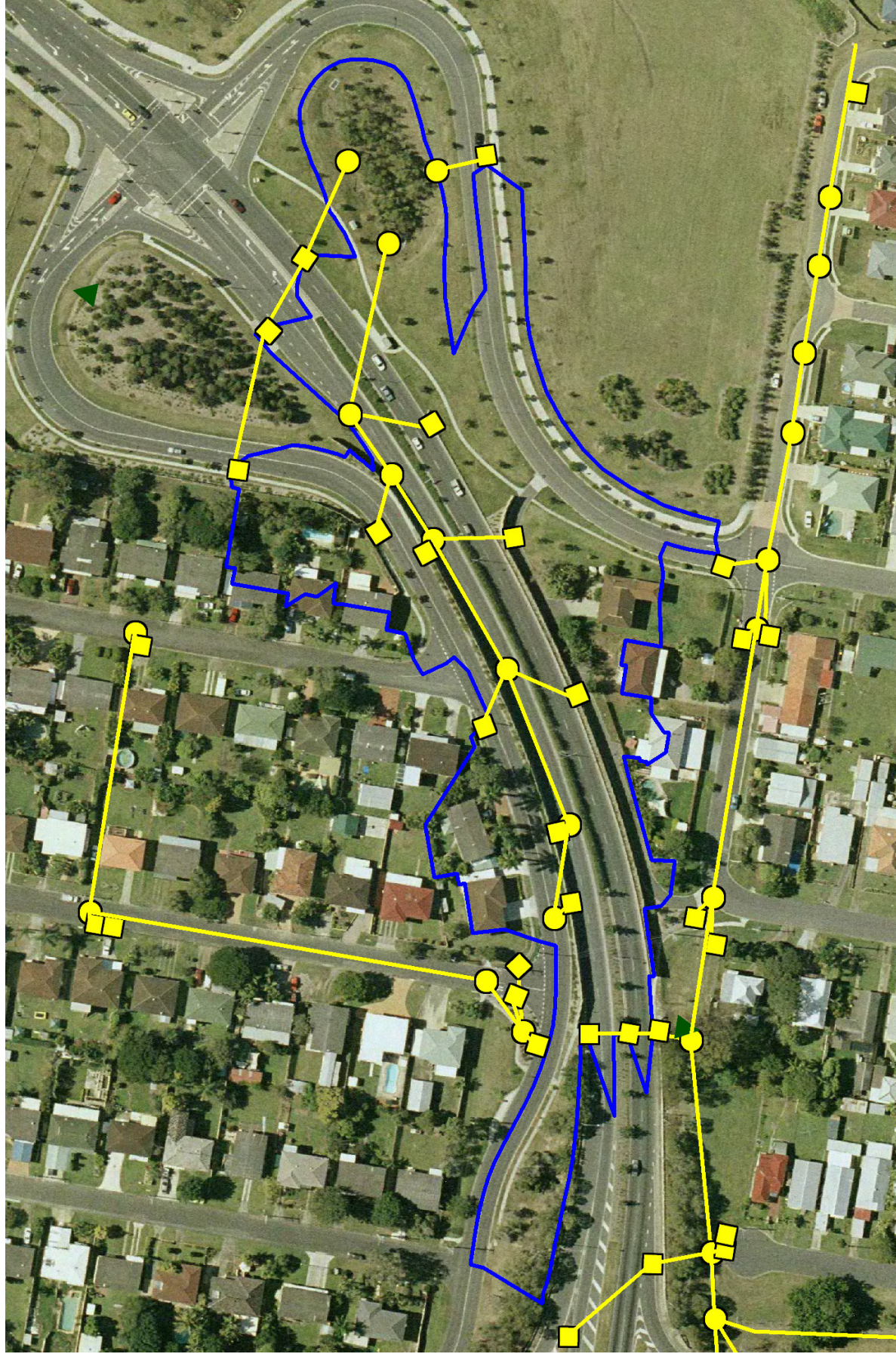
GULLY SETOUT TABLE

POINT #	GULLY NUMBER	EASTING	NORTHING	DESCRIPTION	UP LEVEL	LENGTH	CONNECTION SIZE (mm)	POUR UP LEVEL APPLICABLE
7000	G1	1319.669	8182.914	L LINTEL	8.731	5.00	375	
7001	G2	1326.595	8192.250	S LINTEL (\$560)	9.397	5.70	375	
7002	G3	1340.651	8193.380	S LINTEL (\$560)	8.498	6.30	375	
7003	G4	1357.771	8174.554	L LINTEL (\$560)	8.370	8.30	375	
7004	G5	1371.849	8200.091	M LINTEL	8.276	17.50	375	
7005	G6	1422.344	8173.598	M LINTEL (\$560)	7.785	22.50	375	
7006	G7	1458.373	8191.282	S LINTEL (\$560)	7.391	22.50	375	
7007	G8	1453.187	8215.566	L LINTEL	7.583	4.50	375	
7008	G9	1470.293	8225.992	L LINTEL	7.283	16.00	375	
7009	G10	1498.758	8212.719	L LINTEL	6.829	23.00	375	
7010	G11	1501.759	8252.421	L LINTEL	6.800	32.50	375	
7011	G12	1539.125	8262.421	L LINTEL	6.800	32.50	375	
7012	G13	1540.282	8272.112	M LINTEL	3.951	27.00	375	
7013	G14	1539.986	8314.743	FIELD INLET	3.000	7.50	450	
7014	G15	1534.324	8321.266	S LINTEL (\$560)	4.934	30.30	375	
7015	G16	1574.024	8315.383	S LINTEL (\$560)	5.078	91.00	375	
7016	G17	1556.795	8264.396	M LINTEL (\$560)	9.981	30.00	375	
7017	G18	1607.091	8264.082	S LINTEL (\$560)	9.843	14.50	375	
7018	G19	1578.018	8197.513	S LINTEL (\$560)	3.983	15.50	375	
7019	G20	1578.018	8197.513	L LINTEL	3.983	17.00	375	
7020	G21	1443.098	8182.737	S LINTEL	7.627	16.50	375	
7021	G22	1452.447	8312.792	S LINTEL	7.411	16.50	375	
7022	G23	1600.615	8241.486	GRADED TOP MANHOLE	4.500	33.00	1/4IN 6/3	
7023	G24	1639.793	8343.075	REFER DETAIL				
7024	G25	1630.236	8348.072	S LINTEL	4.877	8.50	375	
7025	G26	1584.861	8348.329	P/C FUTURE L	4.823	11.50	375	
7026	G27	1744.111	8331.649	P/C FUTURE S	3.470	42.00	375	
7027	G28	1753.701	8494.746	P/C FUTURE M	3.413	18.00	375	
7028	G29	1707.538	8503.746	P/C FUTURE L	3.456	12.40	375	
7029	HW13A	1752.291	8393.680	FIELD INLET	3.500	11.50	375	
7030	G30	1737.469	8510.054	P/C FUTURE M	3.543	35.00	375	
7031	F131	1608.637	8627.605	1500 DIA	4.745	35.50	500	
7032	G31	1807.072	8645.396	P/C FUTURE S	3.290	14.70	375	
7033	G32	1791.097	8654.683	P/C FUTURE L	3.479	10.00	375	

MANHOLE SETOUT TABLE

POINT #	MANHOLE NUMBER	EASTING	NORTHING	INVERT LEVEL	MANHOLE CONSTRUCTION
7000	M1	1318.791	8182.914	8.500	1500
7001	M2	1348.803	8178.123	8.875	1500
7002	M3	1383.397	8178.251	8.975	1050
7003	M4	1430.282	8194.438	7.250	1050
7004	M5	1458.914	8213.658	6.215	1050
7005	M6	1501.759	8200.091	6.100	1050
7006	M7	1501.759	8200.091	6.872	1200
7007	M8	1558.177	8153.238	7.520	1500
7008	M9	1626.097	8381.449	3.970	1050
7009	M10	1630.236	8348.072	4.000	1050
7010	M11	1630.236	8348.072	4.100	1050
7011	M12	1630.236	8348.072	4.100	1050
7012	M13	1630.236	8348.072	4.100	1050
7013	M14	1630.236	8348.072	4.100	1050
7014	M15	1630.236	8348.072	4.100	1050
7015	M16	1630.236	8348.072	4.100	1050
7016	M17	1630.236	8348.072	4.100	1050
7017	M18	1630.236	8348.072	4.100	1050
7018	M19	1630.236	8348.072	4.100	1050
7019	M20	1630.236	8348.072	4.100	1050
7020	M21	1630.236	8348.072	4.100	1050
7021	M22	1630.236	8348.072	4.100	1050
7022	M23	1630.236	8348.072	4.100	1050
7023	M24	1630.236	8348.072	4.100	1050
7024	M25	1630.236	8348.072	4.100	1050
7025	M26	1630.236	8348.072	4.100	1050
7026	M27	1630.236	8348.072	4.100	1050
7027	M28	1630.236	8348.072	4.100	1050
7028	M29	1630.236	8348.072	4.100	1050
7029	M30	1630.236	8348.072	4.100	1050
7030	M31	1630.236	8348.072	4.100	1050
7031	M32	1630.236	8348.072	4.100	1050
7032	M33	1630.236	8348.072	4.100	1050
7033	M34	1630.236	8348.072	4.100	1050
7034	M35	1630.236	8348.072	4.100	1050
7035	M36	1630.236	8348.072	4.100	1050
7036	M37	1630.236	8348.072	4.100	1050
7037	M38	1630.236	8348.072	4.100	1050
7038	M39	1630.236	8348.072	4.100	1050
7039	M40	1630.236	8348.072	4.100	1050
7040	M41	1630.236	8348.072	4.100	1050
7041	M42	1630.236	8348.072	4.100	1050
7042	M43	1630.236	8348.072	4.100	1050
7043	M44	1630.236	8348.072	4.100	1050
7044	M45	1630.236	8348.072	4.100	1050
7045	M46	1630.236	8348.072	4.100	1050
7046	M47	1630.236	8348.072	4.100	1050
7047	M48	1630.236	8348.072	4.100	1050
7048	M49	1630.236	8348.072	4.100	1050
7049	M50	1630.236	8348.072	4.100	1050
7050	M51	1630.236	8348.072	4.100	1050
7051	M52	1630.236	8348.072	4.100	1050
7052	M53	1630.236	8348.072	4.100	1050
7053	M54	1630.236	8348.072	4.100	1050
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7056	M57	1630.236	8348.072	4.100	1050
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7059	M60	1630.236	8348.072	4.100	1050
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7061	M62	1630.236	8348.072	4.100	1050
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7066	M67	1630.236	8348.072	4.100	1050
7067	M68	1630.236	8348.072	4.100	1050
7068	M69	1630.236	8348.072	4.100	1050
7069	M70	1630.236	8348.072	4.100	1050
7070	M71	1630.236	8348.072	4.100	1050
7071	M72	1630.236	8348.072	4.100	1050
7072	M73	1630.236	8348.072	4.100	1050
7073	M74	1630.236	8348.072	4.100	1050
7074	M75	1630.236	8348.072	4.100	1050
7075	M76	1630.236	8348.072	4.100	1050
7076	M77	1630.236	8348.072	4.100	1050
7077	M78	1630.236	8348.072	4.100	1050
7078	M79	1630.236	8348.072	4.100	1050
7079	M80	1630.236	8348.072	4.100	1050
7080	M81	1630.236	8348.072	4.100	1050
7081	M82	1630.236	8348.072	4.100	1050
7082	M83	1630.236	8348.072	4.100	1050
7083	M84	1630.236	8348.072	4.100	1050
7084	M85	1630.236	8348.072	4.100	1050
7085	M86	1630.236	8348.072	4.100	1050
7086	M87	1630.236	8348.072	4.100	1050
7087	M88	1630.236	8348.072	4.100	1050
7088	M89	1630.236	8348.072	4.100	1050
7089	M90	1630.236	8348.072	4.100	1050
7090	M91	1630.236	8348.072	4.100	1050
7091	M92	1630.236	8348.072	4.100	1050
7092	M93	1630.236	8348.072	4.100	1050
7093	M94	1630.236	8348.072	4.100	1050
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7095	M96	1630.236	8348.072	4.100	1050
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7098	M99	1630.236	8348.072	4.100	1050
7099	M100	1630.236	8348.072	4.100	1050
7100	M101	1630.236	8348.072	4.100	1050
7101	M102	1630.236	8348.072	4.100	1050
7102	M103	1630.236	8348.072	4.100	1050
7103	M104	1630.236	8348.072	4.100	1050
7104	M105	1630.236	8348.072	4.100	1050
7105	M106	1630.236	8348.072	4.100	1050
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7110	M111	1630.236	8348.072	4.100	1050
7111	M112	1630.236	8348.072	4.100	1050
7112	M113	1630.236	8348.072	4.100	1050
7113	M114	1630.236	8348.072	4.100	1050
7114	M115	1630.236	8348.072	4.100	1050
7115	M116	1630.236	8348.072	4.100	1050
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7117	M118	1630.236	8348.072	4.100	1050
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7120	M121	1630.236	8348.072	4.100	1050
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7124	M125	1630.236	8348.072	4.100	1050
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7126	M127	1630.236	8348.072	4.100	1050
7127	M128	1630.236	8348.072	4.100	1050
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7130	M131	1630.236	8348.072	4.100	1050
7131	M132	1630.236	8348.072	4.100	1050
7132	M133	1630.236	8348.072	4.100	1050
7133	M134	1630.236	8348.072	4.100	1050
7134	M135	1630.236	8348.072	4.100	1050
7135	M136	1630.236	8348.072	4.100	1050
7136	M137	1630.236	8348.072	4.100	1050
7137	M138	1630.236	8348.072	4.100	1050

APPENDIX D: Amended Drainage Infrastructure



APPENDIX E: Hoyland Street Catchments



APPENDIX F: DRAINS Pipe and Catchment Output

PIT / NODE DETAILS										Version 9				
Name	Type	Family	Size	Ponding Volume (cu.m)	Pressure Change Coeff. Ku	Surface Elev (m)	Max Pond Depth (m)	Base Inflow (cu.m/s)	Blocking Factor	x	y	Bolt-down lid	id	Part Full Shock Loss
G4	OnGrade	Brisbane CL Lintel			5	8.462		0	0	102	-393	No		130 1 x Ku
M 2/1	OnGrade	Brisbane CL Lintel			1	8.558		0	0	110	-348	Yes		324 1 x Ku
M 3/1	OnGrade	Brisbane CL Lintel			1.5	8.451		0	0	187	-373	Yes		317 1 x Ku
M 4/1	OnGrade	Brisbane CL Lintel			0.2	7.77		0	0	312	-339	Yes		313 1 x Ku
M 5/1	OnGrade	Brisbane CL Lintel			1.5	7.224		0	0	449	-298	Yes		310 1 x Ku
M 6/1	OnGrade	Brisbane CL Lintel			1.5	6.94		0	0	546	-270	Yes		300 1 x Ku
M 7/1	OnGrade	Brisbane CL Lintel			2	6.687		0	0	626	-246	Yes		291 1 x Ku
N117	Node					2.5		0		976	-326			413
G5	OnGrade	Brisbane CM Lintel			5	8.345		0	0	187	-318	No		136 1 x Ku
G6	OnGrade	Brisbane CM Lintel			5	8.217		0	0	272	-309	No		150 1 x Ku
G9	OnGrade	Brisbane CL Lintel			5	7.519		0	0	429	-278	No		152 1 x Ku
G8	Sag	Brisbane CS Lintel		10	5	7.309	0.5	0	0	454	-408	No		153 1 x Ku
G10	OnGrade	Brisbane CL Lintel			5	7.212		0	0	473	-235	No		154 1 x Ku
G11	OnGrade	Brisbane CL Lintel			5	6.757		0	0	616	-332	No		155 1 x Ku
G20	OnGrade	Brisbane CL Lintel			2	5.933		0	0	916	-411	No		156 1 x Ku
G13	OnGrade	Brisbane CL Lintel			5	6.423		0	0	582	-105	No		170 1 x Ku
G14	OnGrade	Brisbane CM Lintel			0.5	6.069		0	0	735	-173	No		177 1 x Ku
G17	OnGrade	Brisbane CM Lintel			0.5	6.065		0	0	895	-196	No		171 1 x Ku
G7	OnGrade	Brisbane CM Lintel			5	7.75		0	0	305	-419	No		184 1 x Ku
G12	OnGrade	Brisbane CL Lintel			5	6.607		0	0	583	-202	No		337 1 x Ku

DETENTION BASIN DETAILS

Name	Elev	Surf. Area	Init Vol. (cu	Outlet Type	K	Dia(mm)	Centre RL	Pit Family	Pit Type	x	y	HED	Crest RL	Crest Leng	id
Basin3	3.227	1400		0 Orifice		675	3.5645				890	-334	No		398
	3.3	1400													
	3.5	1400													
	3.6	1400													
	3.65	1400													
	3.66	1400													
	3.67	1400													

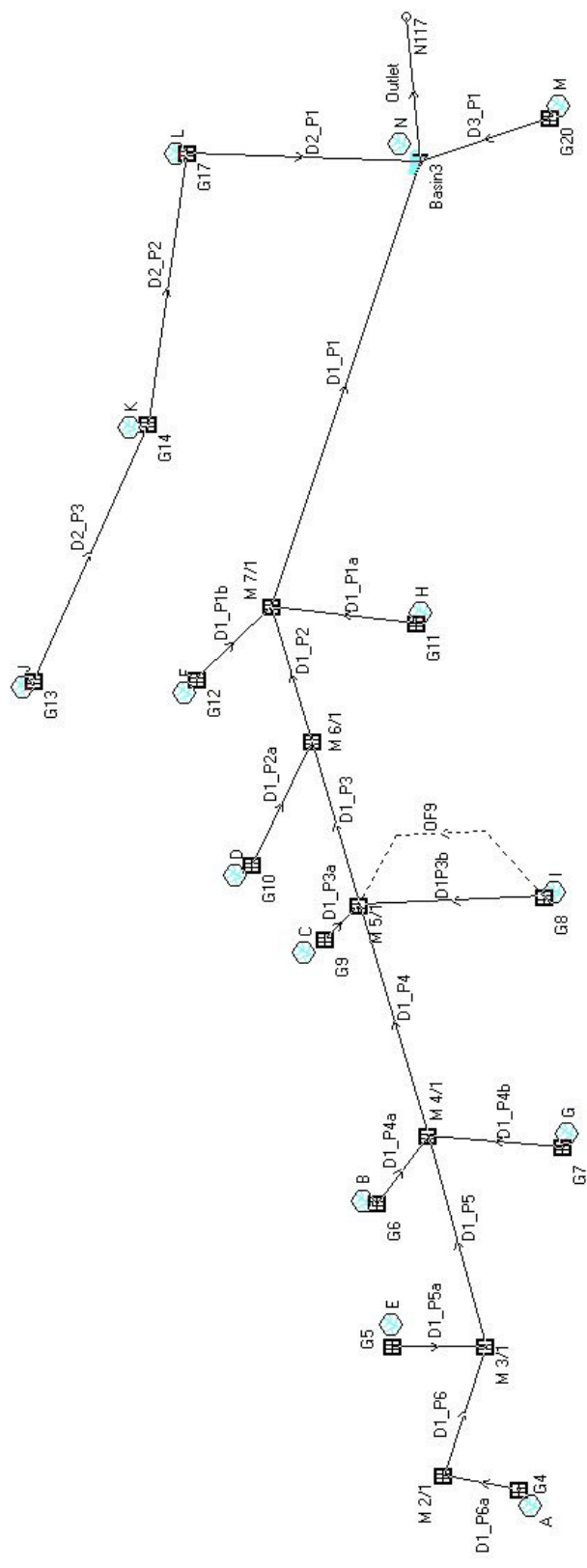
SUB-CATCHMENT DETAILS

Name	Pit or Node	Total Area (ha)	Paved Area %	Grass Area %	Supp Area %	Paved Time (min)	Grass Time (min)	Supp Time (min)	Paved Length (m)	Grass Length (m)	Supp Length (m)	Paved Slope(%)	Grass Slope %	Supp Slope %	Paved Rough	Grass Rough	Supp Rough
A	G4	0.1667	25.8	67.1	7.1	0	25	25	110	20	20	0.5	3	3	0.015	0.13	0.015
N	Basin3	0.2579	0	96.4	3.6	0	0	0	-1	50	50	-1	2	2	-1	0.13	0.015
E	G5	0.0619	100	0	0	0	0	0	82	-1	-1	3.3	-1	-1	0.015	-1	-1
B	G6	0.0967	50	42	8	0	10	10	72	43	30	1	1	1	0.015	0.13	0.015
C	G9	0.0651	68.9	31.1	0	0	10	0	110	1.5	-1	1.5	0.5	-1	0.015	0.13	-1
I	G8	0.1991	42.6	56.3	1.1	0	60	2	120	35	30	1.5	0.5	0.5	0.015	0.13	0.015
D	G10	0.0596	33.6	53.2	13.2	0	60	20	42	12	2	1.5	2	2	0.015	0.13	0.015
H	G11	0.1249	50.6	49.4	0	0	60	0	91	27	-1	1	1	-1	0.015	0.13	-1
M	G20	0.2969	34.3	50.4	15.2	0	10	0	150	29	2	1	1.7	0.5	0.015	0.13	0.015
J	G13	0.0775	66.7	25.1	8.1	0	20	20	65	65	70	1.5	0.5	0.5	0.015	0.13	0.015
K	G14	0.0413	100	0	0	0	0	0	45	-1	-1	1	-1	-1	0.015	-1	-1
L	G17	0.1099	49.7	50.3	0	0	25	0	70	6	-1	1	0.5	-1	0.015	0.13	-1
G	G7	0.1343	61.4	38.6	0	0	15	0	120	9	-1	2.6	1	-1	0.015	0.13	-1
F	G12	0.1251	92.1	7.9	0	0	0	0	140	1.94	-1	1.2	0.5	-1	0.015	0.13	-1

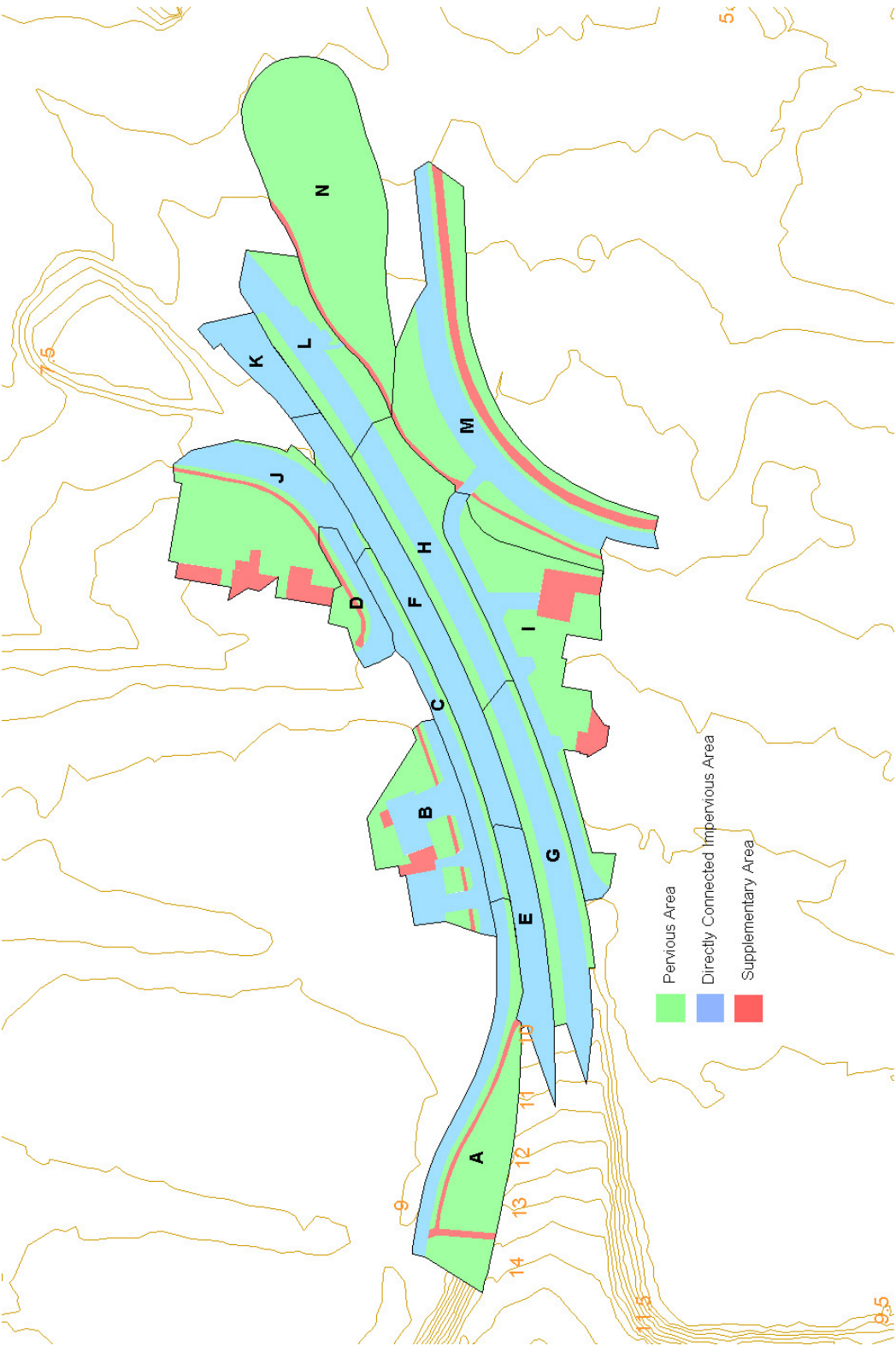
PIPE DETAILS

Name	From	To	Length (m)	U/S IL (m)	D/S IL (m)	Slope (%)	Type	Dia (mm)	I.D. (mm)	Rough	Pipe Is	No. Pipes	Chg From	At Chg	Chg (m)	RI (m)	Chg (m)
D1_P6a	G4	M 2/1	4.30323	7.436	6.858	13.43	Concrete, t	375	375	0.3	Existing	1	G4		0		
D1_P6	M 2/1	M 3/1	23.025	6.828	6.691	0.6	Concrete, t	375	375	0.3	New	1	M 2/1		0		
D1_P5	M 3/1	M 4/1	46.591	6.621	6.22	0.86	Concrete, t	375	375	0.3	Existing	1	M 3/1		0		
D1_P4	M 4/1	M 5/1	41.4138	6.18	5.789	0.94	Concrete, t	450	450	0.3	Existing	1	M 4/1		0		
D1_P3	M 5/1	M 6/1	22.1032	5.719	5.41	1.4	Concrete, t	450	450	0.3	New	1	M 5/1		0		
D1_P2	M 6/1	M 7/1	20.9356	5.365	5.237	0.61	Concrete, t	525	525	0.3	Existing	1	M 6/1		0		
D1_P1	M 7/1	Basin3	51.3529	5.212	4.964	0.48	Concrete, t	525	525	0.3	NewFixed	1	M 7/1		0		
Outlet	Basin3	N117	60	3.227	2.5	1.21	Concrete, t	675	675	0.3	Existing	2	Basin3		0		
D1_P5a	G5	M 3/1	2.52386	7.438	7.301	5.43	Concrete, t	375	375	0.3	Existing	1	G5		0		
D1_P4a	G6	M 4/1	19.1604	7.313	6.23	5.65	Concrete, t	375	375	0.3	Existing	1	G6		0		
D1_P3a	G9	M 5/1	4.819	5.919	5.819	2.08	Concrete, t	375	375	0.3	Existing	1	G9		0		
D1P3b	G8	M 5/1	22.0933	5.889	5.779	0.5	Concrete, t	375	375	0.3	New	1	G8		0		
D1_P2a	G10	M 6/1	15.7087	5.823	5.43	2.5	Concrete, t	375	375	0.3	Existing	1	G10		0		
D1_P1a	G11	M 7/1	23.7636	5.586	5.312	1.15	Concrete, t	375	375	0.3	Existing	1	G11		0		
D3_P1	G20	Basin3	18.6402	4.968	4.756	1.14	Concrete, t	375	375	0.3	Existing	1	G20		0		
D2_P3	G13	G14	41.205	4.877	4.683	0.47	Concrete, t	375	375	0.3	New	1	G13		0		
D2_P2	G14	G17	24.4263	4.686	4.568	0.48	Concrete, t	375	375	0.3	Existing	1	G14		0		
D2_P1	G17	Basin3	24.4452	4.508	4.347	0.66	Concrete, t	375	375	0.3	Existing	1	G17		0		

APPENDIX G: DRAINS Model



APPENDIX H: DRAINS SURFACES



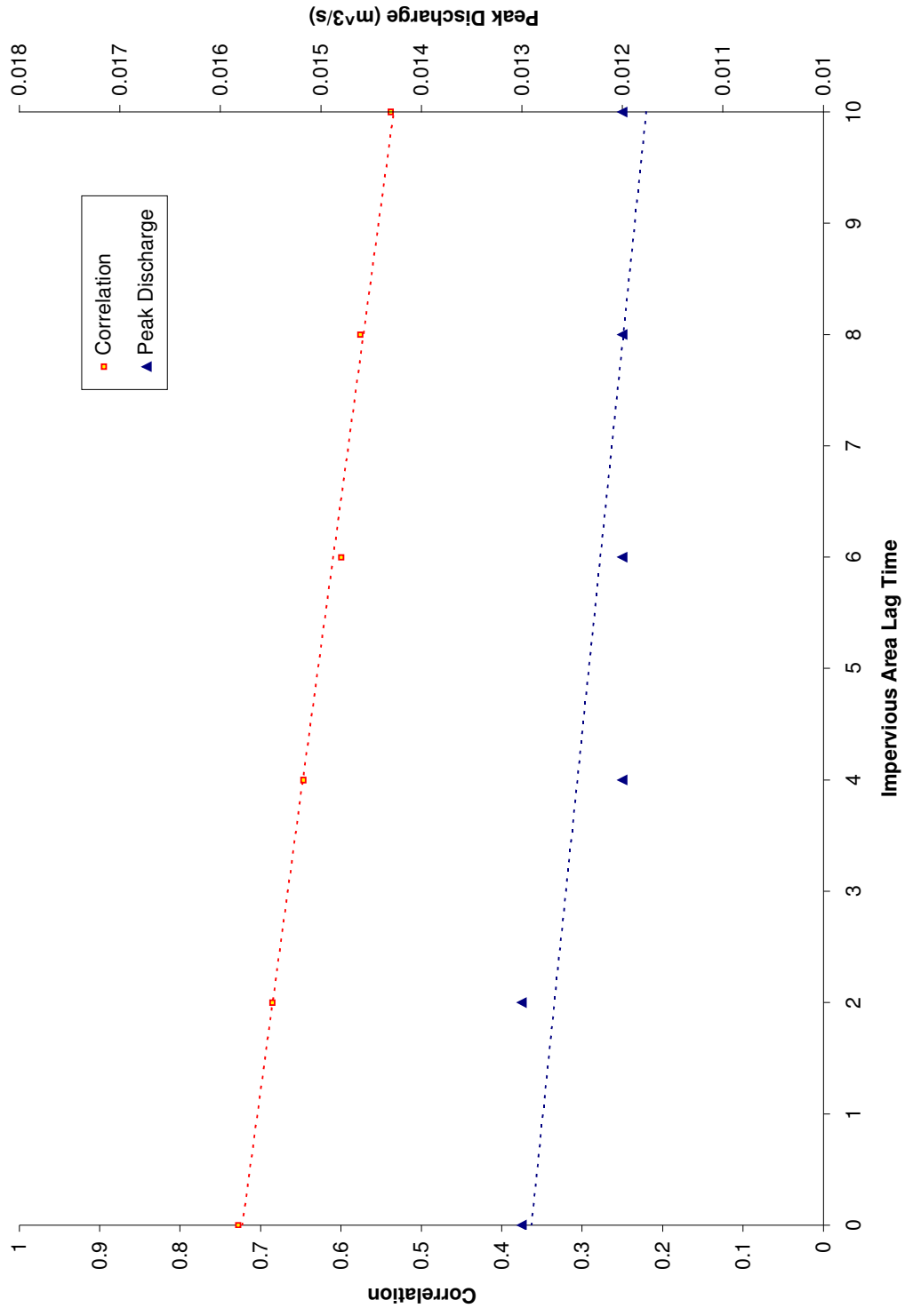
APPENDIX I: Sensitivity Analysis Summary

Paved				Supplementry				Grassed				Pipes		Soil	Results									
Additional Time	Flow Path Slope	Flow Path Roughness	Area Depression Storage	Additional Time	Flow Path Slope	Flow Path Roughness	Area Depression Storage	Lag Time	Additional Time	Flow Path Slope	Flow Path Roughness	Area Depression Storage	Pressure Loss Coefficient	Soil Type	Total Volume	Change in Total Volume	Peak Discharge	Change in Peak Discharge	TS2	TS2 Correlation	Time to Peak	Tp Correlation	Correlation (shape)	Change in R2
min	%		mm	min	%		mm	min	min	%		mm			m³	%	m³/s	%					R	%
0	1	0.015	1.5	5	1.7	0.015	1	10	15	1.7	0.13	3	4	4	22.7	-	0.013	-	115	7	142	-10	0.7275622	
2	-	-	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	115	7	144.5	-12.5	0.684776	-5.88 %
4	-	-	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.012	-7.69 %	115	7	144.5	-12.5	0.6463182	-11.17 %
6	-	-	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.012	-7.69 %	115	7	145	-13	0.5993776	-17.62 %
8	-	-	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.012	-7.69 %	116	6	145.5	-13.5	0.5759351	-20.84 %
10	-	-	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.012	-7.69 %	116	6	145	-13	0.5378972	-26.07 %
-	1.5	-	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	115	7	140.5	-8.5	0.7838455	7.74 %
-	2	-	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	114	8	139.5	-7.5	0.8189314	12.56 %
-	2.5	-	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	114	8	138.5	-6.5	0.8391271	15.33 %
-	0.5	-	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.012	-7.69 %	115	7	144.5	-12.5	0.6246081	-14.15 %
-	0.1	-	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.01	-23.08 %	117	5	149.5	-17.5	0.4054244	-44.28 %
-	-	0.03	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.012	-7.69 %	116	6	145.5	-13.5	0.5243946	-27.92 %
-	-	0.045	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.01	-23.08 %	117	5	149	-17	0.4203241	-42.23 %
-	-	0.08	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.008	-38.46 %	118	4	150.5	-18.5	0.314743	-56.74 %
-	-	0.02	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.012	-7.69 %	115	7	144.5	-12.5	0.6429902	-11.62 %
-	-	0.01	-	-	-	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	114	8	139.5	-7.5	0.8251691	13.42 %
-	-	-	2	-	-	-	-	-	-	-	-	-	-	-	22.2	-2.20 %	0.013	0.00 %	115	7	142	-10	0.7275622	0.00 %
-	-	-	2.5	-	-	-	-	-	-	-	-	-	-	-	21.7	-4.41 %	0.013	0.00 %	116	6	142	-10	0.719969	-1.04 %
-	-	-	1	-	-	-	-	-	-	-	-	-	-	-	23.2	2.20 %	0.013	0.00 %	115	7	142	-10	0.7256699	-0.26 %
-	-	-	0.5	-	-	-	-	-	-	-	-	-	-	-	23.7	4.41 %	0.013	0.00 %	115	7	142	-10	0.725486	-0.29 %
-	-	-	0.1	-	-	-	-	-	-	-	-	-	-	-	24.1	6.17 %	0.013	0.00 %	115	7	142	-10	0.7253151	-0.31 %
-	-	-	-	10	-	-	-	-	-	-	-	-	-	-	22.6	-0.44 %	0.013	0.00 %	115	7	142.5	-10.5	0.7374537	1.36 %
-	-	-	-	15	-	-	-	-	-	-	-	-	-	-	22.3	-1.76 %	0.012	-7.69 %	115	7	142.5	-10.5	0.7376376	1.38 %
-	-	-	-	20	-	-	-	-	-	-	-	-	-	-	22	-3.08 %	0.012	-7.69 %	115	7	142	-10	0.7486449	2.90 %
-	-	-	-	25	-	-	-	-	-	-	-	-	-	-	21.6	-4.85 %	0.012	-7.69 %	115	7	141.5	-9.5	0.7564987	3.98 %
-	-	-	-	0	-	-	-	-	-	-	-	-	-	-	22.9	0.88 %	0.013	0.00 %	115	7	142	-10	0.7254611	-0.29 %
-	-	-	-	-	2.5	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142.5	-10.5	0.7255453	-0.28 %
-	-	-	-	-	3	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142.5	-10.5	0.7255453	-0.28 %
-	-	-	-	-	1	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142	-10	0.7275622	0.00 %
-	-	-	-	-	0.5	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142	-10	0.7275622	0.00 %
-	-	-	-	-	0.1	-	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142	-10	0.7296549	0.29 %
-	-	-	-	-	-	0.03	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142	-10	0.7275622	0.00 %
-	-	-	-	-	-	0.045	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142	-10	0.727203	-0.05 %
-	-	-	-	-	-	0.08	-	-	-	-	-	-	-	-	22.6	-0.44 %	0.013	0.00 %	115	7	142	-10	0.7296549	0.29 %
-	-	-	-	-	-	0.02	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142	-10	0.7275622	0.00 %
-	-	-	-	-	-	0.01	-	-	-	-	-	-	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142.5	-10.5	0.7255453	-0.28 %

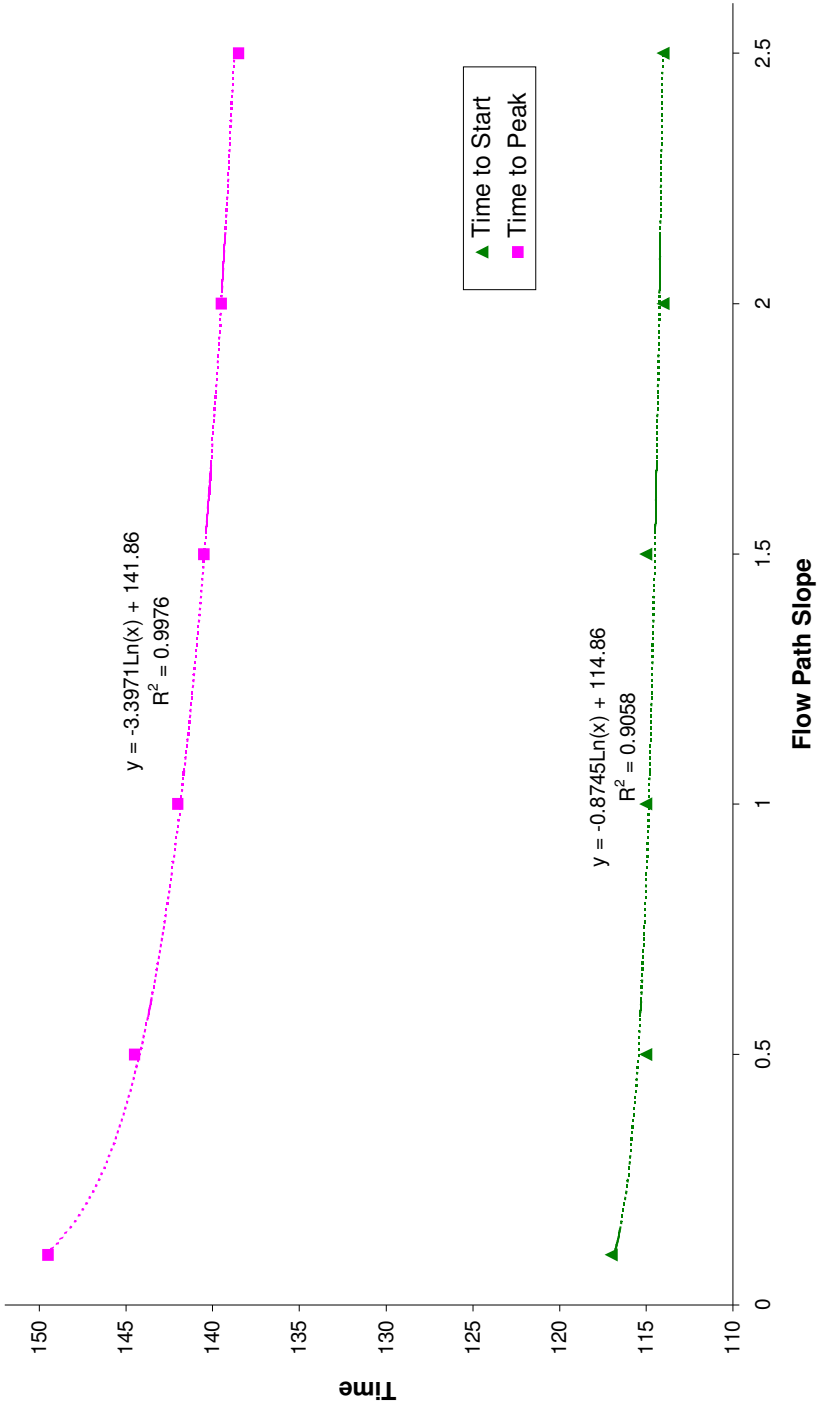
Paved				Results														Pipes		Soil	Change in R2					
Additional Time	Flow Path Slope	Flow Path Roughness	Area Depression Storage	Supplementary		Flow Path Slope	Flow Path Roughness	Area Depression Storage	Lag Time	Grassed		Flow Path Slope	Flow Path Roughness	Area Depression Storage	Pressure Loss Coefficient	Soil Type	Total Volume	Change in Total Volume	Peak Discharge	Change in Peak Discharge	TS2	TS2 Correlation	Time to Peak	Tp Correlation	Correlation (shape)	Change in R2
				Additional Time	min					min	%															
0	1	0.015	1.5	5	1.7	0.015	1	10	15	1.7	0.13	3	4	4	4	4	22.7	-	0.013	-	115	7	142	-10	0.7275622	
-	-	-	-	-	-	-	1.5	-	-	-	-	-	-	-	-	-	22.6	-0.44 %	0.013	0.00 %	115	7	142	-10	0.7292119	0.23 %
-	-	-	-	-	-	-	2	-	-	-	-	-	-	-	-	-	22.5	-0.88 %	0.013	0.00 %	115	7	142	-10	0.7326742	0.70 %
-	-	-	-	-	-	-	2.5	-	-	-	-	-	-	-	-	-	22.4	-1.32 %	0.013	0.00 %	115	7	142	-10	0.7327787	0.72 %
-	-	-	-	-	-	-	0.5	-	-	-	-	-	-	-	-	-	22.8	0.44 %	0.013	0.00 %	115	7	142.5	-10.5	0.7251891	-0.33 %
-	-	-	-	-	-	-	0.1	-	-	-	-	-	-	-	-	-	22.9	0.88 %	0.013	0.00 %	115	7	142.5	-10.5	0.7251891	-0.33 %
-	-	-	-	-	-	-	-	15	-	-	-	-	-	-	-	-	22.7	0.00 %	0.011	-15.38 %	115	7	135.5	-3.5	0.7366714	1.24 %
-	-	-	-	-	-	-	-	20	-	-	-	-	-	-	-	-	22.7	0.00 %	0.011	-15.38 %	115	7	135.5	-3.5	0.7691533	5.72 %
-	-	-	-	-	-	-	-	25	-	-	-	-	-	-	-	-	22.7	0.00 %	0.011	-15.38 %	115	7	135.5	-3.5	0.7897354	8.55 %
-	-	-	-	-	-	-	-	5	-	-	-	-	-	-	-	-	22.7	0.00 %	0.014	7.69 %	115	7	141	-9	0.7582141	4.21 %
-	-	-	-	-	-	-	-	0	-	-	-	-	-	-	-	-	22.7	0.00 %	0.014	7.69 %	115	7	138	-6	0.8184187	12.49 %
-	-	-	-	-	-	-	-	-	25	-	-	-	-	-	-	-	21.7	-4.41 %	0.012	-7.69 %	115	7	142	-10	0.7536821	3.59 %
-	-	-	-	-	-	-	-	-	35	-	-	-	-	-	-	-	20.9	-7.93 %	0.012	-7.69 %	115	7	141.5	-9.5	0.769232	5.73 %
-	-	-	-	-	-	-	-	-	45	-	-	-	-	-	-	-	20.3	-10.57 %	0.012	-7.69 %	115	7	142	-10	0.7789695	7.07 %
-	-	-	-	-	-	-	-	-	10	-	-	-	-	-	-	-	23.2	2.20 %	0.013	0.00 %	115	7	142.5	-10.5	0.7104711	-2.35 %
-	-	-	-	-	-	-	-	-	2	-	-	-	-	-	-	-	24.1	6.17 %	0.014	7.69 %	115	7	143.5	-11.5	0.6750113	-7.22 %
-	-	-	-	-	-	-	-	-	-	2.5	-	-	-	-	-	-	23	1.32 %	0.013	0.00 %	115	7	142.5	-10.5	0.7201928	-1.01 %
-	-	-	-	-	-	-	-	-	-	3	-	-	-	-	-	-	23.1	1.76 %	0.013	0.00 %	115	7	143	-11	0.7127013	-2.04 %
-	-	-	-	-	-	-	-	-	-	1	-	-	-	-	-	-	22.2	-2.20 %	0.013	0.00 %	115	7	142.5	-10.5	0.7391966	1.60 %
-	-	-	-	-	-	-	-	-	-	0.5	-	-	-	-	-	-	21.6	-4.85 %	0.012	-7.69 %	115	7	142	-10	0.7569072	4.03 %
-	-	-	-	-	-	-	-	-	-	0.1	-	-	-	-	-	-	20.1	-11.45 %	0.012	-7.69 %	115	7	141.5	-9.5	0.7883488	8.35 %
-	-	-	-	-	-	-	-	-	-	-	0.2	-	-	-	-	-	21.9	-3.52 %	0.012	-7.69 %	115	7	142.5	-10.5	0.7480827	2.82 %
-	-	-	-	-	-	-	-	-	-	-	0.3	-	-	-	-	-	21.1	-7.05 %	0.012	-7.69 %	115	7	142	-10	0.7641996	5.04 %
-	-	-	-	-	-	-	-	-	-	-	0.4	-	-	-	-	-	20.6	-9.25 %	0.012	-7.69 %	115	7	141.5	-9.5	0.7749155	6.51 %
-	-	-	-	-	-	-	-	-	-	-	0.075	-	-	-	-	-	23.5	3.52 %	0.013	0.00 %	115	7	143	-11	0.7034288	-3.32 %
-	-	-	-	-	-	-	-	-	-	-	0.01	-	-	-	-	-	24.9	9.69 %	0.016	23.08 %	115	7	144.5	-12.5	0.6301682	-13.39 %
-	-	-	-	-	-	-	-	-	-	-	-	4	-	-	-	-	21.2	-6.61 %	0.012	-7.69 %	115	7	142.5	-10.5	0.7469658	2.67 %
-	-	-	-	-	-	-	-	-	-	-	-	5	-	-	-	-	19.9	-12.33 %	0.012	-7.69 %	115	7	142	-10	0.7676475	5.51 %
-	-	-	-	-	-	-	-	-	-	-	-	2	-	-	-	-	24.2	6.61 %	0.013	0.00 %	115	7	142.5	-10.5	0.7067909	-2.85 %
-	-	-	-	-	-	-	-	-	-	-	-	1	-	-	-	-	25.7	13.22 %	0.014	7.69 %	115	7	142.5	-10.5	0.6903463	-5.12 %
-	-	-	-	-	-	-	-	-	-	-	-	0.1	-	-	-	-	27	18.94 %	0.015	15.38 %	115	7	142	-10	0.6750294	-7.22 %
-	-	-	-	-	-	-	-	-	-	-	-	-	-	6	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142.5	-10.5	0.7254434	-0.29 %
-	-	-	-	-	-	-	-	-	-	-	-	-	-	8	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142.5	-10.5	0.7249873	-0.35 %
-	-	-	-	-	-	-	-	-	-	-	-	-	-	10	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142.5	-10.5	0.7253421	-0.31 %
-	-	-	-	-	-	-	-	-	-	-	-	-	-	2	-	-	22.7	0.00 %	0.013	0.00 %	115	7	142	-10	0.7275622	0.00 %
-	-	-	-	-	-	-	-	-	-	-	-	-	-	0.5	-	-	22.7	0.00 %	0.013	0.00 %	114	8	142	-10	0.7349656	1.02 %
-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	3	16.9	-25.55 %	0.011	-15.38 %	115	7	135.5	-3.5	0.8509259	16.96 %
-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	2	16.9	-25.55 %	0.011	-15.38 %	115	7	135.5	-3.5	0.8509259	16.96 %
-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	1	16.9	-25.55 %	0.011	-15.38 %	115	7	135.5	-3.5	0.8509259	16.96 %

APPENDIX J: Sensitivity Analysis Plots

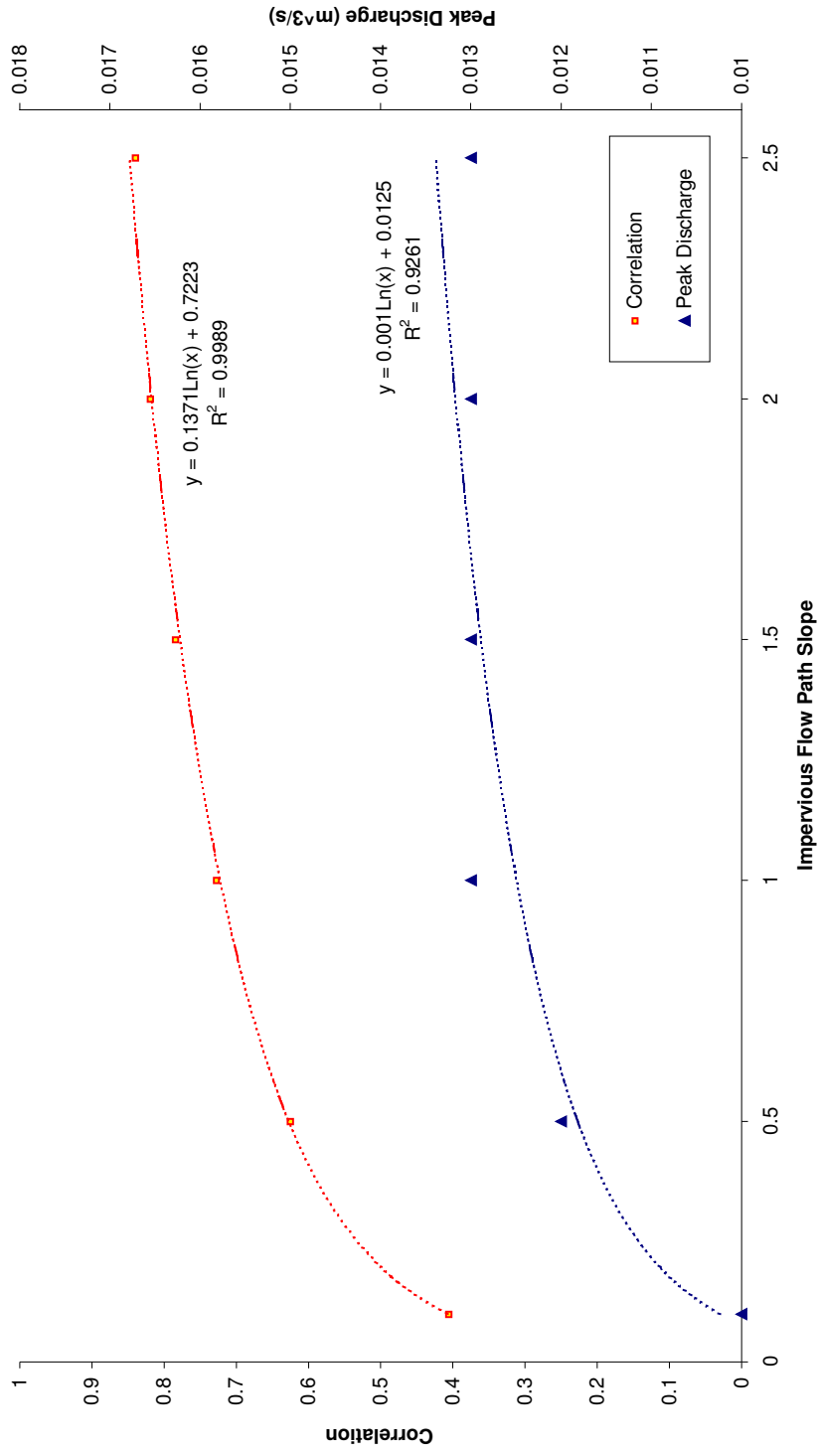
Sensitivity of Impervious Area Lag Time on Model Output



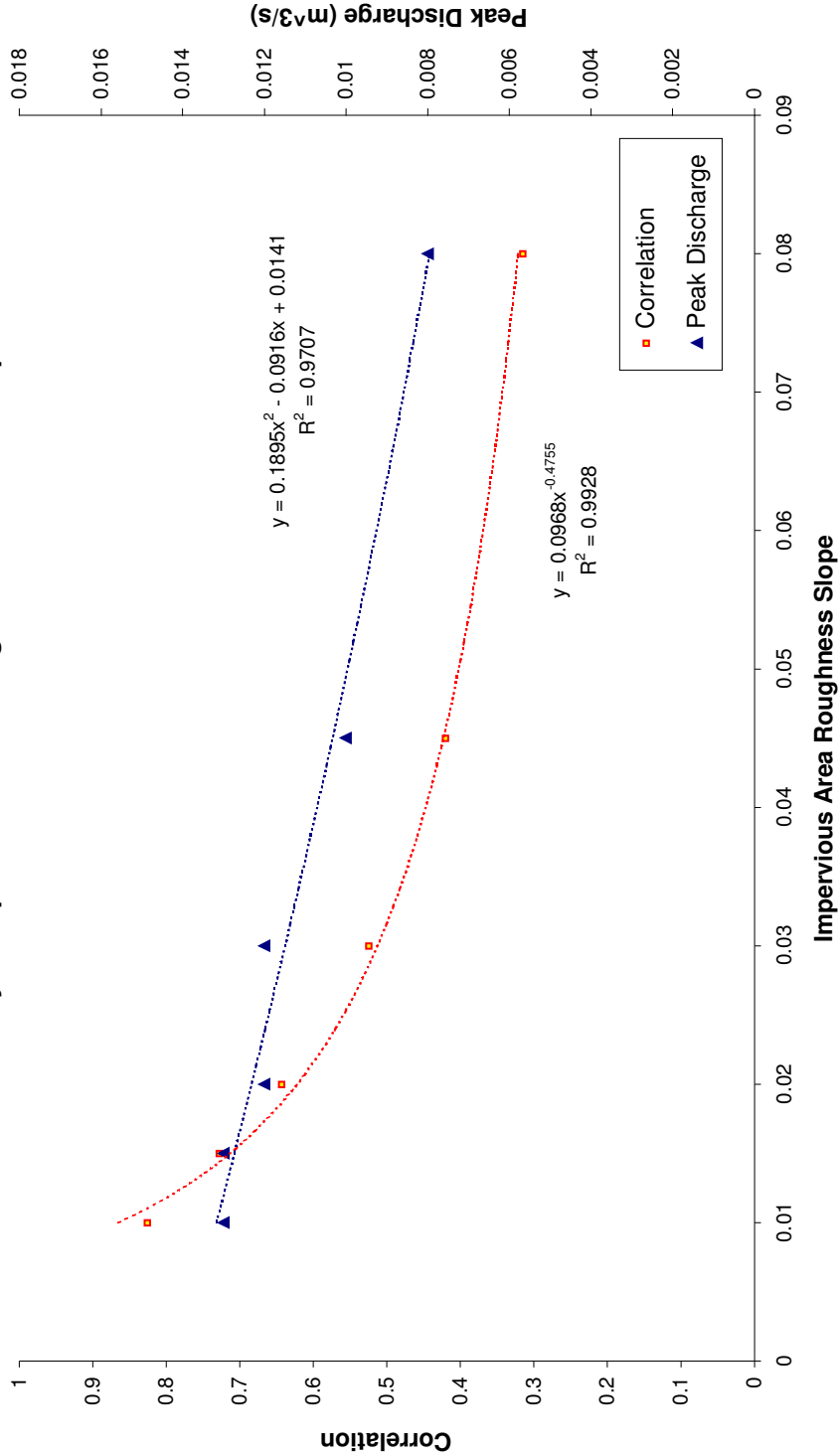
Sensitivity of Impervious Area Flow Path Slope on Model Output



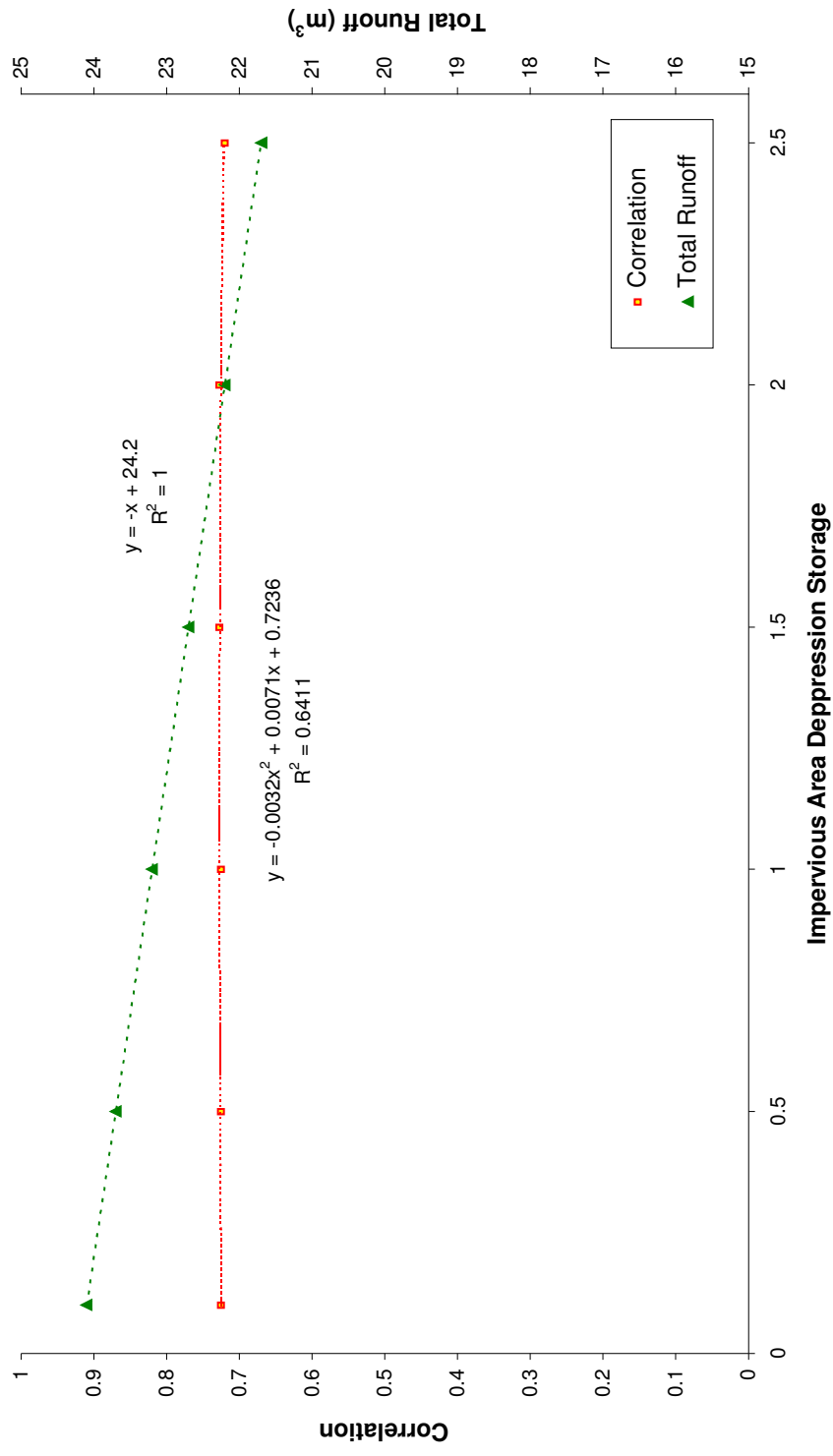
Sensitivity of Impervious Flow Path Slope on Model Output



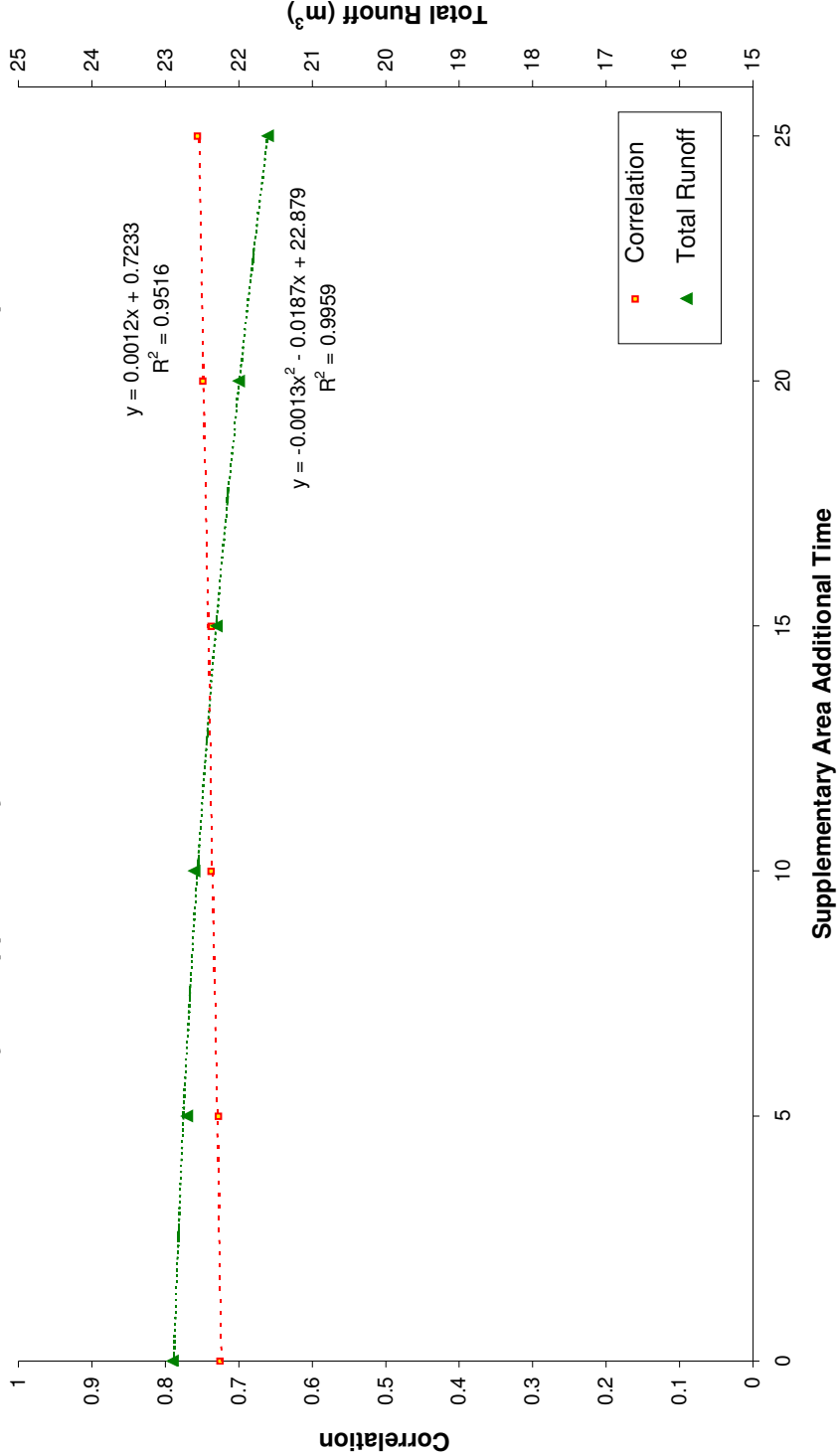
Sensitivity of Impervious Area Roughness on Model Output



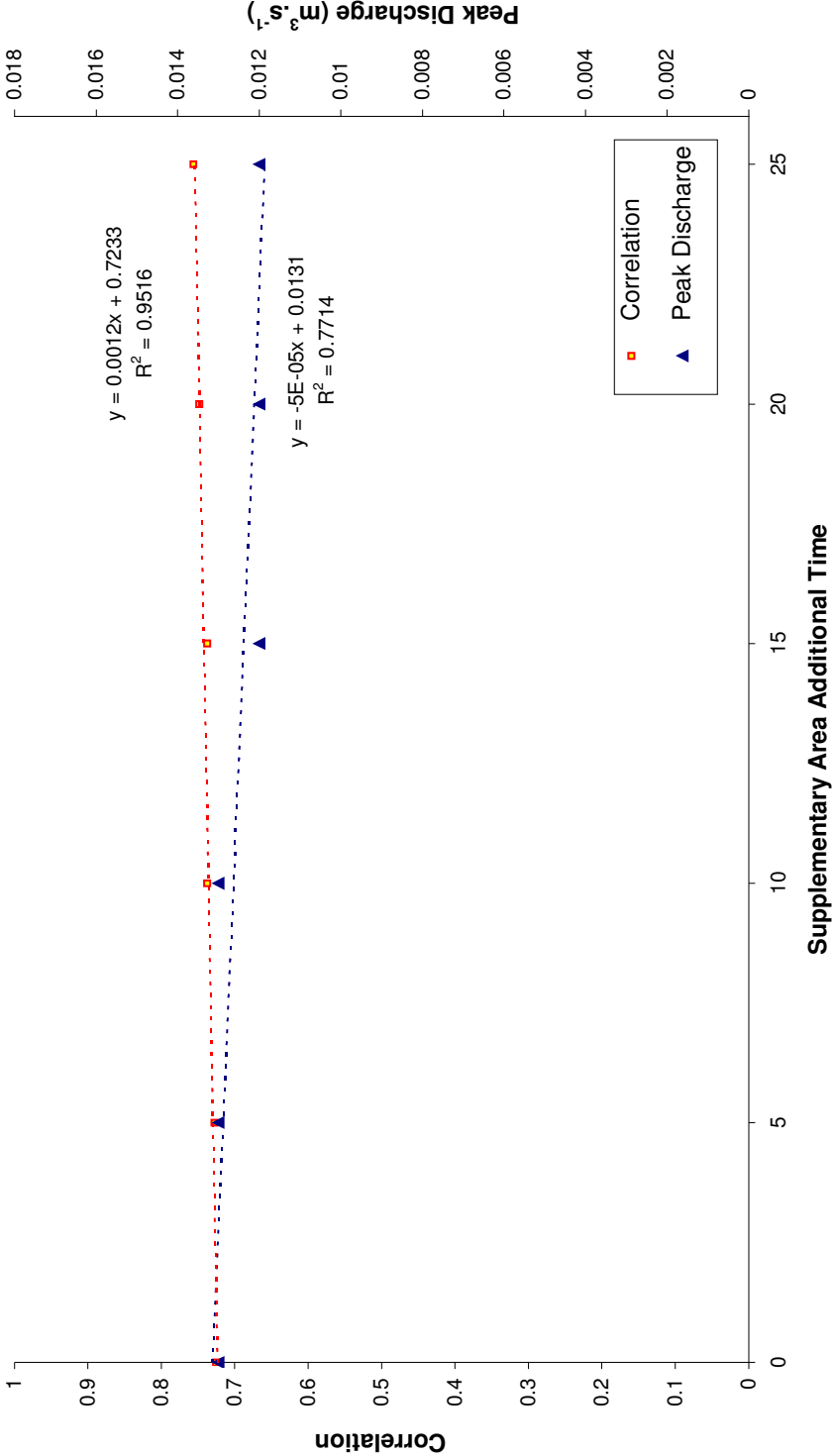
Sensitivity of Impervious Area Depression Storage on Model Output



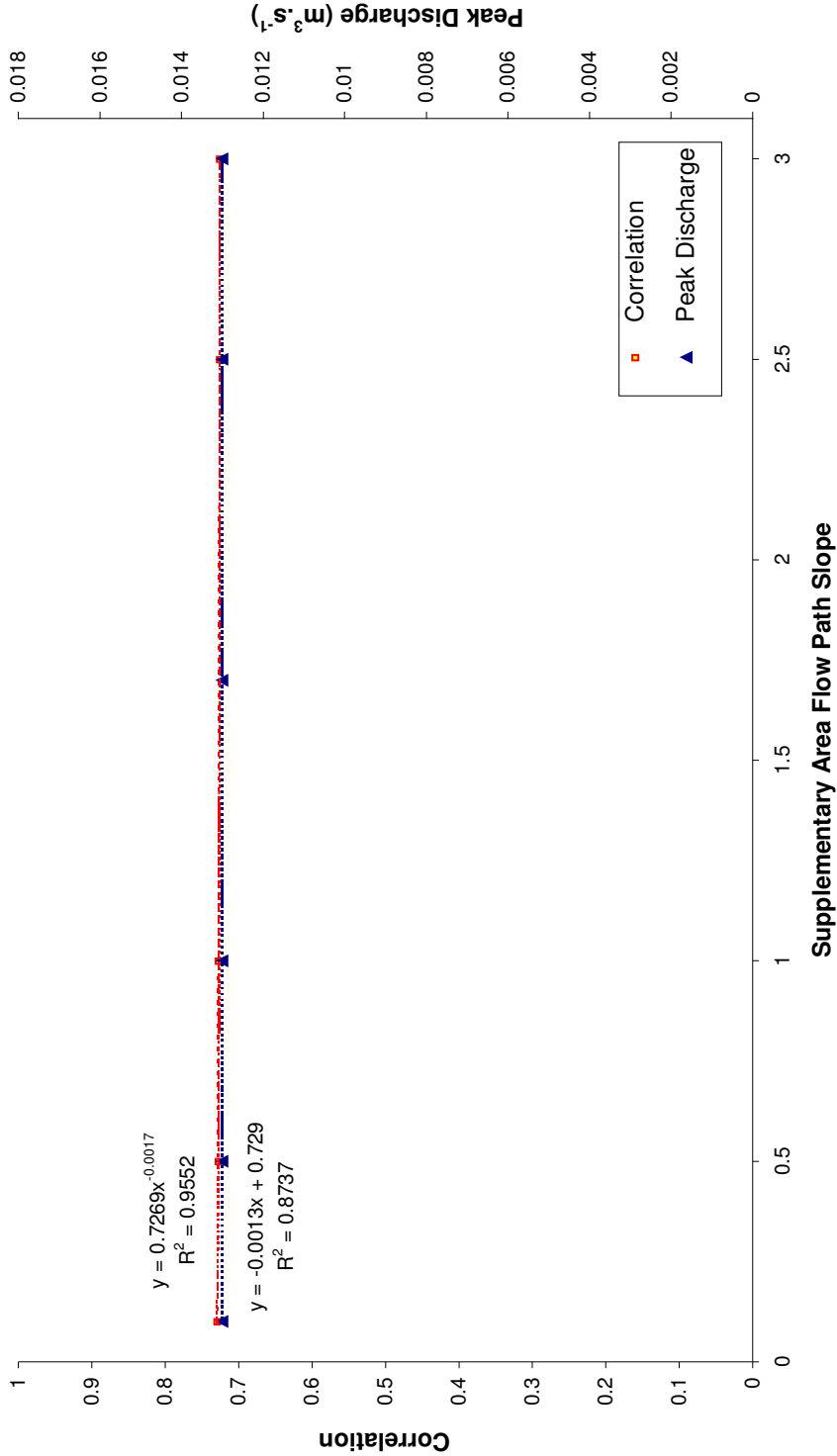
Sensitivity of Supplementary Area Additional Time on Model Output



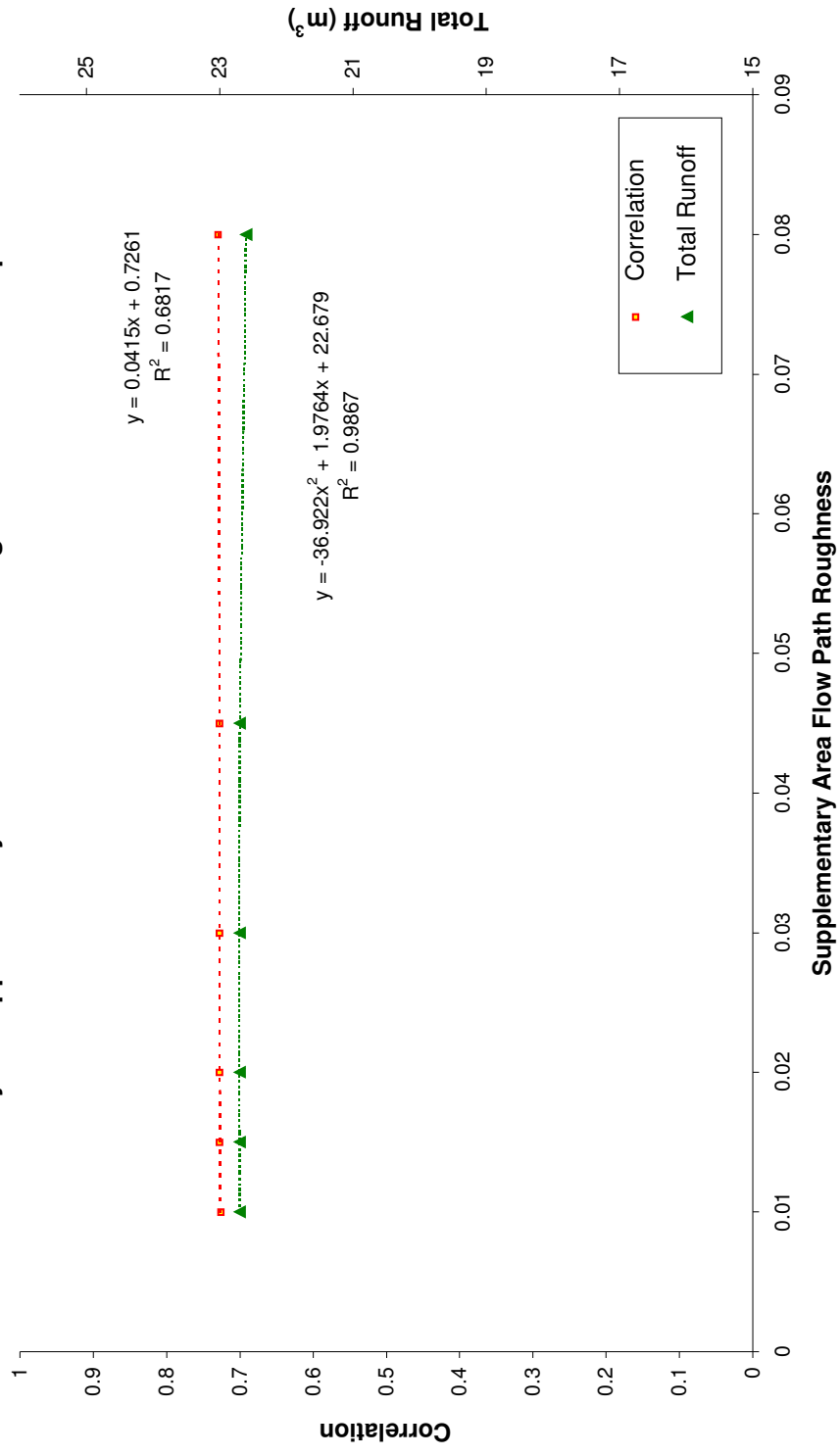
Sensitivity of Supplementary Area Additional Time on Model Output



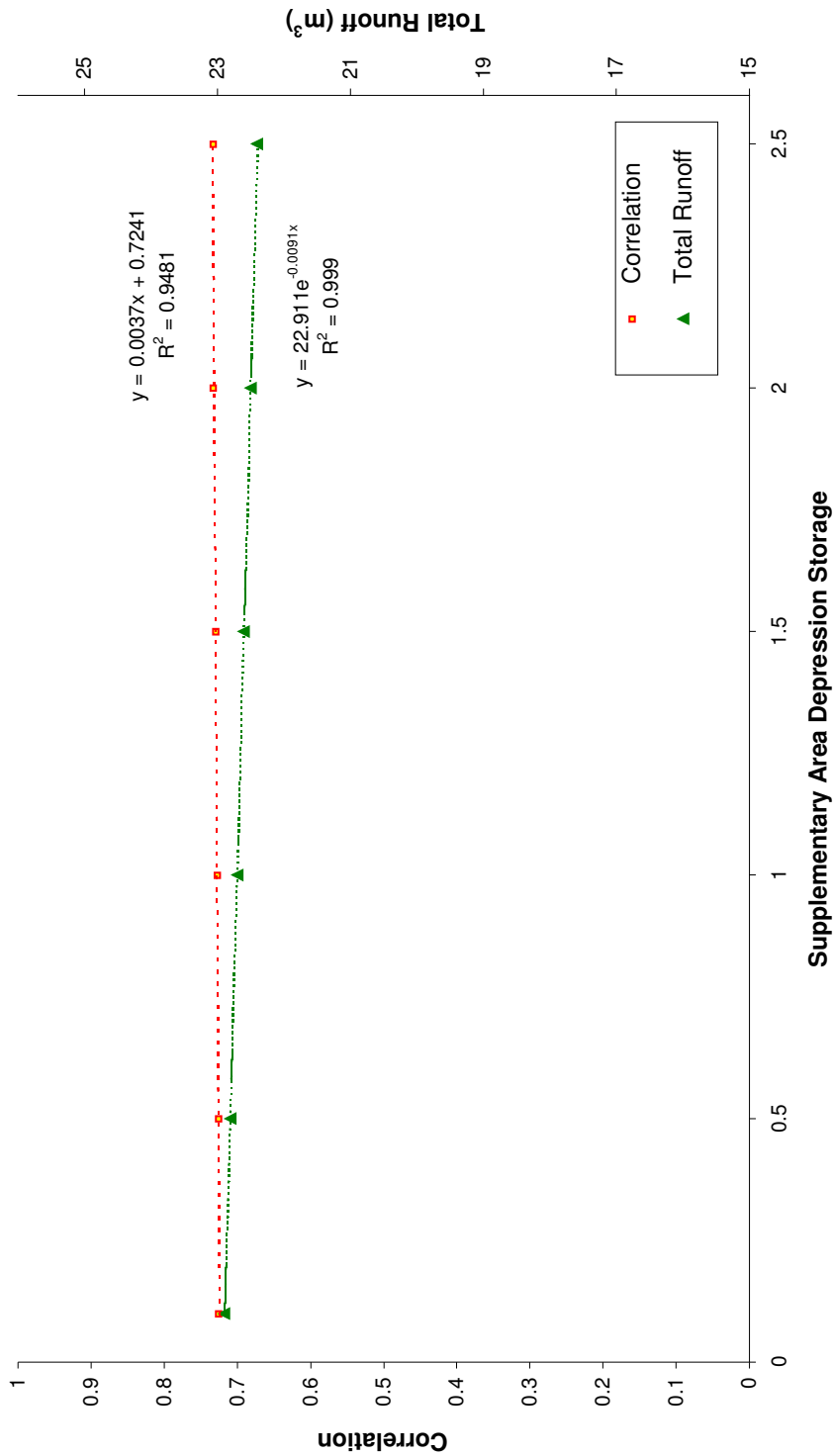
Sensitivity of Supplementary Area Flow Path Slope on Model Output



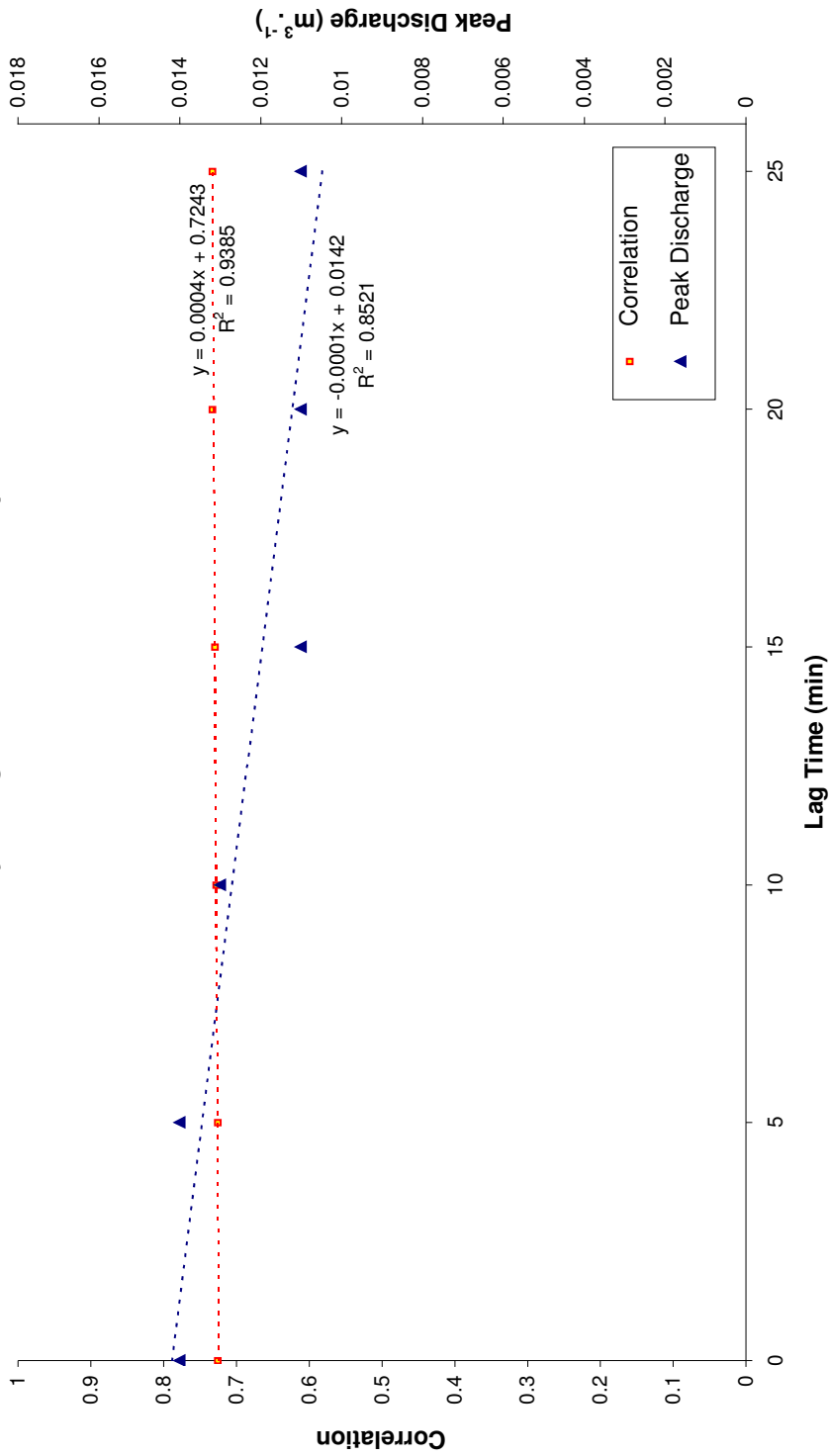
Sensitivity of Supplementary Area Flow Path Roughness on Model Output



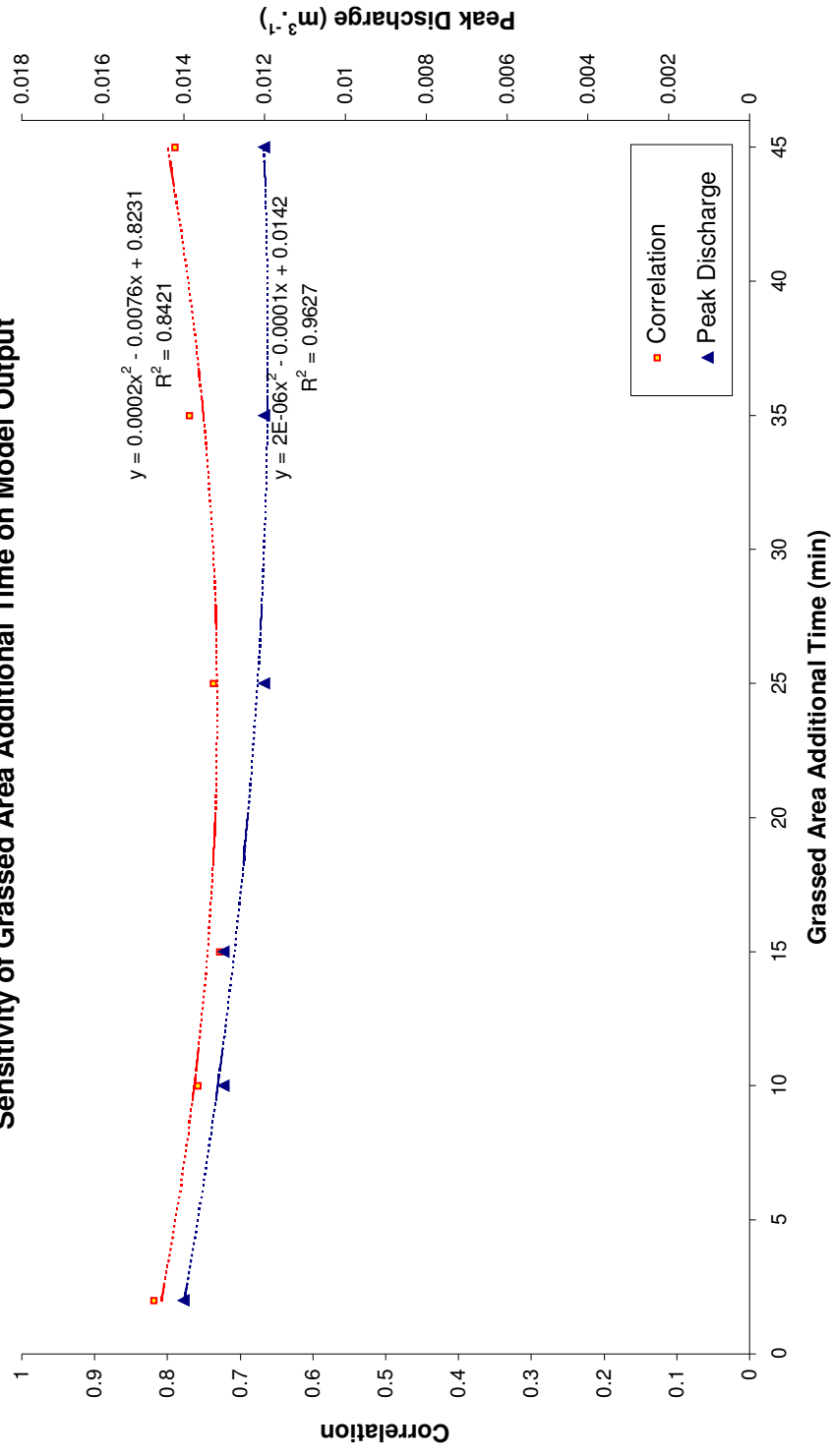
Sensitivity of Supplementary Area Depression Storage on Model Output



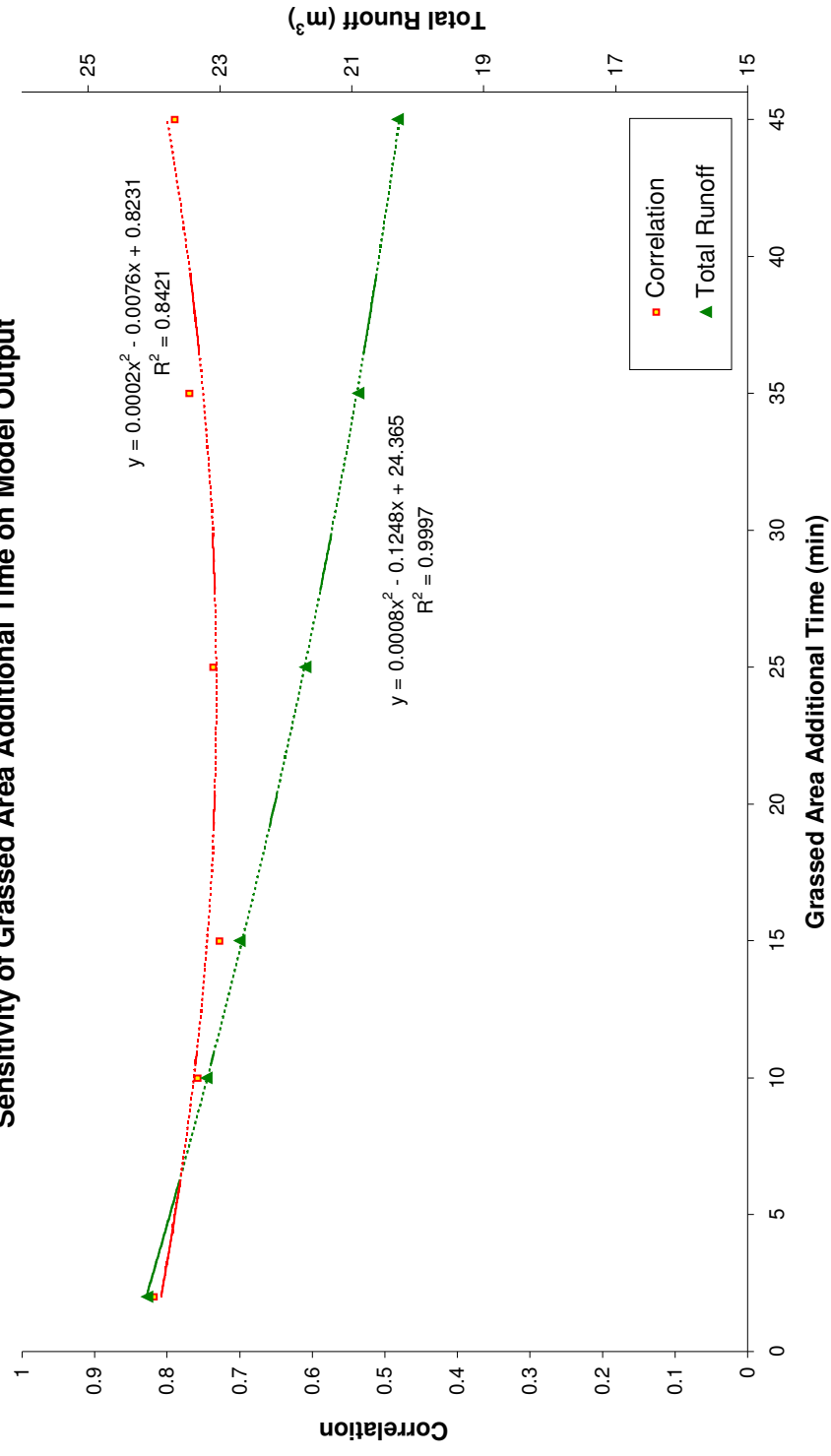
Sensitivity of Lag Time on Model Output



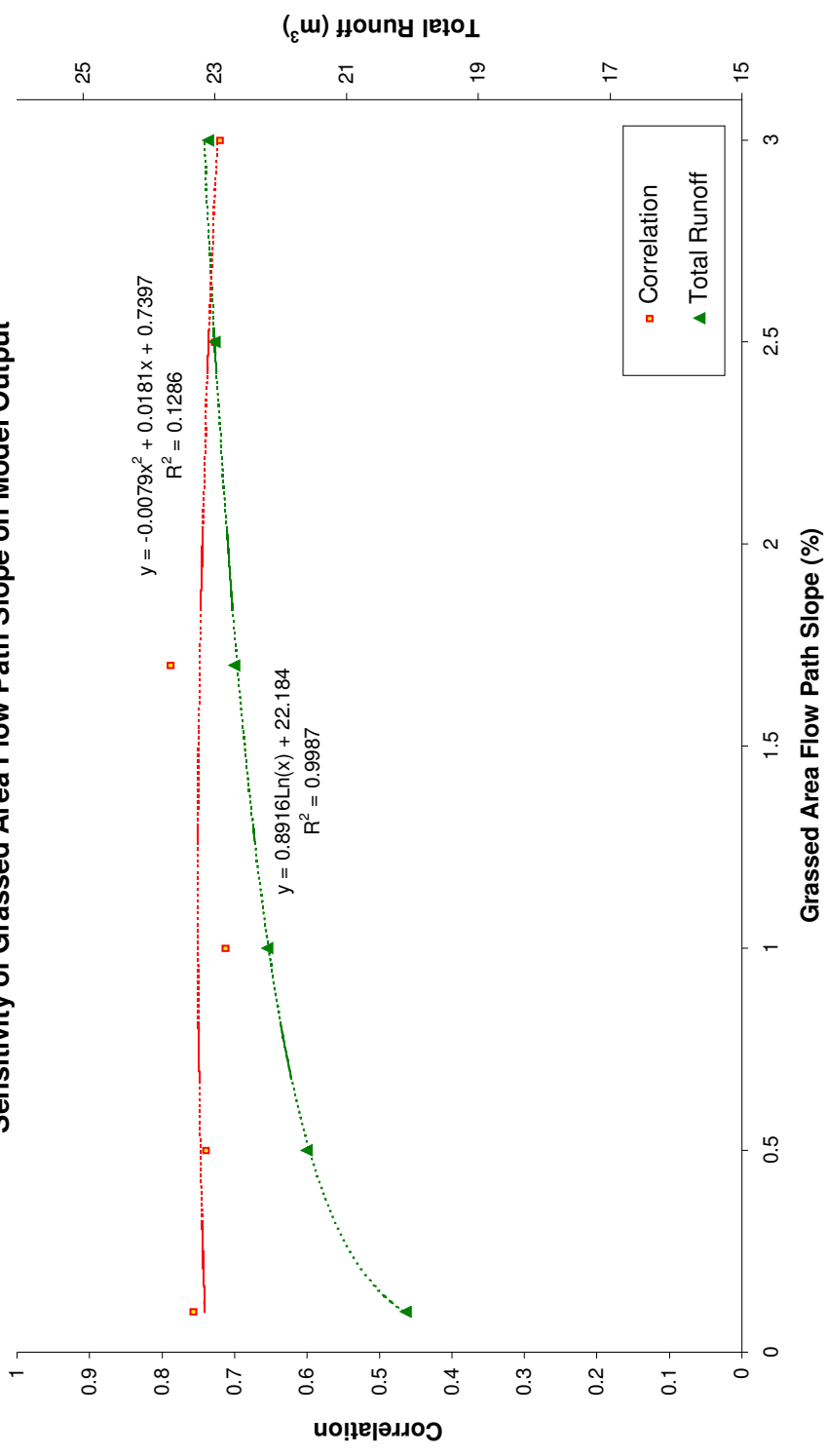
Sensitivity of Grassed Area Additional Time on Model Output



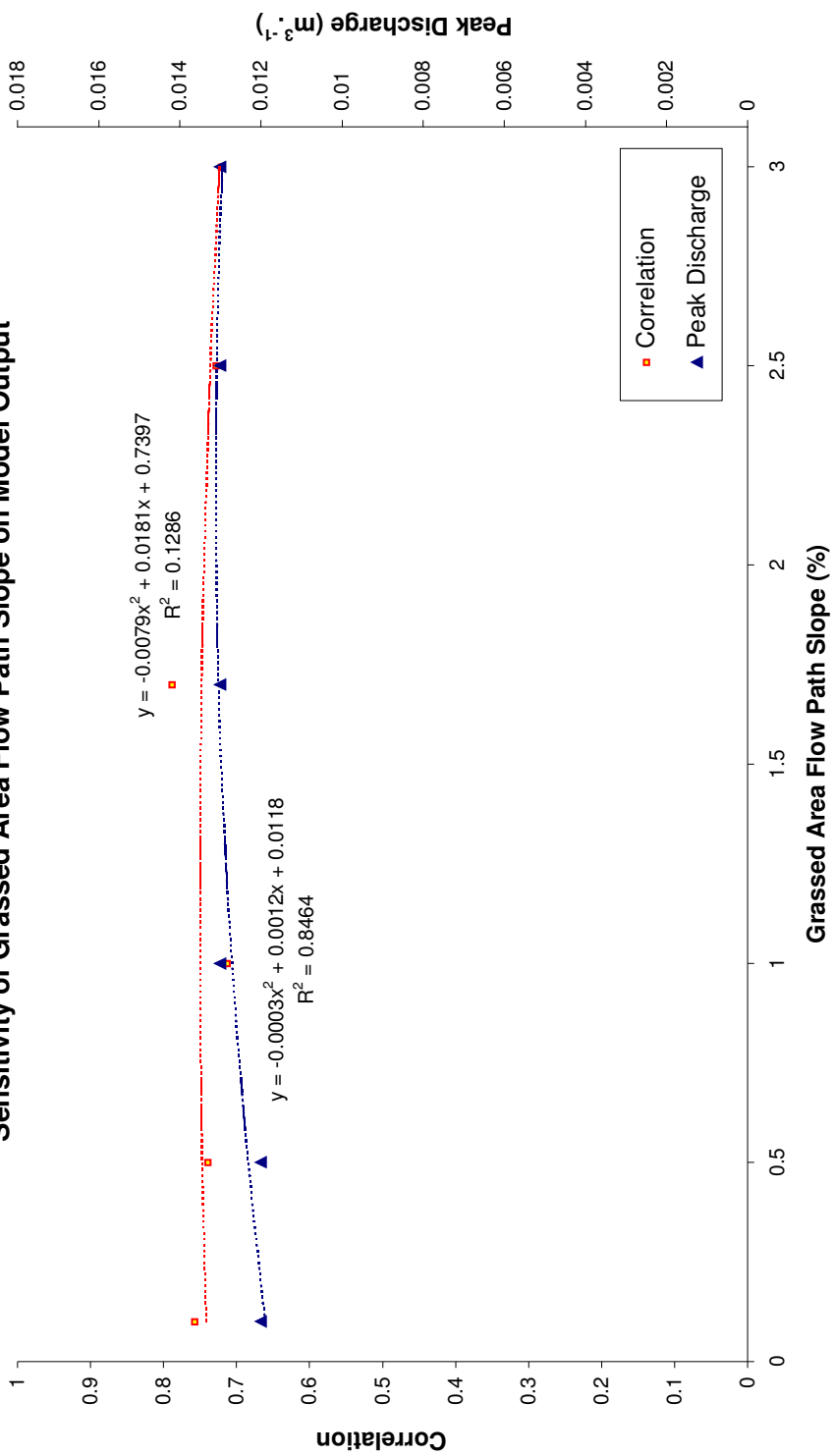
Sensitivity of Grassed Area Additional Time on Model Output



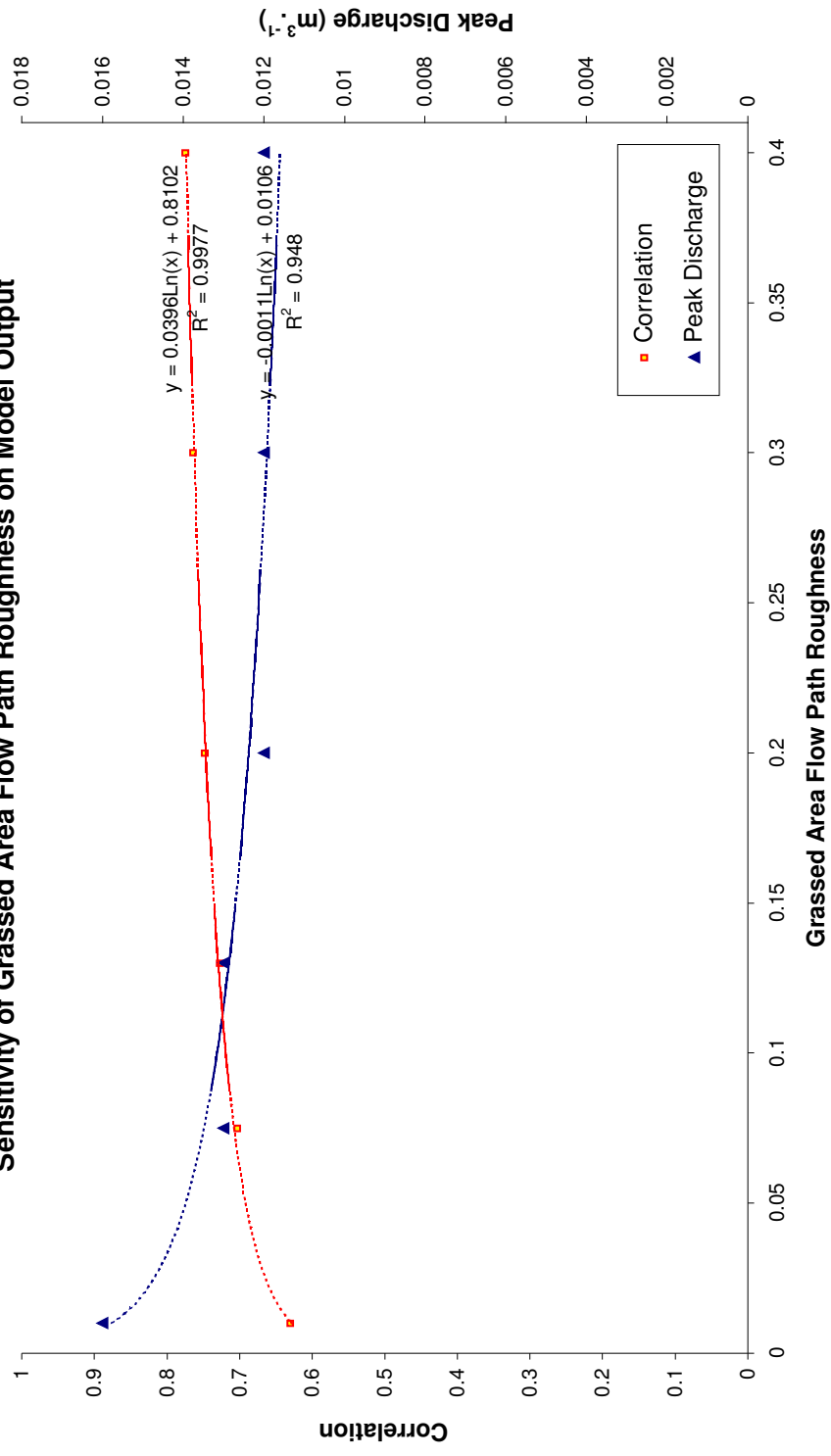
Sensitivity of Grassed Area Flow Path Slope on Model Output



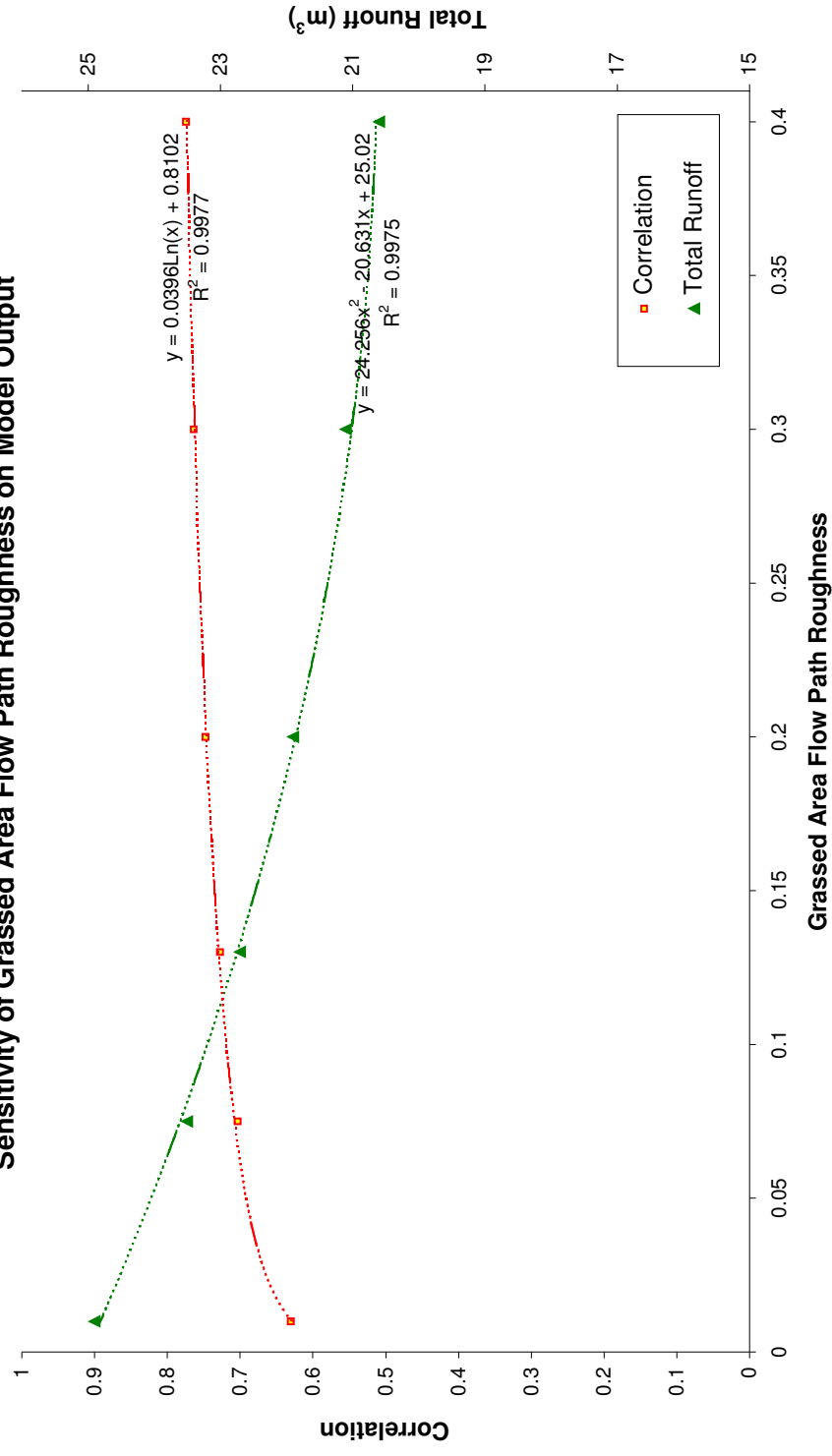
Sensitivity of Grassed Area Flow Path Slope on Model Output



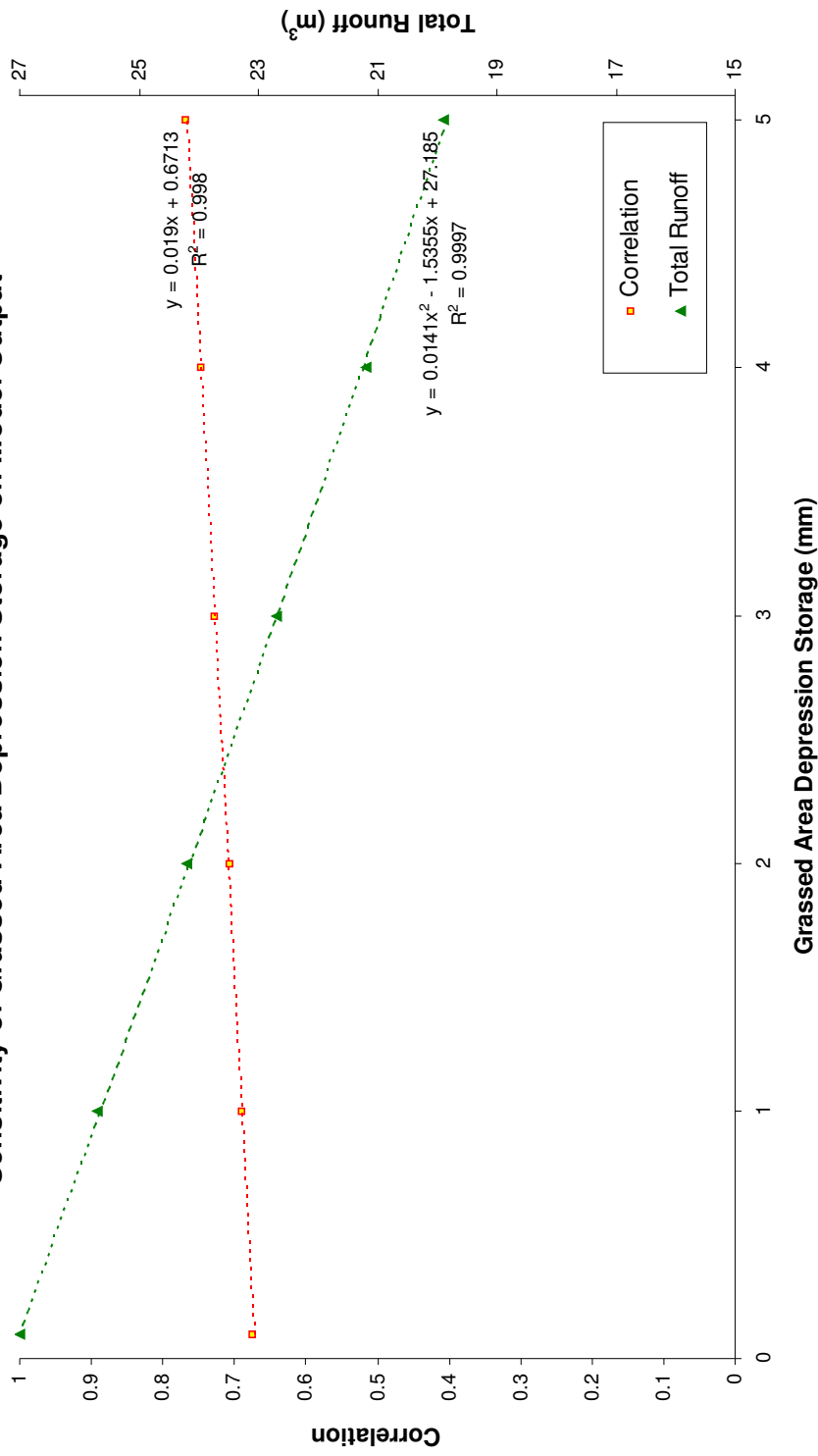
Sensitivity of Grassed Area Flow Path Roughness on Model Output



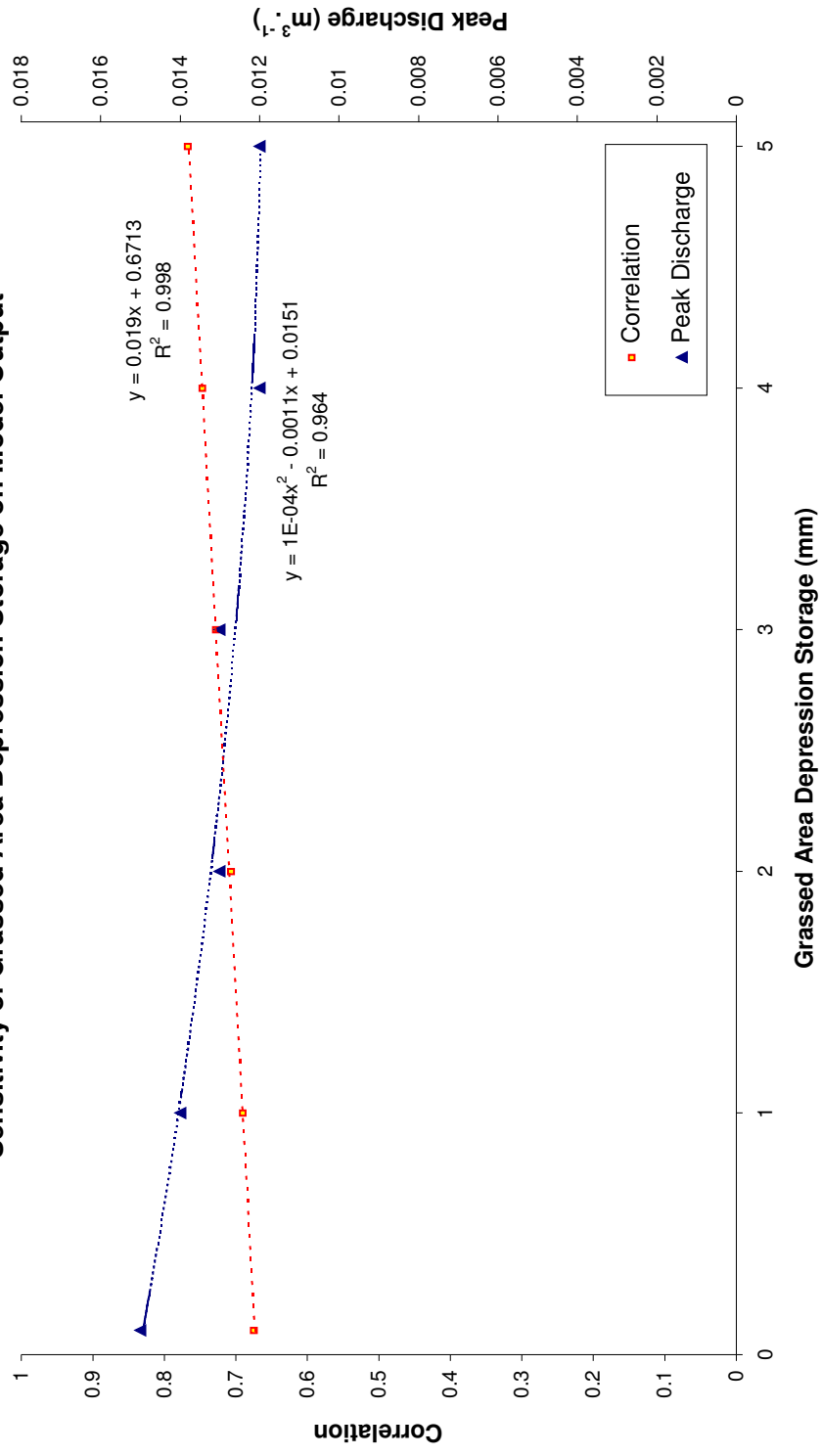
Sensitivity of Grassed Area Flow Path Roughness on Model Output



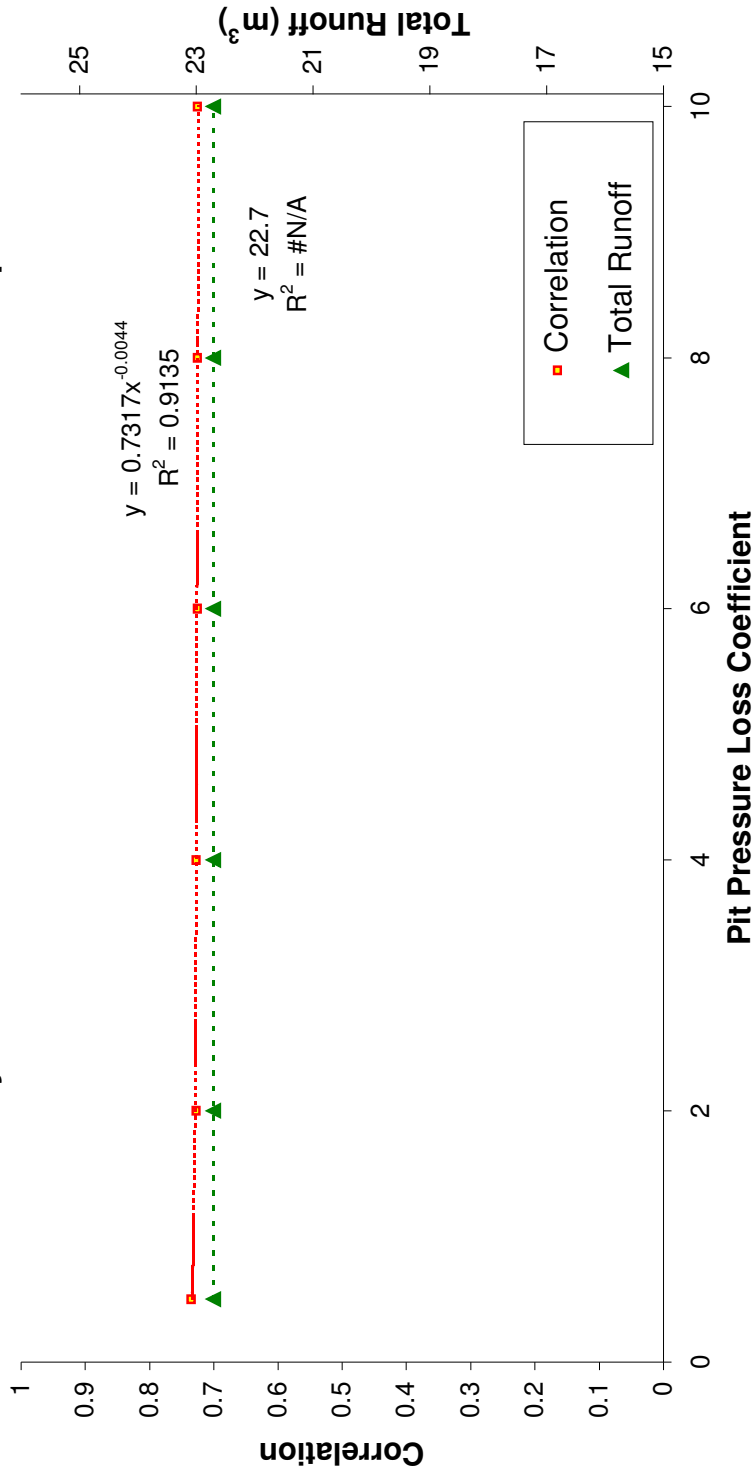
Sensitivity of Grassed Area Depression Storage on Model Output



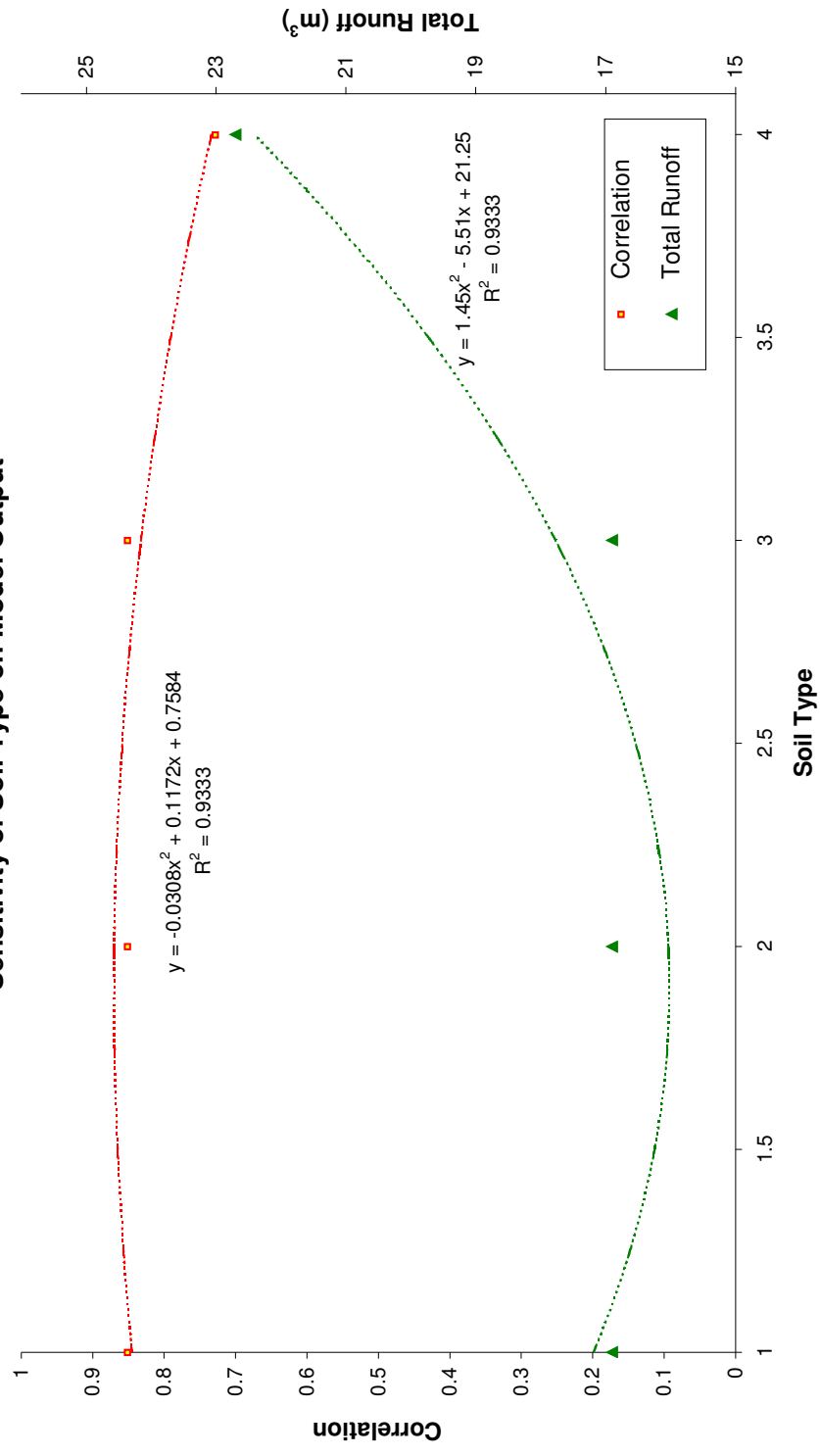
Sensitivity of Grassed Area Depression Storage on Model Output



Sensitivity of Pit Pressure Loss Coefficient on Model Output

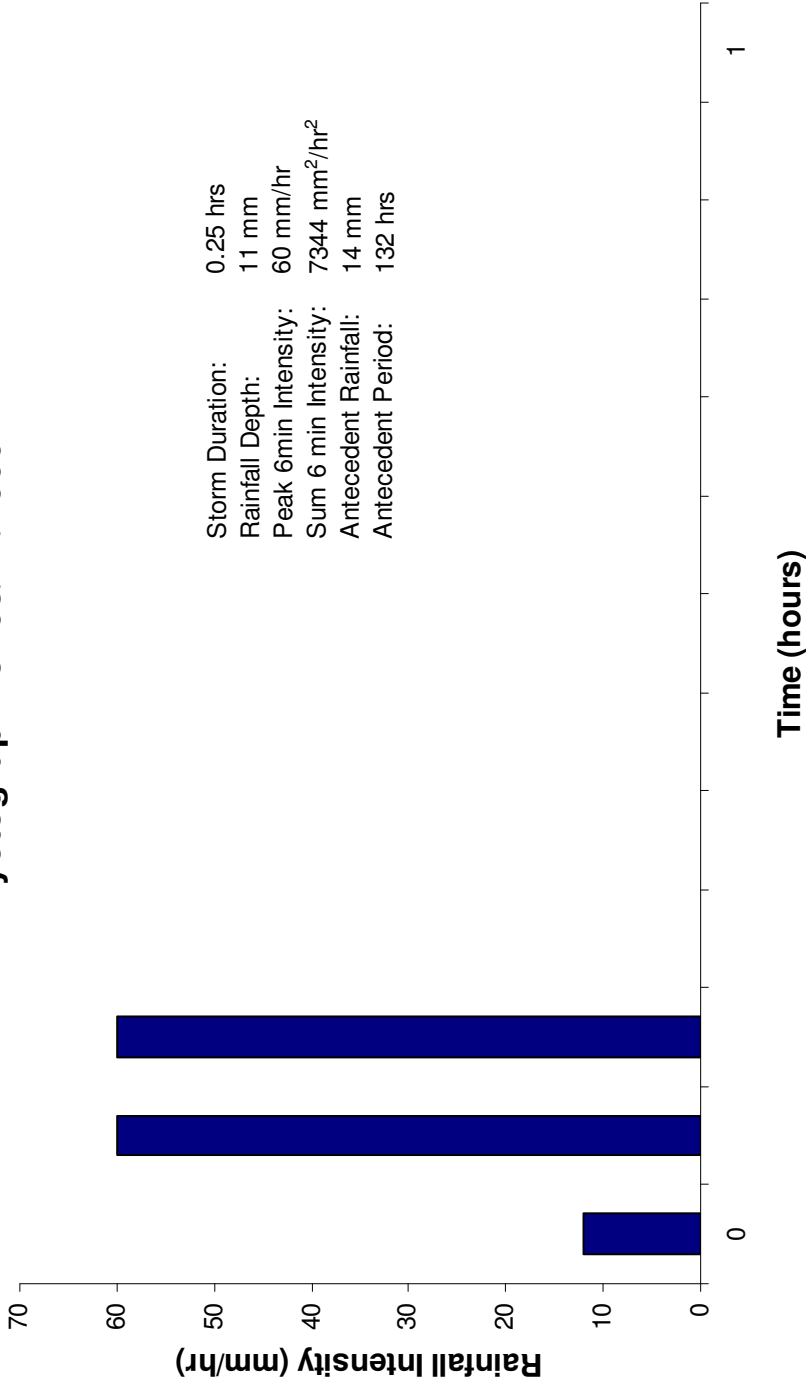


Sensitivity of Soil Type on Model Output

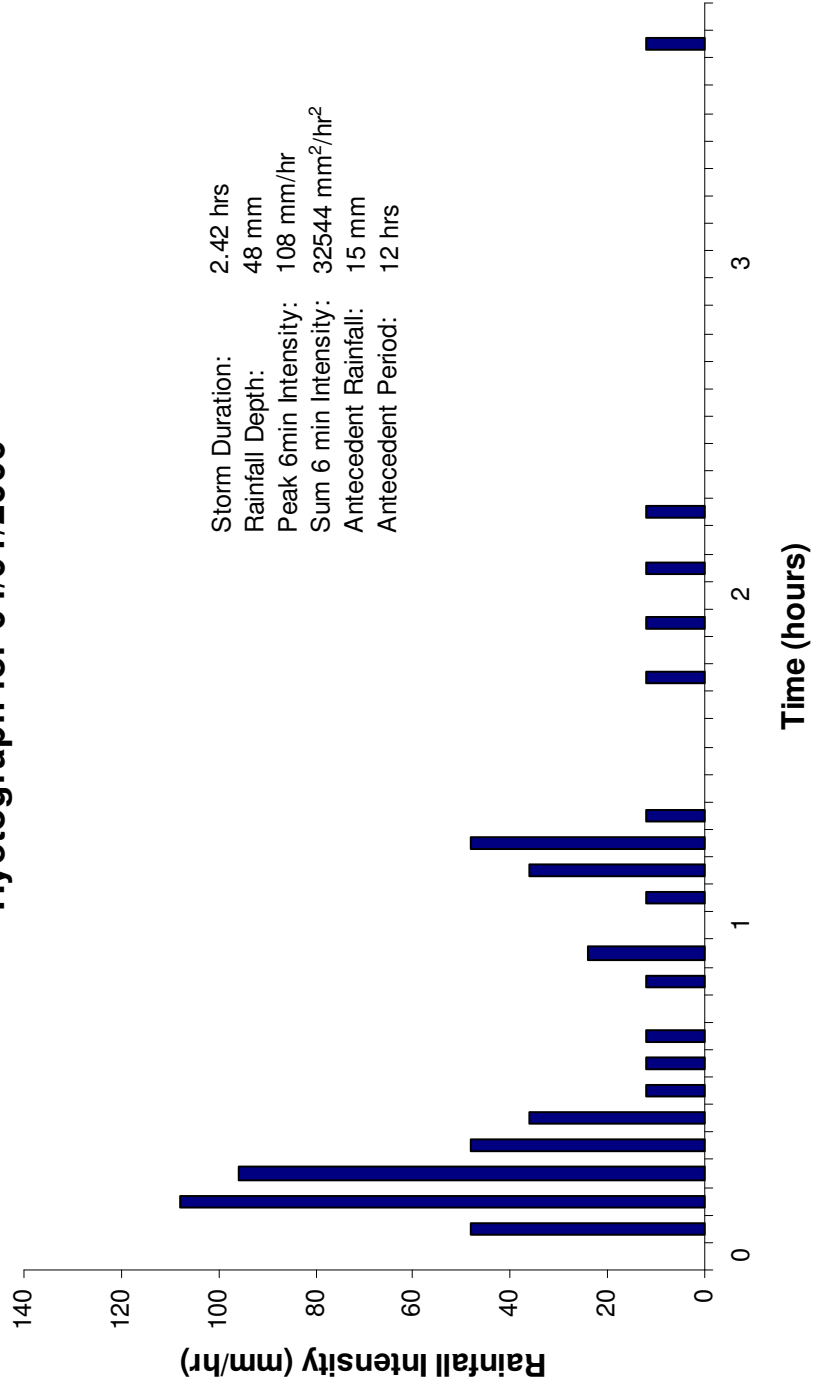


APPENDIX L: Hyetographs for Rainfall Events used to test Mass
Balance Model

Hyetograph for 08/12/2006



Hyetograph for 04/01/2006



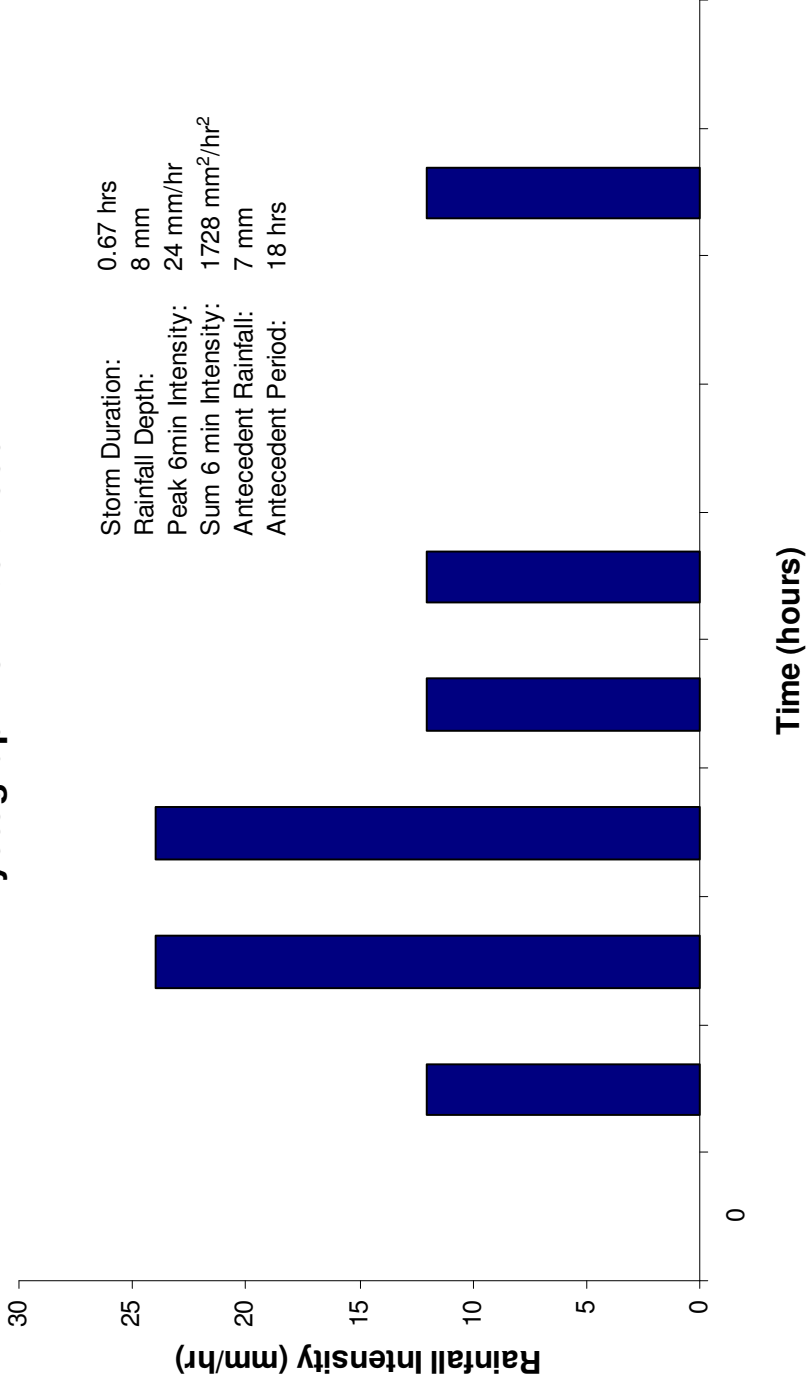
Hyetograph for Event 12021200

Storm Duration: 2.17 hrs
 Rainfall Depth: 61 mm
 Peak 6min Intensity: 84 mm/hr
 Sum 6 min Intensity: 32976 mm²/hr²
 Antecedent Rainfall: 1 mm
 Antecedent Period: 91 hrs

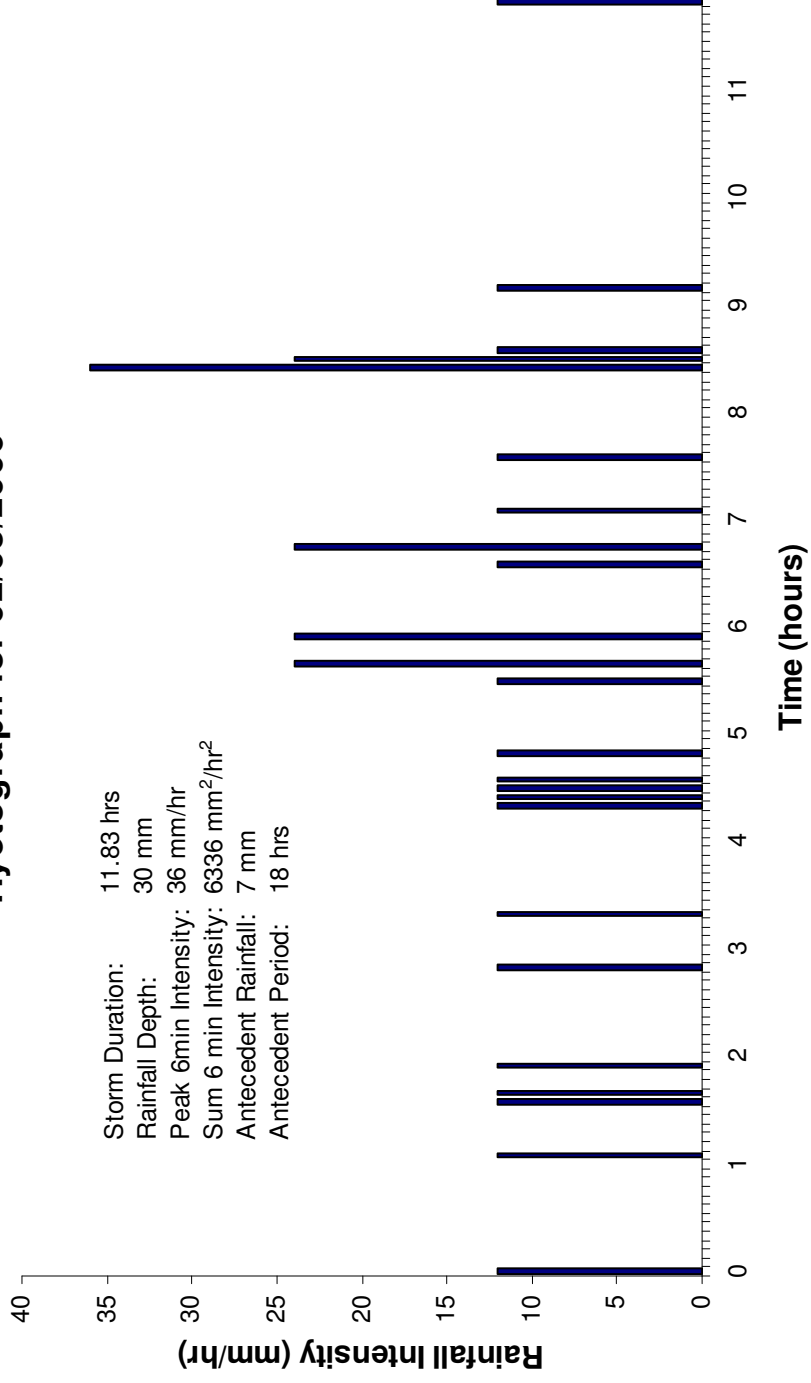
Time (hours)	Rainfall Intensity (mm/hr)
0.00 - 0.15	10
0.15 - 0.30	10
0.30 - 0.45	10
0.45 - 0.60	10
0.60 - 0.75	85
0.75 - 0.90	85
0.90 - 1.05	85
1.05 - 1.20	45
1.20 - 1.35	35
1.35 - 1.50	25
1.50 - 1.65	25
1.65 - 1.80	25
1.80 - 1.95	25
1.95 - 2.10	25
2.10 - 2.25	25
2.25 - 2.40	25
2.40 - 2.55	25
2.55 - 2.70	25
2.70 - 2.85	25
2.85 - 3.00	25
3.00 - 3.15	25
3.15 - 3.30	25
3.30 - 3.45	25
3.45 - 3.60	25
3.60 - 3.75	25
3.75 - 3.90	25
3.90 - 4.05	25
4.05 - 4.20	25
4.20 - 4.35	25
4.35 - 4.50	25
4.50 - 4.65	25
4.65 - 4.80	25
4.80 - 4.95	25
4.95 - 5.10	25
5.10 - 5.25	25
5.25 - 5.40	25
5.40 - 5.55	25
5.55 - 5.70	25
5.70 - 5.85	25
5.85 - 6.00	25
6.00 - 6.15	25
6.15 - 6.30	25
6.30 - 6.45	25
6.45 - 6.60	25
6.60 - 6.75	25
6.75 - 6.90	25
6.90 - 7.05	25
7.05 - 7.20	25
7.20 - 7.35	25
7.35 - 7.50	25
7.50 - 7.65	25
7.65 - 7.80	25
7.80 - 7.95	25
7.95 - 8.10	25
8.10 - 8.25	25
8.25 - 8.40	25
8.40 - 8.55	25
8.55 - 8.70	25
8.70 - 8.85	25
8.85 - 9.00	25
9.00 - 9.15	25
9.15 - 9.30	25
9.30 - 9.45	25
9.45 - 9.60	25
9.60 - 9.75	25
9.75 - 9.90	25
9.90 - 10.05	25
10.05 - 10.20	25
10.20 - 10.35	25
10.35 - 10.50	25
10.50 - 10.65	25
10.65 - 10.80	25
10.80 - 10.95	25
10.95 - 11.10	25
11.10 - 11.25	25
11.25 - 11.40	25
11.40 - 11.55	25
11.55 - 11.70	25
11.70 - 11.85	25
11.85 - 12.00	25
12.00 - 12.15	25
12.15 - 12.30	25
12.30 - 12.45	25
12.45 - 12.60	25
12.60 - 12.75	25
12.75 - 12.90	25
12.90 - 13.05	25
13.05 - 13.20	25
13.20 - 13.35	25
13.35 - 13.50	25
13.50 - 13.65	25
13.65 - 13.80	25
13.80 - 13.95	25
13.95 - 14.10	25
14.10 - 14.25	25
14.25 - 14.40	25
14.40 - 14.55	25
14.55 - 14.70	25
14.70 - 14.85	25
14.85 - 15.00	25
15.00 - 15.15	25
15.15 - 15.30	25
15.30 - 15.45	25
15.45 - 15.60	25
15.60 - 15.75	25
15.75 - 15.90	25
15.90 - 16.05	25
16.05 - 16.20	25
16.20 - 16.35	25
16.35 - 16.50	25
16.50 - 16.65	25
16.65 - 16.80	25
16.80 - 16.95	25
16.95 - 17.10	25
17.10 - 17.25	25
17.25 - 17.40	25
17.40 - 17.55	25
17.55 - 17.70	25
17.70 - 17.85	25
17.85 - 18.00	25
18.00 - 18.15	25
18.15 - 18.30	25
18.30 - 18.45	25
18.45 - 18.60	25
18.60 - 18.75	25
18.75 - 18.90	25
18.90 - 19.05	25
19.05 - 19.20	25
19.20 - 19.35	25
19.35 - 19.50	25
19.50 - 19.65	25
19.65 - 19.80	25
19.80 - 19.95	25
19.95 - 20.1	

Storm Duration:	2.17 hrs
Rainfall Depth:	61 mm
Peak 6min Intensity:	84 mm/hr
Sum 6 min Intensity:	32976 mm ² /hr ²
Antecedent Rainfall:	1 mm
Antecedent Period:	91 hrs

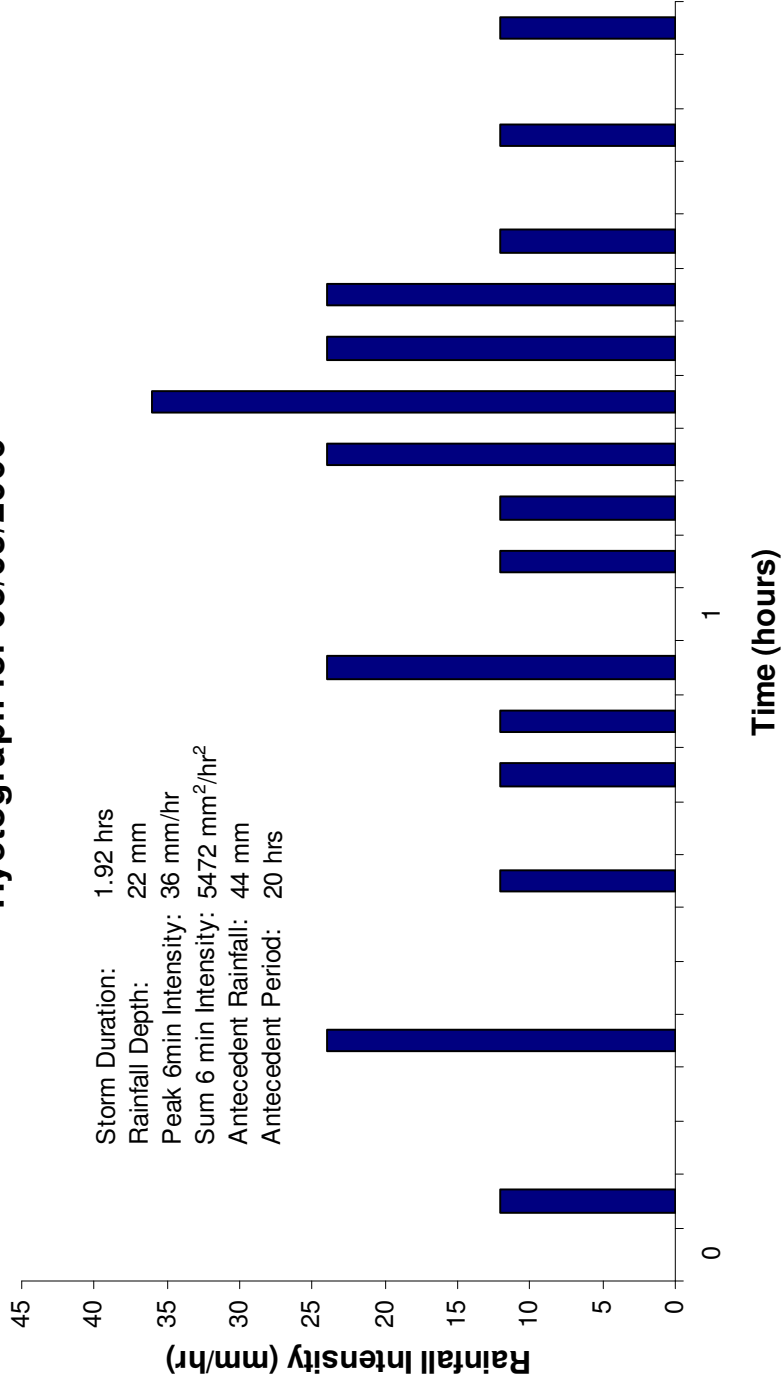
Hyetograph for 17/02/2006



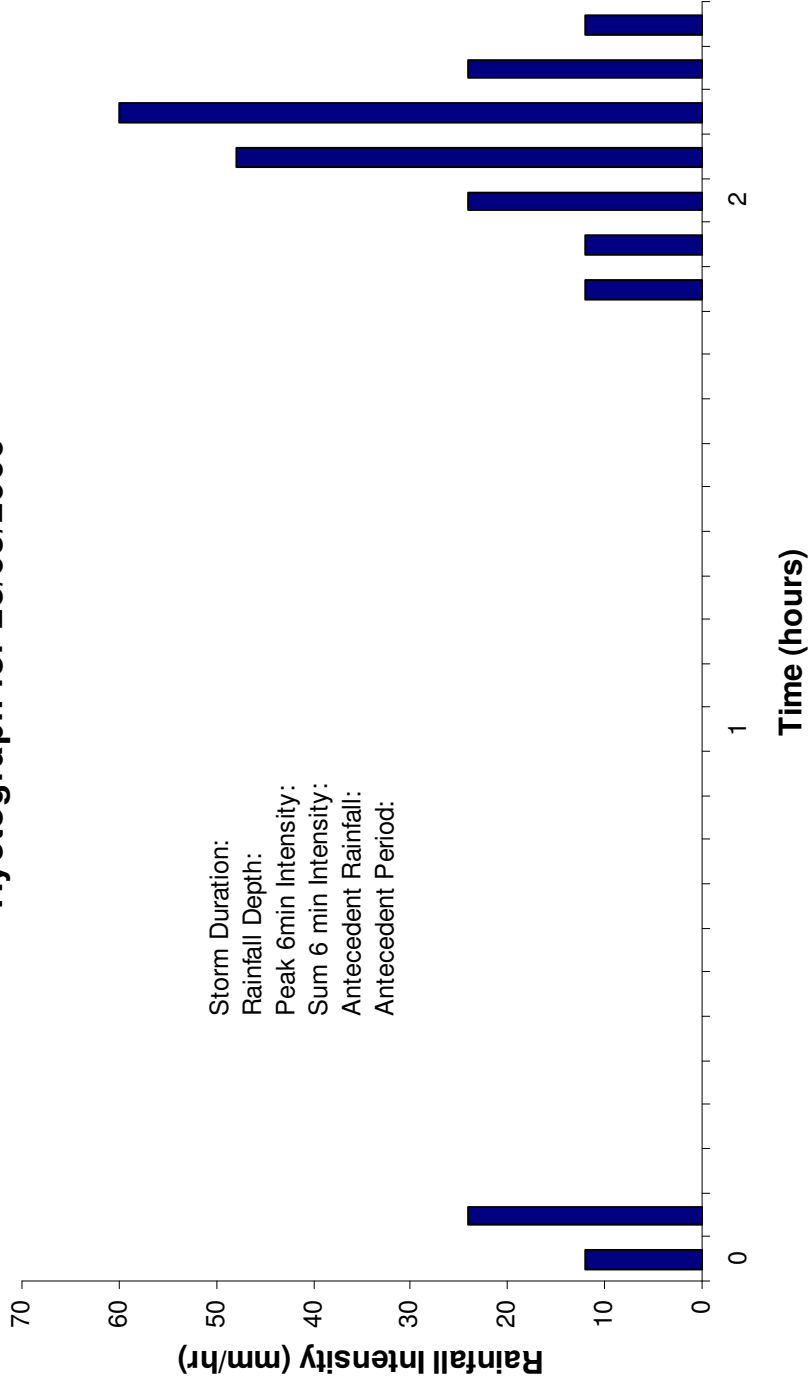
Hyetograph for 02/03/2006



Hyetograph for 05/03/2006



Hyetograph for 23/03/2006



APPENDIX M: Mass Balance Model Spreadsheet Output

Catchment HSA

MASS BALANCE SPREADSHEET TO ESTIMATE NCP EVENT MEAN CONCENTRATION FROM URBAN CATCHMENTS
CATCHMENT HSA

MASS BALANCE SPREADSHEET TO ESTIMATE NCP EVENT MEAN CONCENTRATION FROM URBAN CATCHMENTS																													
CATCHMENT HSA																													
ROAD															CARPARK														
TEdrain curve parameters															WEFP curve parameters														
TEdrain curve parameters															WEFP curve parameters														
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Catchment HSB

MASS BALANCE SPREADSHEET TO ESTIMATE HCP EVENT MEAN CONCENTRATION FROM URBAN CATCHMENTS

ROAD

Tetradrain curve parameters

WEFP curve parameters

TERmin

25.00%

WEFPmin

30.00%

PeakImin

5.0

RPmin

40.0

PeakIcritical

15

RPcri

200

RPDPcr

1.3000

Max FPDry

1600

ARFPwet

2000.0

CP

25%

Max FFWet

6000

ARFPdry

1000.0

MaxDPDry

2600

AROPdry

30.0

LRRPdrain

0.4

Storm Event

Art Rain

Ant Period

Storm Dur

Rain Depth

Sum Rain

Sum I

Peak I₆

Peak I₂

Sum I₂

Sum I₂²

Sum I₂/²D

14.00

132.00

0.25

11.00

11

44.0

60

3600

7344

29376

15.00

12.00

2.42

48.00

48

19.8

108

11664

32544

13448

1.00

91.00

2.17

61.00

61

28.1

84

7056

32976

15196

7.00

18.00

0.67

8.00

8

11.9

24

576

1728

2579

21.00

6.00

11.83

30.00

30

2.5

36

1296

6336

536

44.00

20.00

1.92

22.00

22

11.5

36

1296

5472

2860

3.00

4.00

2.33

19.00

19

8.2

60

3600

8208

3523

STATISTICAL DATA

1st quartile (25 %ile)

Minimum

Median

Maximum

3rd quartile (75 %ile)

Mean

Stand Dev

COV

n

Sum

5.0

1.0

14.0

44.0

18.0

15.0

14.6

1.0

7.0

105.0

9.0

4.0

18.0

132.0

55.5

40.4

50.3

1.2

7.0

283.0

1.3

0.3

2.2

11.8

2.4

3.1

3.9

1.3

7.0

21.6

15.0

8.0

22.0

61.0

39.0

28.4

19.6

0.7

7.0

199.0

9.8

2.5

11.9

44.0

24.0

18.0

14.1

0.8

7.0

126.0

36.0

24.0

60.0

108.0

72.0

56.3

29.7

0.5

7.0

408.0

1296.0

576.0

3600.0

11664.0

5328.0

4155.4

3970.2

1.0

7.0

29088.0

5904.0

1728.0

7344.0

32976.0

20376.0

13515.4

13305.6

1.0

7.0

94508.0

2714.6

535.6

3522.7

29376.0

14322.1

9644.0

10417.6

7.0

7.0

67507.7

TERmin

20.00%

WEFPmin

60.00%

PeakImin

6.0

RPmin

40.0

PeakIcritical

25

RPcri

100

RPDPcr

940.0

Max FPDry

500

ARFPwet

300.0

CP

30%

Max FFWet

500

ARFPdry

300.0

MaxDPDry

1000

AROPdry

5.0

LRRPdrain

0.5

Storm Event

Art Rain

Ant Period

Storm Dur

Rain Depth

Sum Rain

Sum I

Peak I₆

Peak I₂

Sum I₂

Sum I₂²

Sum I₂/²D

105.0

283.0

21.6

199.0

199.0

126.0

408.0

29088.0

94508.0

67507.7

100%

100%

100%

96%

100%

100%

100%

100%

100%

100%

100%

100%

100%

100%

100%

100%

100%

100%

100%

100%

0.0

1.1

1.4

0.0

0.0

0.0

0.0

0.0

0.0

0.0

0

324

426

0

0

0

0

0

0

0

132.0

12.0

91.0

18.0

6.0

20.0

4.0

20

20

20

500

500

455

500

500

500

500

500

500

500

17

0

0

0

24

0

0

0

0

0

500

500

955

500

500

500

500

500

500

500

1177

560

955

565

524

630

520

520

520

520

0

0

0

25

0

0

0

0

0

0

Catchment HSC

MASS BALANCE SPREADSHEET TO ESTIMATE NCP EVENT MEAN CONCENTRATION FROM URBAN CATCHMENTS
CATCHMENT HSC

ROAD																											
Storm Event		RAINFALL DATA										WEFP curve parameters															
		Ant Rain (mm)	Ant Period (hrs)	Storm Dur (hrs)	Rain Depth (mm)	Sum Rain Depth for Storm (mm)	Mean I (mm/hr)	Peak I _e (mm/hr)	Peak I _e ²	Sum I _e ²	SumI2/D	TEdrain (%)	WEFP (%)	Twet=IBP (hrs)	LFPwet (mg/m2)	Tdry (hrs)	Tdry x ARFPdry (mg/m2)	LDPsurf (mg/m2)	LDPdrain (mg/m2)	LFPi PreStorm	LFPsurf Storm	LRPdrain (mg/m2)	LRPsurf Post Storm	LRPdrain PostStorm	LRPsurf Post Storm		
8/12/2006		14.00	132.00	0.25	11.00	11	44.0	60	3600	7344	29376	100%	100%	0.0	0	132.0	1600	2600	1600	1600	24	1600	4224	4224	4224	0	0
4/01/2006		15.00	12.00	2.42	48.00	48	19.8	108	11864	32544	13448	100%	100%	1.1	2160	12.0	1600	360	1600	1600	3760	0	3760	4120	4120	0	0
12/02/2006		1.00	91.00	2.17	61.00	61	28.1	84	7056	32976	15196	100%	100%	1.4	2840	91.0	1600	2600	1600	1600	4440	0	4440	7040	7040	0	0
17/02/2006		7.00	18.00	0.67	8.00	8	11.9	24	576	1728	2579	100%	100%	0.0	0	18.0	1600	540	1600	1600	1600	0	1600	2140	2140	0	0
2/03/2006		21.00	6.00	11.83	30.00	30	2.5	36	1296	6336	536	100%	100%	0.0	0	6.0	1600	180	1600	1600	1600	0	1600	1600	1600	0	0
5/03/2006		44.00	20.00	1.92	22.00	22	11.5	36	1296	5472	2650	100%	100%	0.0	0	20.0	1600	780	1600	1600	1600	0	1600	2380	2380	0	0
23/03/2006		3.00	4.00	2.33	19.00	19	8.2	60	3600	8208	3523	100%	100%	0.0	0	4.0	1600	120	1600	1600	1600	0	1600	1720	1720	0	0
CARPARK																											
Storm Event		RAINFALL DATA										WEFP curve parameters															
		Ant Rain (mm)	Ant Period (hrs)	Storm Dur (hrs)	Rain Depth (mm)	Sum Rain Depth for Storm (mm)	Mean I (mm/hr)	Peak I _e (mm/hr)	Peak I _e ²	Sum I _e ²	SumI2/D	TEdrain (%)	WEFP (%)	Twet=IBP (hrs)	LFPwet (mg/m2)	Tdry (hrs)	Tdry x ARFPdry (mg/m2)	LDPsurf (mg/m2)	LDPdrain (mg/m2)	LFPi PreStorm	LFPsurf Storm	LRPdrain (mg/m2)	LRPsurf Post Storm	LRPdrain PostStorm	LRPsurf Post Storm		
1st quartile (25 %ile)		5.0	9.0	1.3	15.0	15	9.8	36.0	1296.0	5904.0	2714.6	100%	100%	0.0	0	132.0	500	660	500	500	17	500	1177	1177	500	0	0
Minimum		1.0	4.0	0.3	8.0	8	2.5	24.0	576.0	1728.0	535.6	100%	100%	1.1	324	12.0	500	60	500	60	500	0	500	560	560	0	0
Median		14.0	18.0	2.2	22.0	22	11.9	60.0	3600.0	7344.0	3522.7	100%	100%	1.4	426	91.0	500	455	500	455	500	0	500	955	955	0	0
Maximum		44.0	132.0	11.8	61.0	61	44.0	108.0	11864.0	32976.0	29376.0	100%	100%	0.0	0	18.0	500	90	500	90	500	0	500	590	565	25	0
3rd quartile (75 %ile)		18.0	55.5	2.4	39.0	39	24.0	72.0	5328.0	20376.0	14322.1	100%	100%	0.0	0	6.0	300.0 CP	30	500	30	500	24	500	524	524	0	0
Mean		15.0	40.4	3.1	28.4	28	18.0	58.3	4155.4	13515.4	9644.0	100%	100%	0.0	0	20.0	300.0	130	500	130	500	0	500	630	630	0	0
Stand Dev		14.6	50.3	3.9	19.6	19.6	14.1	29.7	3970.2	13305.6	10417.6	100%	100%	0.0	0	4.0	5.0	20	500	20	500	0	500	520	520	0	0
COV		1.0	1.2	1.3	0.7	0.7	0.8	0.5	1.0	1.0	1.1	100%	100%	0.0	0	4.0	0.5	20	500	20	500	0	500	520	520	0	0
n		7.0	7.0	7.0	7.0	7.0	7.0	7.0	7.0	7.0	7.0	100%	100%	0.0	0	4.0	0.5	20	500	20	500	0	500	520	520	0	0
Sum		105.0	283.0	21.6	199.0	199.0	126.0	408.0	29088.0	94508.0	67507.7	100%	100%	0.0	0	4.0	500	660	500	500	17	500	1177	1177	500	0	0

ROOF

TEdrain curve parameters				WEFP curve parameters			
TEmin	5.00%	WEFPmin	80.00%	TEmin	5.00%	WEFPmin	80.00%
Peakmin	4.0	RPmin	10.0	Peakmin	4.0	RPmin	10.0
Peakcritical	25	RPcrit	100	Peakcritical	25	RPcrit	100

30%

3.0 CP

160

Max FPDry

Max FFWet

MaxDPdry

250

ARFPwet

ARFPdry

5.0

ARDPdry

0.0

LRPDrain

0.5

TEdrain (%)	WEFP (%)	Twet=D (hrs)	LFPwet (mg/m2)	Tdry (hrs)	Tdry x ARFPdry (mg/m2)	LDPsurf (mg/m2)	LDPdrain (mg/m2)	LDPi PreStorm (mg/m2)	LFPsurf Storm (mg/m2)	LRPdrain (mg/m2)	LFPdrain (mg/m2)	Ldrain (mg/m2)	L (mg/m2)	Sum L (mg/m2)	LRPdrain PostStorm	LRPsurf Post Storm
-------------	----------	--------------	----------------	------------	------------------------	-----------------	------------------	-----------------------	-----------------------	------------------	------------------	----------------	-----------	---------------	--------------------	--------------------

100%	100%	0.3	1	132.0	160	0	0	0	161	17	161	178	178	178	50	0
100%	100%	2.4	7	12.0	60	0	0	0	60	67	0	67	67	67	0	0
100%	100%	2.2	7	91.0	160	0	0	0	160	167	0	167	167	167	0	0
95%	100%	0.7	2	18.0	90	0	0	0	90	92	0	92	92	88	4	0
100%	100%	11.8	35	6.0	30	0	0	0	30	65	4	65	70	70	0	0
100%	100%	1.9	6	20.0	100	0	0	0	100	106	0	106	106	106	0	0
100%	100%	2.3	7	4.0	20	0	0	0	20	27	0	27	27	27	0	0

EMC AND LOAD CALCULATIONS

Road		Carpark		Roof		Grass		Total	Road Flow	Carpark Flow	Roof Flow (with 0.4)	Grass Flow	Predicted Flow	Sum Pre Flow	Predicted EMC	Meas EMC	Modelled Flow	Predicted Load	Modelled Load
1020		453		0		1497													
(kg)	(mg/L)	(kg)	(mg/L)	(kg)	(mg/L)	(kg)	(mg/L)	(kg)	(ML)	(ML)	(ML)	(ML)	(ML)	(ML)	(mg/L)	(mg/L)	(ML)	(g)	(g)
4	422	1	118	0	#DIV/0!	0.7	0	0	5	0.01	0.0000	0.0000	0.0147	0.0147	373.4	60	0.0009	5499.9330	54.0000
4	88	0	12	0	#DIV/0!	2.9	110	7	7	0.05	0.0000	0.0261	0.0954	0.0954	76.9	220	0.0036	7330.3200	792.0000
7	117	0	16	0	#DIV/0!	3.7	96	11	11	0.06	0.0000	0.0382	0.1266	0.1266	89.0	180	0.0042	11266.0950	756.0000
2	306	0	81	0	#DIV/0!	0.5	0	3	3	0.01	0.0000	0.0000	0.0103	0.0103	283.0	60	0.0002	2917.8565	12.0000
2	55	0	18	0	#DIV/0!	1.8	190	4	4	0.03	0.0000	0.0094	0.0521	0.0521	70.3	140	0.0020	3665.8159	280.0000
2	113	0	30	0	#DIV/0!	1.3	657	4	4	0.02	0.0000	0.0020	0.0329	0.0329	122.4	15	0.0020	4030.3500	30.0000
2	96	0	29	0	#DIV/0!	1.1	0	3	3	0.02	0.0000	0.0000	0.0265	0.0265	118.0	170	0.0039	3127.6800	663.0000

STATISTICAL DATA

1st quartile	2.0	91.6	0.2	17.0	0.0	#DIV/0!	0.9	0.0	38		7.0	161.8	120.7	7.0
Minimum	1.6	55.2	0.2	11.9	0.0	#DIV/0!	0.5	0.0						
Median	2.4	113.3	0.3	28.9	0.0	#DIV/0!	1.3	95.6						
Maximum	7.2	422.4	0.5	117.7	0.0	#DIV/0!	3.7	656.7						
3rd quartile	4.3	211.5	0.4	55.4	0.0	#DIV/0!	2.3	150.2						
Mean	3.38	171.02	0.32	43.32	0.00	#DIV/0!	1.70	150.40	234.68					
Stand Dev	2.00	137.59	0.12	40.19	0.00	#DIV/0!	1.17	234.68						
COV	0.6	0.8	0.4	0.9	#DIV/0!	#DIV/0!	0.7	1.6						
n	7.0	7.0	7.0	7.0	7.0	7.0	7.0	7.0	12					
Sum	24	2	2	0	0	0	12	12						

APPENDIX N: Mass Balance Model Input Error Analysis

Mass Balance Parameters	Units	Multiplication Factor	HSA		HSB		HSC		TOTAL		Percentage Change in Total Error			Average Percentage Change in Total Error		
			Cumulative EMC Error	Maximum EMC Error	Cumulative EMC Error	Maximum EMC Error	Cumulative EMC Error	Maximum EMC Error	Cumulative EMC Error	Maximum EMC Error	Cumulative EMC Error	Maximum EMC Error	Maximum EMC Error	Cumulative EMC Error	Maximum EMC Error	Maximum EMC Error
Unmodified	-	-	56%	528%	35%	475%	34%	716%	41%	716%	0%	0%	0%	0%	0%	0%
Rainfall Parameters	Antecedent Rainfall Total	0.5 2	56%	528%	35%	475%	34%	716%	41%	716%	0%	0%	0%	0%	0%	0%
	Antecedent Rainfall Period	0.5 2	40%	444%	20%	396%	21%	629%	27%	629%	-35%	-12%	24%	34%	18%	18%
	Storm Duration	0.5 2	56%	528%	35%	475%	34%	713%	41%	713%	0%	0%	16%	6%	8%	8%
	Rainfall Depth	0.5 2	226%	1241%	186%	1145%	171%	1426%	193%	1426%	369%	99%	-56%	267%	78%	78%
	Peak 6 min Intensity	0.5 2	-19%	207%	-32%	178%	-28%	316%	-27%	316%	-14%	0%	0%	7%	0%	0%
	Sum of the Square of all 6 min Intensities	0.5 2	50%	528%	29%	475%	30%	827%	36%	827%	-13%	16%	0%	6%	8%	8%
	Interburst Period	0.5 2	56%	528%	35%	475%	34%	716%	41%	716%	0%	0%	0%	11%	0%	0%
Particle Accumulation and Washoff Parameters	Maximum Free Particle Load Prior to Storm	0.5 2	52%	528%	31%	475%	31%	716%	38%	716%	-9%	0%	0%	11%	0%	0%
	Maximum Free Particle Load During Storm	0.5 2	62%	528%	41%	475%	38%	716%	47%	716%	14%	0%	0%	119%	36%	36%
	Maximum Detained Particle Load Prior to Storm	0.5 2	19%	417%	1%	371%	5%	528%	8%	528%	-81%	-26%	46%	6%	2%	2%
	Wet Weather Accumulation Rate of Free Particles During the Storm	0.5 2	129%	744%	102%	681%	86%	1046%	106%	1046%	156%	46%	0%	6%	0%	0%
	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	52%	528%	31%	474%	28%	693%	37%	693%	-11%	-3%	0%	35%	0%	0%
	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	56%	528%	35%	475%	34%	716%	41%	716%	0%	0%	0%	11%	0%	0%
	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	39%	369%	19%	325%	22%	716%	26%	716%	-36%	0%	0%	0%	0%	0%
	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	72%	711%	49%	650%	45%	716%	55%	716%	33%	0%	0%	35%	0%	0%
	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	52%	528%	31%	475%	31%	716%	38%	716%	-9%	0%	0%	11%	0%	0%
	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	62%	528%	41%	475%	38%	716%	47%	716%	14%	0%	0%	0%	0%	0%
Particle Accumulation and Washoff Parameters	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	56%	528%	35%	475%	34%	716%	41%	716%	0%	0%	0%	0%	0%	0%
	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	40%	439%	20%	394%	21%	629%	26%	629%	-36%	-12%	24%	34%	18%	18%
	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	71%	530%	48%	475%	46%	889%	55%	889%	32%	24%	0%	0%	0%	0%
	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	56%	530%	35%	476%	34%	716%	41%	716%	0%	0%	0%	0%	0%	0%
	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	56%	530%	35%	476%	34%	716%	41%	716%	0%	0%	0%	0%	0%	0%
	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	56%	528%	35%	475%	34%	716%	41%	716%	0%	0%	0%	0%	0%	0%
	Dry Weather Accumulation Rate of Free Particles During the Storm	0.5 2	56%	528%	35%	475%	34%	716%	41%	716%	0%	0%	0%	0%	0%	0%
Particle Accumulation and Washoff Parameters	Adjustment Factor for Total Particle Load in Lateral Drain	0.5 2	44%	503%	27%	460%	18%	582%	29%	582%	-28%	-19%	37%	43%	28%	28%
	Particle Washoff Multiplier for Grassed Areas	0.5 2	81%	576%	49%	503%	66%	982%	65%	982%	57%	37%	0%	43%	28%	28%